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APPENDIX E

TOXICITY REFERENCE VALUES

Screening Level Ecological Risk Assessment Protocol

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APPENDIX E

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APPENDIX E

TOXICITY REFERENCE VALUES

Appendix E presents implementation of the recommended approach (described in Chapter 5) for identifying toxicity reference values (*TRVs*) for measurement receptors. Discussion is provided for determining compound-specific *TRV* values for community and wildlife measurement receptors.

Following the guidance in Sections E-1.0 through E-1.2, U.S. EPA OSW has identified default *TRV* values for the measurement receptors of the seven example food webs (listed in Chapter 4) and the compounds commonly identified in ecological risk assessments for combustion facilities (identified in Chapter 2). Section E-1.0 describes the determination of *TRV* values for surface water, sediment, and soil community measurement receptors in the example food webs. Section E-2.0 describes determination of *TRV* values for wildlife measurement receptors in the example food webs. Tables E-1 through E-8 present the default *TRV* values selected, the basis for selection of each value, and the references evaluated in determination of each value.

TRV values for a limited number of compounds are included in this appendix (see Tables E-1 through E-3) to facilitate the completion of screening ecological risk assessments. However, it is expected that *TRV* values for additional compounds and receptors may be required for evaluation on a site specific basis. In such cases, *TRV* values for these additional compounds could be determined following the same guidance used in determination of the *TRV* values reported in this appendix. For the determination of *TRV* values for measurement receptors not specifically represented in Sections E-1.0 through E-2.0 (e.g., amphibians and reptiles), an approach consistent to that presented in this appendix could be utilized by applying data applicable to those measurement receptors being evaluated.

The default *TRVs* provided in Tables E-1 through E-8 are based on values reported in available scientific literature. Toxicity values identified in secondary reference sources were verified, where possible, by reviewing the primary reference source. As noted in Chapter 5, *TRV* values may change as additional toxicity research is conducted and the availability of toxicity data in the scientific literature increases. As a result, U.S. EPA OSW recommends evaluating the latest toxicity data before completing a risk assessment to ensure that the toxicity data used in the risk assessment is the most current. If more appropriate *TRV* values can be documented, they should be used presented to the respective permitting authority for approval.

TRVs were not identified for amphibians and reptiles because of the paucity of toxicological information on these receptors. Additional guidance on determination and use of *TRV* values in the screening level ecological risk assessment is provided in Chapter 5.

E-1.0 *TRVs* FOR COMMUNITY MEASUREMENT RECEPTORS IN SURFACE WATER, SEDIMENT, AND SOIL

TRV values provided in this appendix for community measurement receptors in surface water, sediment, and soil were identified from screening toxicity values developed and/or adopted by federal and/or state regulatory agencies. As discussed in Chapter 5, these screening toxicity values are generally provided in the form of standards, criteria, guidance, or benchmarks. For compounds with no available screening toxicity value, *TRVs* were determined using toxicity values from available scientific literature. The

equilibrium partitioning (EqP) approach was used to compute several sediment *TRVs*. Uncertainty factors (UFs) were applied to toxicity values, as necessary, to meet the *TRV* criteria discussed in Chapter 5. The following sections discuss determination of *TRV* values for community receptors in surface water, sediment, and soil.

Freshwater TRVs Freshwater *TRVs* should be used for freshwater and estuarine ecosystems with a salinity less than 5 parts per thousand. Freshwater *TRVs*, based on the dissolved concentration of the compound in surface water, are listed in Table E-1. *TRVs* were identified using the following hierarchy:

1. Federal chronic ambient water quality criteria (AWQC) calculated for with no final residue value (U.S. EPA 1999; 1996b). Federal AWQC for cadmium, copper, lead, nickel, and zinc were multiplied by a chemical-specific conversion factor to determine a *TRV* based on dissolved concentration (U.S. EPA 1999; 1996b).
2. Final chronic values (FCV) for COPCs for which their AWQC included a final residue value (U.S. EPA 1996b).
3. If inadequate data (insufficient number of families of aquatic life with toxicity data) were available to compute an AWQC or FCV, U.S. EPA (1999; 1996b) also reported secondary chronic values (SCV) calculated using the Tier II method in the Great Lakes Water Quality Initiative (GLWQI) (reported in 40 CFR Part 122). This method is similar to the procedures for calculating an FCV. It uses statistically-derived “adjustment factors” to address deficiencies in available data. The adjustment factor decreases as the number of representative families increases.
4. If an AWQC, FCV, or GLWQI Tier II SCV value were not available, toxicity values cited by U.S. EPA (1987) were identified. These toxicity values represent the lowest available values. Further, additional toxicity values available from the AQUIRE database in U.S. EPA’s *ECOTOXicology Database System* (U.S. EPA 1996a) were identified. If collected from a secondary source (such as AQUIRE), original studies were obtained and reviewed for accuracy. The toxicity values reported in Table E-1 represent the lowest (most conservative), ecologically relevant, available value.
5. If toxicity data were unavailable, a surrogate *TRV* from a COPC with a similar structure was identified.
6. If no surrogate was available, a *TRV* was not listed. The potential toxicity of a COPC with no *TRV* should be addressed as an uncertainty (see Chapter 6)

Standard AQUIRE report summaries on tests were screened for duration, endpoint, effect, and concentration. Studies were also screened for ecologically relevant effects by focusing on studies that evaluated effects on survival, reproduction, and growth. Aspects of endpoint, duration, and test organism in each toxicity study were evaluated to identify the most appropriate study. Several compounds, most notably metals, had a large number of toxicity values based on various endpoints, organisms, and exposure durations. In these instances, best scientific judgment was used to identify the most appropriate toxicity value (see Chapter 5).

Chronic NOAEL-based values were not adjusted, but rather were carried through unchanged to become the *TRV*. Toxicity values identified as “less than” a particular concentration were divided by 2 to represent an average value because the true value is unknown, and it occurs between 0 and the noted concentration. *UFs* discussed in Chapter 5 were applied to toxicity values not meeting *TRV* criteria.

Saltwater TRVs Saltwater *TRVs* are applicable to marine water bodies and estuarine systems with a salinity greater than 5 ppt. Saltwater *TRVs* are listed in Table E-2. Saltwater water *TRV* development followed the same procedure as described above for freshwater receptors, except no GLWQI Tier II SCVs were available. In addition, if no saltwater *TRV* for a surrogate compound was available, the corresponding freshwater *TRV* was adopted.

Freshwater Sediment TRVs Freshwater sediment *TRVs* are listed in Table E-3. They are applicable to water bodies with a salinity less than 5 ppt. Freshwater sediment *TRVs* were identified from various sets of screening values and ecotoxicity review documents. The lowest available screening values among the following sources were identified:

1. No effect level (NEL) and lowest effect level (LEL) values from “Ontario’s Approach to Sediment Assessment and Remediation” (Persaud et al. 1993)
2. Apparent effects threshold (AET) values for the amphipod, *Hyallela azteca*, reported in “Creation of Freshwater Sediment Quality Database and Preliminary Analysis of Freshwater Apparent Effects Thresholds” (Washington State Department of Ecology 1994)
3. Sediment effect concentrations jointly published by the National Biological Service and the U.S. EPA (Ingersoll et al. 1996).

If a screening value was not available in the sources listed above, toxicity studies and other values compiled and reported by Jones, Hull, and Suter (1997) were reviewed to identify possible *TRVs*. Relevant studies were prioritized based on the criteria listed in Chapter 5, and uncertainty factors were applied, as applicable, based on criteria presented (see Chapter 5).

If a screening or sediment toxicity value was not available for an organic COPC, a freshwater sediment *TRV* was computed, using the EqP approach (see Chapter 5), from the compounds corresponding freshwater *TRV* and K_{oc} value. The U.S. EPA Office of Water utilizes the EqP approach to develop sediment quality criteria for nonionic (neutral) organic chemicals (U.S. EPA 1993). The EqP approach assumes that the toxicity of a compound in sediment is a function of the concentration in pore water and that to be nontoxic, the pore water must meet the surface water final chronic value. The EqP approach also assumes that the concentration of a compound in sediment pore water depends on the carbon content of the sediment and the compound’s organic carbon partitioning coefficient (U.S. EPA 1993). A *TRV* may be calculated using the following equation (U.S. EPA 1993):

$$TRV_{sed} = K_{oc} \cdot f_{oc} \cdot TRV_{sw} \quad \text{Equation E-1}$$

where

$$TRV_{sed} = \text{Sediment } TRV \text{ (}\mu\text{g/kg)}$$

K_{oc}	=	Organic carbon partition coefficient (L/kg)
f_{oc}	=	Fraction of organic carbon in sediment (unitless)—default value = 4% (0.04)
TRV_{sw}	=	Corresponding surface water TRV ($\mu\text{g/L}$)

Marine Sediment $TRVs$ Marine sediment $TRVs$ are listed in Table E-4. They are applicable to sediments of marine water bodies and estuarine systems with a salinity greater than 5 ppt. Marine sediment $TRVs$ were developed following the procedures used to identify the freshwater sediment $TRVs$. Screening values were compiled from the following sources:

1. No observed effect level (NOEL) sediment quality assessment guidelines for State of Florida coastal waters (MacDonald 1993).
2. Marine and estuarine effects range low (ERL) values from “Incidence of Adverse Biological Effects Within Ranges of Chemical Concentrations in Marine and Estuarine Sediments” (Long et al. 1995)
3. ERL values from “The Potential for Biological Effects of Sediment-Sorbed Contaminants Tested in the National Status and Trends Program” (Long and Morgan 1991)
4. Marine sediment quality criteria from “Sediment Management Standards” (Washington State Department of Ecology 1991)

Screening values were adopted directly as $TRVs$. If a screening value was not available in the sources listed above, toxicity values from a search of the scientific literature and those compiled and reported by Hull and Suter (1994) were reviewed to identify possible $TRVs$. Original studies were obtained, where possible, and toxicity values were verified. Relevant studies were prioritized based on the criteria listed in Chapter 5, and uncertainty factors were applied, as appropriate, based on criteria (see Chapter 5). If a screening or ecologically relevant sediment toxicity value from the scientific literature were not available for an organic COPC, a marine sediment TRV was computed, using the EqP approach, from the COPC’s corresponding saltwater TRV and K_{oc} value (see Equation E-1).

Terrestrial Plant $TRVs$ The terrestrial plant $TRVs$ listed in Table E-5 are based on bulk soil exposures. Available terrestrial plant toxicity values from the scientific literature were used to develop presented TRV values. Toxicity values were first identified from the following secondary sources:

1. Studies cited in *Toxicological Benchmarks for Screening Potential Contaminants of Concern for Effects on Terrestrial Plants: 1997 Revision* (Efroymson, Will, Suter, and Wooten 1997). Available studies were obtained and reviewed for accuracy of toxicity values. UFs were applied depending on study endpoint and available information.
2. Toxicity values in the Phytotox database in U.S. EPA’s *ECOTOXicology Database System*. Available studies were obtained and toxicity values were verified. UFs were applied depending on study endpoint and available information.
3. Toxicity values in U.S. EPA Region 5 *Ecological Data Quality Levels (EDQL) Database* (PRC 1995). The database contains media-specific EDQLs for the RCRA Appendix IX constituents (40 CFR Part 264). The EDQLs represent conservative media concentrations

protective of media receptors and wildlife that might be exposed through food chains based in these media. Available studies were obtained and toxicity values were verified. UFs were applied depending on study endpoint and available information.

Original studies were obtained, where possible, and prioritized based on criteria listed in Chapter 5. Uncertainty factors were applied, as appropriate, based on criteria (discussed in Chapter 5) to develop *TRV* values. For COPCs without toxicity data, the *TRV* for a surrogate COPC was adopted. If an appropriate surrogate *TRV* was not available, no *TRV* value was identified. Generally, review of toxicity data available in the scientific literature indicates that limited *TRVs* are available for organic compounds; while *TRVs* for metals are available.

Soil Invertebrate TRVs The soil invertebrate *TRVs* listed in Table E-6 are based on bulk soil exposures. Available soil invertebrate toxicity values from the scientific literature were used to develop *TRVs* for these receptors. Soil invertebrate toxicity values were first identified from the following secondary sources:

1. Studies cited in *Toxicological Benchmarks for Potential Contaminants of Concern for Effects on Soil and Litter Invertebrates and Heterotrophic Process* (Will and Suter II 1995a). Available studies were obtained and toxicity values were verified. UFs were applied depending on study endpoint and available information.
2. Scientific literature was searched for toxicity values for outstanding compounds. Relevant studies were obtained, toxicity values were verified, and UFs were applied as described.

Original studies were obtained, where possible, and prioritized based on criteria listed in Chapter 5. Uncertainty factors were applied, as appropriate, based on criteria to develop *TRVs*. If no toxicity value was available for a COPC, the *TRV* for a surrogate COPC was adopted.

E-2.0 *TRVs* FOR WILDLIFE MEASUREMENT RECEPTORS

TRV values for wildlife measurement receptors are listed in Tables E-7 (mammals) and E-8 (birds). *TRVs* were not developed for each avian and mammalian measurement receptor in the seven example food webs because of the paucity of species-specific data. Rather, U.S. EPA OSW focused on identifying a set of avian *TRVs* and a set of mammalian *TRVs* for the classes of compounds listed in Section 2.3. U.S. EPA OSW assumed that, among the literature reviewed for a particular guild, the lowest available toxicity value across orders in class Aves and across orders in class Mammalia would provide a conservative estimate of toxicity. Available mammalian and avian toxicity values from the scientific literature were used to develop *TRVs* for these receptors. Also, as previously noted, *TRV* values were not identified for amphibians and reptiles because of the paucity of toxicological information on these receptors. Wildlife measurement receptors *TRV* values were first identified from the following secondary sources:

1. Toxicity values compiled in *Toxicological Benchmarks for Wildlife: 1996 Revision* (Sample, Opresko, and Suter 1996).
2. Toxicity values listed in the Terretox database of U.S. EPA's *ECOTOXicology Database System* (U.S. EPA 1996b) were screened to identify studies potentially meeting the criteria listed in Chapter 5.

Original studies were compiled, where possible, and reviewed to verify their accuracy based on criteria listed in Chapter 5. In many cases, best scientific judgement was used to screen out studies with poor experimental design (see Chapter 5). Uncertainty factors were applied, as appropriate, to develop *TRVs* based on criteria presented in Chapter 5.

Conversions Some avian and mammalian toxicity data are expressed in terms of compound concentration in the food of the test organism. To convert to daily dose, it is necessary to determine the exposure duration and organism body weight. If the study does not report this information, the results should not be used to compute a *TRV*. If information on exposure duration and organism body weight is available, dietary concentration can be computed to dose using the following generic equation:

$$DD = \frac{C \cdot IR}{BW} \quad \text{Equation E-2}$$

where

<i>DD</i>	=	COPC dose (mg COPC/kg BW/day)
<i>C</i>	=	Concentration of COPC in diet (mg COPC/kg food)
<i>IR</i>	=	Food ingestion rate (kg/day)
<i>BW</i>	=	Test organism body weight (kg)

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APPENDIX E TEXT

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TABLES OF TOXICITY REFERENCE (TRV) VALUES

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TABLE E-1
FRESHWATER TOXICITY REFERENCE VALUES

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Compound	Toxicity Value		Uncertainty Factor ^b	TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Concentration			
Polychlorinateddibenzo-p-dioxins (μg/L)					
2,3,7,8-TCDD	Chronic LOEL	0.000038	0.1	0.0000038	Mehrle et al. (1988). 2,3,7,8-TCDD toxicity value for rainbow trout (<i>Oncorhynchus mykiss</i>).
Polynuclear aromatic hydrocarbons (PAH) (μg/L)					
Total high molecular weight (HMW) PAHs	--	--	--	0.014	Benzo(a)pyrene toxicity used as surrogate measure of toxicity. This TRV should be used if assessing the risk of total HMW PAHs.
Benzo(a)pyrene	Tier II value	0.014	Not applicable	0.014	U.S. EPA (1996). Calculated using Great Lakes Water Quality Initiative Tier II methodology.
Benzo(a)anthracene	Tier II SCV	0.027	Not applicable	0.027	Suter and Tsao (1996). Calculated using Great Lakes Water Quality Initiative Tier II methodology.
Benzo(b)fluoranthene	--	--	--	0.027	Toxicity value not available. Benzo(a)anthracene used as surrogate.
Benzo(k)fluoranthene	--	--	--	0.027	Toxicity value not available. Benzo(a)anthracene used as surrogate.
Chrysene	--	--	--	0.027	Toxicity value not available. Benzo(a)anthracene used as surrogate.
Dibenz(a,h)anthracene	--	--	--	0.027	Toxicity value not available. Benzo(a)anthracene used as surrogate.
Indeno(1,2,3-cd)pyrene	--	--	--	0.027	Toxicity value not available. Benzo(a)anthracene used as surrogate.
Polychlorinated biphenyls (PCB) (μg/L)					
Aroclor 1016	--	0.19	Not applicable	0.19	Adopted from U.S. EPA (1996) value for Total PCB. Calculated using Great Lakes Water Quality Initiative Tier II methodology.

TABLE E-1
FRESHWATER TOXICITY REFERENCE VALUES

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Compound	Toxicity Value		Uncertainty Factor ^b	TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Concentration			
Aroclor 1254	--	0.19	Not applicable	0.19	Adopted from U.S. EPA (1996) value for Total PCB. Calculated using Great Lakes Water Quality Initiative Tier II methodology.
Nitroaromatics ($\mu\text{g/L}$)					
1,3-Dinitrobenzene	Subchronic NOEC	260	0.1	26	van der Schalie (1983). Algal growth test with <i>Selenastrum capricornutum</i> .
2,4-Dinitrotoluene	Chronic LOEL	230	0.1	23	U.S. EPA (1987)
2,6-Dinitrotoluene	Chronic NOEC	60	Not applicable	60	Kuhn et al. (1989). Toxicity value for water flea (<i>Daphnia magna</i>).
Nitrobenzene	Acute LOEL	27,000	0.01 ^e	270	U.S. EPA (1987)
Pentachloronitrobenzene	LC50	1,000	0.01	10	Hashimoto and Nishiuchi (1981). Toxicity value for common carp (<i>Cyprinus carpio</i>).
Phthalate esters ($\mu\text{g/L}$)					
Bis(2-ethylhexyl)phthalate	Tier II SCV	3.0	Not applicable	3.0	Suter and Tsao (1996). Calculated using Great Lakes Water Quality Initiative Tier II methodology.
Di(n)octyl phthalate	Chronic NOEL	320	Not applicable	320	McCarthy and Whitmore (1985). Toxicity value for water flea (<i>D. magna</i>).
Volatile organic compounds ($\mu\text{g/L}$)					
Acetone	Tier II SCV	1,500	Not applicable	1,500	Suter and Tsao (1996). Calculated using Great Lakes Water Quality Initiative Tier II methodology.
Acrylonitrile	Chronic LOEL	2,600	0.1	260	U.S. EPA (1987)
Chloroform	Tier II SCV	28	Not applicable	28	Suter and Tsao (1996). Calculated using Great Lakes Water Quality Initiative Tier II methodology.

TABLE E-1
FRESHWATER TOXICITY REFERENCE VALUES
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Compound	Toxicity Value		Uncertainty Factor ^b	TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Concentration			
Crotonaldehyde	Acute LC50	3,500	0.01	35	Dawson et al. (1977). Toxicity value for bluegill sunfish (<i>Lepomis macrochirus</i>).
1,4-Dioxane	Acute EC0	6,210,000	0.01	62,100	Bringmann and Kühn (1982). Toxicity value for water flea (<i>D. magna</i>).
Formaldehyde	Acute LC50	4,960	0.01	49.6	Reardon and Harrell (1990). No data available for formaldehyde. Formalin containing 37 percent formaldehyde used as a surrogate. Endpoint based on formaldehyde concentration.
Vinyl chloride	Subchronic LC100	388,000	0.01 ^e	3,880	Brown et al. (1977)
Other chlorinated organics (µg/L)					
Hexachlorobenzene	Proposed chronic criterion	3.68	Not applicable	3.68	U.S. EPA (1987)
Hexachlorobutadiene	Chronic LOEL	9.3	0.1	0.93	U.S. EPA (1987)
Hexachlorocyclopentadiene	Chronic LOEL	5.2	0.1	0.52	U.S. EPA (1987)
Pentachlorobenzene	Tier II value	0.47	Not applicable	0.47	U.S. EPA (1996). Calculated using Great Lakes Water Quality Initiative Tier II methodology.
Pentachlorophenol	Chronic criterion	15	Not applicable	15	U.S. EPA (1999). Value expressed as a function of pH and calculated as follows: TRV = exp(1.005(pH)-5.134). A pH of 7.8 is assumed to calculate the displayed value.
Pesticides (µg/L)					
4,4'-DDE	Acute LOEL	1,050	0.01 ^e	10.5	U.S. EPA (1987)
Heptachlor	Chronic criterion	0.0038	Not applicable	0.0038	U.S. EPA (1987)

TABLE E-1
FRESHWATER TOXICITY REFERENCE VALUES

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Compound	Toxicity Value		Uncertainty Factor ^b	TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Concentration			
Hexachlorophene	Subchronic NOEC	8.8	0.1	0.88	Call et al. (1989). Toxicity value for fathead minnow (<i>P. promelas</i>).
Inorganics (mg/L) ^f					
Aluminum	FCV	0.087	Not applicable	0.087	U.S. EPA (1988)
Antimony	Proposed chronic criterion	0.03	Not applicable	0.03	U.S. EPA (1987)
Arsenic (trivalent)	Chronic criterion	0.15	Not applicable	0.15	U.S. EPA (1999)
Barium	Tier II SCV	0.004	Not applicable	0.004	Suter and Tsao (1996). Calculated using Great Lakes Water Quality Initiative Tier II methodology.
Beryllium	Tier II SCV	0.00066	Not applicable	0.00066	Suter and Tsao (1996). Calculated using Great Lakes Water Quality Initiative Tier II methodology.
Cadmium	Chronic criterion	0.0022 (dissolved)	Not applicable	0.0022	U.S. EPA (1999). Value expressed as a function of water hardness and calculated as follows: $TRV = \exp(m_c[\ln(\text{hardness})] + b_c)$ where $m_c = 0.7852$ and $b_c = -2.715$. Criterion was converted to dissolved concentration using the following conversion factor: $1.101672 - [(\ln \text{hardness})(0.041838)]$. A assumed hardness of 100 mg/L and a conversion from mg/L to $\mu\text{g/L}$ were used to calculate the displayed value.
Chromium (hexavalent)	Chronic criterion	0.011	Not applicable	0.011	U.S. EPA (1999).
Copper	Chronic criterion	0.009 (dissolved)	Not applicable	0.009	U.S. EPA (1999). Value expressed as a function of water hardness and calculated as follows: $TRV = \exp(m_c[\ln(\text{hardness})] + b_c)$ where $m_c = 0.8545$ and $b_c = -1.702$. Criterion was converted to dissolved concentration using a conversion factor of 0.960. A assumed hardness of 100 mg/L and a conversion from mg/L to $\mu\text{g/L}$ were used to calculate the displayed value.

TABLE E-1
FRESHWATER TOXICITY REFERENCE VALUES

(Page 5 of 8)

Compound	Toxicity Value		Uncertainty Factor ^b	TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Concentration			
Total Cyanide	Chronic criterion	0.0052	Not applicable	0.0052	U.S. EPA (1999). This value is expressed as mg free cyanide (as CN)/L.
Lead	Chronic criterion	0.0025 (dissolved)	Not applicable	0.0025	U.S. EPA (1999). Value expressed as a function of water hardness and calculated as follows: $TRV = \exp(m_c[\ln(\text{hardness})] + b_c)$ where $m_c = 1.273$ and $b_c = -4.705$. Criterion was converted to dissolved concentration using the following conversion factor: $1.46203 - [(\ln \text{hardness})(0.145712)]$. A assumed hardness of 100 mg/L and a conversion from mg/L to $\mu\text{g/L}$ were used to calculate the displayed value.
Mercuric chloride	Chronic criterion	0.00077	Not applicable	0.00077	U.S. EPA (1999). This value was from data for inorganic mercury (II).
Methyl mercury	Tier II SCV	0.0000028	Not applicable	0.0000028	Suter and Tsao (1996). Calculated using Great Lakes Water Quality Initiative Tier II methodology.
Nickel	Chronic criterion	0.052 (dissolved)	Not applicable	0.052	U.S. EPA (1999). Value expressed as a function of water hardness and calculated as follows: $TRV = \exp(m_c[\ln(\text{hardness})] + b_c)$ where $m_c = 0.8460$ and $b_c = 0.0584$. Criterion was converted to dissolved concentration using a conversion factor of 0.997. A assumed hardness of 100 mg/L and a conversion from mg/L to $\mu\text{g/L}$ were used to calculate the displayed value.
Selenium	Chronic criterion	0.005	Not applicable	0.005	U.S. EPA (1999)
Silver	Proposed chronic criterion	0.00012	Not applicable	0.00012	U.S. EPA (1987)
Thallium	Chronic LOEL	0.04	0.1	0.004	U.S. EPA (1987)

TABLE E-1

FRESHWATER TOXICITY REFERENCE VALUES

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Compound	Toxicity Value		Uncertainty Factor ^b	TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Concentration			
Zinc	Chronic criterion	0.118 (dissolved)	Not applicable	0.118	U.S. EPA (1999). Value expressed as a function of water hardness and calculated as follows: $TRV = \exp(m_c[\ln(\text{hardness})] + b_c)$ where $m_c = 0.8473$ and $b_c = 0.884$. Criterion was converted to dissolved concentration using a conversion factor of 0.986. A assumed hardness of 100 mg/L and a conversion from mg/L to $\mu\text{g/L}$ were used to calculate the displayed value.

Notes:

- a The duration of exposure is defined as chronic if it represents about 10 percent or more of the test animals lifetime expectancy. Acute exposures represent single exposures or multiple exposures occurring within a short time. For evaluating exposure duration, the following general guidelines were used. For invertebrates and other lower trophic level aquatic biota: (1) chronic duration lasted for 7 or more days, (2) subchronic duration lasted from 3 to 6 days, and (3) acute duration lasted 2 days or less. For fish: (1) chronic duration lasted for more than 90 days, (2) subchronic duration lasted from 14 to 90 days, and (3) acute duration lasted less than 2 weeks.
- b Uncertainty factors are used to extrapolate a toxicity value to a chronic NOAEL TRV. See Chapter 5 (Section 5.4) of the SLERAP for a discussion of the use of uncertainty factors.
- c TRV was calculated by multiplying the toxicity value with the uncertainty factor.
- d The references refer to the source of the toxicity value. Complete reference citations are provided below.
- e Best scientific judgment used to identify uncertainty factor. See Chapter 5 (Section 5.4.1.2) for a discussion the use of best scientific judgement. Factors evaluated include test duration, ecological relevance of endpoint, experimental design, and availability of toxicity data.
- f TRVs for metals are based on the dissolved metal concentration. According to U.S. EPA (1993) policy, concentrations of dissolved metal more closely approximate the bioavailable fraction of metal in the water column.

EC0	=	Effective concentration for zero percent of the test organisms.
FCV	=	Final Chronic Value
HMW	=	High molecular weight
LC50	=	Lethal concentration for 50 percent of the test organisms.
LC100	=	Lethal concentration for 100 percent of the test organisms.
LOEL	=	Lowest Observed Effect Level
NOEC	=	No Observed Effect Concentration
NOEL	=	No Observed Effect Level
SCV	=	Secondary Chronic Value
TRV	=	Toxicity Reference Value

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FRESHWATER TOXICITY REFERENCE VALUES

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TABLE E-2

MARINE/ESTUARINE SURFACE WATER TOXICITY REFERENCE VALUES

(Page 1 of 8)

Compound	Toxicity Value		Uncertainty Factor ^b	Toxicity Reference Value ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Concentration			
Polychlorinated dibenzo-p-dioxins (μg/L)					
2,3,7,8-TCDD	LOEC	0.000038	0.1	0.0000038	No saltwater data were available, therefore, corresponding freshwater toxicity value was used (rainbow trout, <i>Oncorhynchus mykiss</i>) from Mehrle et al. (1988). 2,3,4,5-TCDD toxicity value used.
Polynuclear aromatic hydrocarbons (PAH) (μg/L)					
Total high molecular weight (HMW) PAHs	Acute LC50	>50	0.01 ^e	0.5	Rossi and Neff (1978) evaluated toxicity of three HMW (three or more aromatic rings) PAHs to the polychaete, <i>Neanthes arenaceodentata</i> . LC50 of each HMW PAH exceeded 50 μg/L. This TRV should be used if assessing the risk of total HMW PAHs.
Benzo(a)pyrene	Acute LC50	>50	0.01 ^e	0.5	Rossi and Neff (1978). Toxicity value for polychaete (<i>N. arenaceodentata</i>).
Benzo(a)anthracene	Acute LC50	>50	0.01 ^e	0.5	Toxicity value not available. TRV for benzo(a)pyrene used as surrogate.
Benzo(b)fluoranthene	Acute LC50	>50	0.01 ^e	0.5	Toxicity value not available. TRV for benzo(a)pyrene used as surrogate.
Benzo(k)fluoranthene	Acute LC50	>50	0.01 ^e	0.5	Toxicity value not available. TRV for benzo(a)pyrene used as surrogate.
Chrysene	Acute LC50	>50	0.01 ^e	0.5	Rossi and Neff (1978). Toxicity of several PAHs was evaluated. LC50 of each individual HMW PAH exceeded 50 μg/L.
Dibenz(a,h)anthracene	Acute LC50	>50	0.01 ^e	0.5	Rossi and Neff (1978). Toxicity of several PAHs was evaluated. LC50 of individual HMW PAHs exceeded 50 μg/L.
Indeno(1,2,3-cd)pyrene	Acute LC50	>50	0.01 ^e	0.5	Toxicity value not available. TRV for benzo(a)pyrene used as surrogate.
Polychlorinated biphenyls (PCB) (μg/L)					
Aroclor 1016	--	0.03	Not applicable	0.03	U.S. EPA (1987) chronic criterion for ambient water quality.

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MARINE/ESTUARINE SURFACE WATER TOXICITY REFERENCE VALUES

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Compound	Toxicity Value		Uncertainty Factor ^b	Toxicity Reference Value ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Concentration			
Aroclor 1254	--	0.03	Not applicable	0.03	U.S. EPA (1987) chronic criterion for ambient water quality.
Nitroaromatics ($\mu\text{g/L}$)					
1,3-Dinitrobenzene	--	--	--	66.8	Toxicity data not available. TRV for nitrobenzene used as surrogate.
2,4-Dinitrotoluene	Chronic criterion	370	Not applicable	370	U.S. EPA (1987)
2,6-Dinitrotoluene	--	--	--	370	Toxicity data not available. TRV for 2,4-dinitrotoluene used as surrogate.
Nitrobenzene	Acute criterion	6,680	0.01	66.8	U.S. EPA (1987)
Pentachloronitrobenzene	Acute LC50	1,000	0.01	10	No toxicity value or surrogate TRV available, therefore, corresponding freshwater toxicity value (common carp, <i>Cyprinus carpio</i>) from Hashimoto and Nishiuchi (1981) adopted.
Phthalate esters ($\mu\text{g/L}$)					
Bis(2-ethylhexyl)phthalate	Acute LC50	>170	0.01	1.7	Adams et al. (1995). Toxicity value for sheepshead minnow (<i>Cyprinodon variegatus</i>).
Di(n)octyl phthalate	NOEL	320	Not applicable	320	No toxicity value or surrogate TRV available, therefore, corresponding freshwater toxicity value used (water flea, <i>D. magna</i>) from McCarthy and Whitmore (1985).
Volatile organic compounds ($\mu\text{g/L}$)					
Acetone	Acute LC50	2,100,000	0.01	21,000	Price et al. (1974). Toxicity value for brine shrimp (<i>Artemia</i> sp.).
Acrylonitrile	Acute LC50	10,000	0.01	100	Portmann and Wilson (1971). Toxicity value for common shrimp (<i>Crangon crangon</i>).

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MARINE/ESTUARINE SURFACE WATER TOXICITY REFERENCE VALUES

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Compound	Toxicity Value		Uncertainty Factor ^b	Toxicity Reference Value ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Concentration			
Chloroform	Acute LC 50	18,000	0.01	180	Anderson and Luster (1980). Toxicity value for Rainbow trout (<i>Salmo gairdneri</i>).
Crotonaldehyde	Acute LC50	1,300	0.01	13	Dawson et al. (1977). Toxicity value for inland silverside (<i>Menidia beryllina</i>).
1,4-Dioxane	Acute LC50	6,700,000	0.01	67,000	Dawson et al. (1977). Toxicity value for inland silverside (<i>M. beryllina</i>).
Formaldehyde	Acute LC50	4,960	0.01	49.6	No toxicity value or surrogate TRV available for this constituent, therefore, corresponding freshwater toxicity value used (Striped bass, <i>Morone saxatilis</i>) from Reardon and Harell (1990). No data available for formaldehyde. Formalin containing 37 percent formaldehyde used as surrogate. TRV expressed on formaldehyde basis.
Vinyl chloride	Subchronic LC100	388,000	0.01 ^e	3,880	No toxicity value of surrogate TRV available, therefore, corresponding freshwater toxicity value used (Northern pike, <i>Esox lucius</i>) from Brown et al. (1977).
Other chlorinated organics ($\mu\text{g/L}$)					
Hexachlorobenzene	Acute EC50	>1,000	0.01	10	Zaroogian (1981). Toxicity value for American oyster (<i>Crassostrea virginica</i>).
Hexachlorobutadiene	Acute LOEL	32	0.01 ^e	0.32	U.S. EPA (1987)
Hexachlorocyclopentadiene	Acute LOEL	7.0	0.01 ^e	0.07	U.S. EPA (1987)
Pentachlorobenzene	Subchronic NOEC	18	0.1	1.8	Hansen and Cripe (1991). Toxicity value for sheepshead minnow (<i>Cyprinodon variegatus</i>).
Pentachlorophenol	Chronic criterion	7.9	Not applicable	7.9	U.S. EPA (1987)
Pesticides ($\mu\text{g/L}$)					

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MARINE/ESTUARINE SURFACE WATER TOXICITY REFERENCE VALUES

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Compound	Toxicity Value		Uncertainty Factor ^b	Toxicity Reference Value ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Concentration			
4,4'-DDE	Acute LOEL	14	0.01 ^e	0.14	U.S. EPA (1987)
Heptachlor	Chronic criterion	0.0036	Not applicable	0.0036	U.S. EPA (1987)
Hexachlorophene	Acute LC50	3.3	0.01	0.033	Calleja et al. (1994). Toxicity value for brine shrimp (<i>Artemia salina</i>).
Inorganics (mg/L)					
Aluminum	Acute LT50	0.271	0.01	0.00271	Study examined influence of pH and temperature on acute (48-hour) toxicity (as time to mortality) of aluminum to smoltifying Atlantic salmon (<i>Salmo salar</i>). Endpoint concentration based on sum of inorganic and organic aluminum for exposure at pH 6.5 (Poleo and Muniz 1993).
Antimony	Proposed chronic criterion	0.5	Not applicable	0.5	U.S. EPA (1987)
Arsenic (trivalent)	Chronic criterion	0.036	Not applicable	0.036	U.S. EPA (1987)
Barium	Subchronic LC50	>500.	0.01 ^e	5.0	U.S. EPA (1978)
Beryllium	Tier II SCV	0.00066	Not applicable	0.00066	No toxicity value or surrogate TRV available, therefore, corresponding freshwater TRV adopted. Suter and Tsao (1996); value calculated using Great Lakes Water Quality Initiative Tier II methodology.
Cadmium	Chronic criterion	0.0093	Not applicable	0.0093	U.S. EPA (1987)
Chromium (hexavalent)	Chronic criterion	0.05	Not applicable	0.05	U.S. EPA (1987)
Copper	Chronic criterion	0.0031	Not applicable	0.0031	U.S. EPA 1999. When the concentration of dissolved organic carbon is elevated, copper is substantially less toxic and use of a water effects ratio may be appropriate.

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MARINE/ESTUARINE SURFACE WATER TOXICITY REFERENCE VALUES

(Page 5 of 8)

Compound	Toxicity Value		Uncertainty Factor ^b	Toxicity Reference Value ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Concentration			
Total Cyanide	Chronic criterion	0.001	Not applicable	0.001	U.S. EPA (1987)
Lead	Chronic criterion	0.0081	Not applicable	0.0081	U.S. EPA (1999)
Mercuric chloride	Chronic criterion	0.00094	Not applicable	0.00094	U.S. EPA (1999). This value was from data for inorganic mercury (II).
Methyl mercury	Subchronic NOAEL	0.030	0.1	0.003	Sharp and Neff (1982). Toxicity value for mummichog (<i>Fundulus heteroclitus</i>).
Nickel	Chronic criterion	0.0082	Not applicable	0.0082	U.S. EPA (1999)
Selenium	Chronic criterion	0.071	Not applicable	0.071	U.S. EPA (1987)
Silver	Chronic criterion/ proposed criterion	0.0023	Not applicable	0.0023	U.S. EPA (1987)
Thallium	Acute LOEL	2.13	0.01 ^e	0.02	U.S. EPA (1987)
Zinc	Chronic criterion	0.081	1.0	0.081	U.S. EPA (1999)

TABLE E-2

MARINE/ESTUARINE SURFACE WATER TOXICITY REFERENCE VALUES

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Notes:

- a The duration of exposure is defined as chronic if it represents about 10 percent or more of the test animals lifetime expectancy. Acute exposures represent single exposures or multiple exposures occurring within a short time. For evaluating exposure duration, the following general guidelines were used. For invertebrates and other lower trophic level aquatic biota: (1) chronic duration lasted for 7 or more days, (2) subchronic duration lasted from 3 to 6 days, and (3) acute duration lasted 2 days or less. For fish: (1) chronic duration lasted for more than 90 days, (2) subchronic duration lasted from 14 to 90 days, and (3) acute duration lasted less than 2 weeks.
- b Uncertainty factors are used to extrapolate a toxicity value to a chronic NOAEL TRV. See Chapter 5 (Section 5.4) of the SLERAP for a discussion of the use of uncertainty factors.
- c TRV was calculated by multiplying the toxicity value with the uncertainty factor.
- d The references refer to the source of the toxicity value. Complete reference citations are provided at the end of this appendix.
- e Best scientific judgment used to identify uncertainty factor. See Chapter 5 (Section 5.4.1.2) for a discussion of the use of best scientific judgement. Factors evaluated include test duration, ecological relevance of endpoint, experimental design, and availability of toxicity data.

EC50	=	Effective concentration for 50 percent of the test organisms.
FCV	=	Final Chronic Values
HMV	=	High molecular weight
LC50	=	Lethal concentration for 50 percent of the test organisms.
LC100	=	Lethal concentration for 100 percent of the test organisms.
LOEC	=	Lowest Observed Effect Concentration
LOEL	=	Lowest Observed Effect Level
LT50	=	Lethal threshold concentration for 50 percent of the test organisms.
NOAEL	=	No Observed Adverse Effect Level
NOEL	=	No Observed Effect Level
SCV	=	Secondary Chronic Value
TRV	=	Toxicity Reference Value

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TABLE E-3
FRESHWATER SEDIMENT TOXICITY REFERENCE VALUES

(Page 1 of 7)

Compound	Freshwater TRV ^a	K _{oc} Value ^b	Bed Sediment TRV (dry weight)	Reference and Notes ^c
Polychlorinateddibenzo-p-dioxins ($\mu\text{g/kg}$)				
2,3,7,8-TCDD	0.0000038	2,691,535	0.41	TRV was calculated using equilibrium partitioning (EqP) approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Polynuclear aromatic hydrocarbons (PAH) ($\mu\text{g/kg}$)				
Total high molecular weight (HMW) PAH	Not applicable	Not applicable	170	TRV is ERL value computed by Ingersoll et al. (1996) based on 28-day amphipod (<i>Hyalella azteca</i>) toxicity tests. This TRV may be used if risk of total HMW PAHs is assessed.
Benzo(a)pyrene	Not applicable	Not applicable	84	TRV is an ERL value calculated by Ingersoll et al. (1996) based on 28-day <i>H. azteca</i> toxicity tests.
Benzo(a)anthracene	Not applicable	Not applicable	19	TRV is an ERL value calculated by Ingersoll et al. (1996) based on 28-day <i>H. azteca</i> toxicity tests.
Benzo(b)fluoranthene	Not applicable	Not applicable	37	TRV is an ERL value calculated by Ingersoll et al. (1996) based on 28-day <i>H. azteca</i> toxicity tests.
Benzo(k)fluoranthene	Not applicable	Not applicable	37	TRV is an ERL value calculated by Ingersoll et al. (1996) based on 28-day <i>H. azteca</i> toxicity tests.
Chrysene	Not applicable	Not applicable	30	TRV is an ERL value calculated by Ingersoll et al. (1996) based on 28-day <i>H. azteca</i> toxicity tests.
Dibenz(a,h)anthracene	Not applicable	Not applicable	10	TRV is an ERL value calculated by Ingersoll et al. (1996) based on 28-day <i>H. azteca</i> toxicity tests.
Indeno(1,2,3-cd)pyrene	Not applicable	Not applicable	30	TRV is an ERL value calculated by Ingersoll et al. (1996) based on 28-day <i>H. azteca</i> toxicity tests.

TABLE E-3
FRESHWATER SEDIMENT TOXICITY REFERENCE VALUES

(Page 2 of 7)

Compound	Freshwater TRV ^a	K _{oc} Value ^b	Bed Sediment TRV (dry weight)	Reference and Notes ^c
Polychlorinated biphenyls (PCB) ($\mu\text{g/kg}$)				
Aroclor 1016	Not applicable	Not applicable	50	TRV is an ERL value for Total PCB calculated by Ingersoll et al. (1996) based on 28-day <i>H. azteca</i> toxicity tests.
Aroclor 1254	Not applicable	Not applicable	50	TRV is an ERL value for Total PCB calculated by Ingersoll et al. (1996) based on 28-day <i>H. azteca</i> toxicity tests.
Nitroaromatics ($\mu\text{g/kg}$)				
1,3-Dinitrobenzene	26	20.6	21.4	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
2,4-Dinitrotoluene	23	51	46.9	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
2,6-Dinitrotoluene	60	41.9	100.6	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Nitrobenzene	270	119	1285.2	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Pentachloronitrobenzene	10	5,890	2356	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Phthalate esters ($\mu\text{g/kg}$)				
Bis(2-ethylhexyl)phthalate	3	111,000	1.33×10^{-4}	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Di(n)octyl phthalate	320	9.03×10^{-8}	1.16×10^{-10}	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d

TABLE E-3
FRESHWATER SEDIMENT TOXICITY REFERENCE VALUES
(Page 3 of 7)

Compound	Freshwater TRV ^a	K _{oc} Value ^b	Bed Sediment TRV (dry weight)	Reference and Notes ^c
Volatile organic compounds (μg/kg)				
Acetone	1,500	0.951	57.1	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Acrylonitrile	260	2.22	23.1	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Chloroform	28	53.0	59.4	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Crotonaldehyde	35	Not available	Not calculated	No TRV was calculated because no K _{oc} or K _{ow} values were identified for this constituent.
1,4-Dioxane	62,100	0.876	2176.0	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Formaldehyde	49.6	2.62	5.2	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Vinyl chloride	3,880	11.1	1722.7	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Other chlorinated organics (μg/kg)				
Hexachlorobenzene	Not applicable	Not applicable	20	TRV is an LEL value (Persaud et al. 1993).
Hexachlorobutadiene	0.93	6,940	258.2	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Hexachlorocyclopentadiene	0.52	9,510	197.8	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d

TABLE E-3
FRESHWATER SEDIMENT TOXICITY REFERENCE VALUES

(Page 4 of 7)

Compound	Freshwater TRV ^a	K _{oc} Value ^b	Bed Sediment TRV (dry weight)	Reference and Notes ^c
Pentachlorobenzene	0.47	32,148	604.4	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Pentachlorophenol	Not applicable	Not applicable	7,000	TRV is an AET value for <i>H. azteca</i> (Washington State Department of Ecology 1994).
Pesticides ($\mu\text{g/kg}$)				
4,4'-DDE	Not applicable	Not applicable	5	TRV is an LEL value (Persaud et al. 1993). p,p'-DDE used as a surrogate.
Heptachlor	Not applicable	Not applicable	0.3	TRV is an NEL value (Persaud et al. 1993). The NEL was selected because no LEL was available.
Hexachlorophene	0.88	1,800,000	63,360	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Inorganics (mg/kg)				
Aluminum	Not applicable	Not applicable	14,000	TRV is an ERL value calculated by Ingersoll et al. (1996) based on 28-day <i>H. azteca</i> toxicity tests.
Antimony	Not applicable	Not applicable	64.0	TRV is an AET for <i>H. azteca</i> (Washington State Department of Ecology 1994).
Arsenic	Not applicable	Not applicable	6.0	TRV is an LEL value (Persaud et al. 1993).
Barium	Not applicable	Not applicable	20	TRV is a U.S. EPA Region 5 guideline value for classification of sediments for determining the suitability of dredged sediments for open water disposal, as cited in Hull and Suter II (1994).
Beryllium	Not applicable	Not applicable	Not available	Regulatory or toxicity value not available.
Cadmium	Not applicable	Not applicable	0.6	TRV is an LEL value (Persaud et al. 1993).

TABLE E-3
FRESHWATER SEDIMENT TOXICITY REFERENCE VALUES

(Page 5 of 7)

Compound	Freshwater TRV ^a	K _{oc} Value ^b	Bed Sediment TRV (dry weight)	Reference and Notes ^c
Chromium (total)	Not applicable	Not applicable	26	TRV is an LEL value (Persaud et al. 1993).
Copper	Not applicable	Not applicable	16	TRV is an LEL value (Persaud et al. 1993).
Total Cyanide	Not applicable	Not applicable	0.1	TRV is a U.S. EPA Region 5 guideline value for classification of sediments for determining the suitability of dredged sediments for open water disposal, as cited in Hull and Suter II (1994).
Lead	Not applicable	Not applicable	31	TRV is an LEL value (Persaud et al. 1993).
Mercuric chloride	Not applicable	Not applicable	0.2	No toxicity data available for divalent inorganic mercury. Total mercury used as surrogate for divalent inorganic mercury. TRV is an LEL value (Persaud et al. 1993).
Methyl mercury	Not applicable	Not applicable	0.2	No toxicity data available for methyl mercury. Total mercury used as surrogate for methylmercury. TRV is an LEL value (Persaud et al. 1993).
Nickel	Not applicable	Not applicable	16	TRV is an LEL value (Persaud et al. 1993).
Selenium	Not applicable	Not applicable	0.1	TRV is an AET for <i>H. azteca</i> (Washington State Department of Ecology 1994).
Silver	Not applicable	Not applicable	4.5	TRV is an AET for <i>H. azteca</i> (Washington State Department of Ecology 1994).
Thallium	Not applicable	Not applicable	Not available	Regulatory value or toxicity value not available.
Zinc	Not applicable	Not applicable	110	TRV is an ERL value calculated by Ingersoll et al. (1996) based on 28-day <i>H. azteca</i> toxicity tests.

TABLE E-3
FRESHWATER SEDIMENT TOXICITY REFERENCE VALUES
(Page 6 of 7)

Notes:

- a Toxicity reference values are in units of micrograms per kilogram ($\mu\text{g/kg}$) and milligrams per kilograms (mg/kg) for organic and inorganic constituents, respectively.
- b Values are in units of liters per kilogram (L/kg). K_{oc} = Organic carbon normalized sorption coefficient. References and equations used to calculate K_{oc} values are provided in Appendix A.
- c The references refer to the study from which the TRV was identified. Complete reference citations are provided below.
- d Freshwater sediment TRV calculated with the following equation:

$$\text{Freshwater sediment TRV} = \text{Freshwater TRV (Table E-1)} * K_{oc} * f_{oc,bs}$$

where,
 K_{oc} = organic carbon partition coefficient, and
 $f_{oc,bs}$ = fraction of organic carbon in bed sediment, assumed to be 4 percent = 0.04.

K_{oc} values discussed in Appendix A.

AET	=	Apparent Effects Threshold
ERL	=	Effects Range-Low
EqP	=	Equilibrium Partitioning
HMV	=	High molecular weight
LEL	=	Lowest Effect Level
NEL	=	No Effect Level
TRV	=	Toxicity Reference Value

TABLE E-3

FRESHWATER SEDIMENT TOXICITY REFERENCE VALUES

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REFERENCES

Default TRVs for sediments in freshwater habitats were identified from the three sets of freshwater toxicity values presented below. While some compound-specific freshwater sediment toxicity information is available in the scientific literature, available toxicity values were not used because of the complexity in understanding the role of naturally-occurring sediment features (such as grain size, ammonia, sulfide, soil type, and organic carbon content) in toxicity to benthic invertebrates. Among these sets of value, the lowest available toxicity value for a particular compound was adopted as the TRV. In many cases, a default TRV was calculated from the corresponding freshwater TRV using EPA's equilibrium partitioning approach, assuming a 4 percent organic carbon content.

Hull, R.N. and G.W. Suter II. 1994. *Toxicological Benchmarks for Screening Contaminants of Potential Concern for Effects on Sediment-Associated Biota: 1994 Revision*. ES/ER/TM-95/R1. Environmental Sciences Division, Oak Ridge National Laboratory. Oak Ridge, Tennessee. June.

Ingersoll, C.G., P.S. Haverland, E.L. Brunson, T.J. Canfield, F.J. Dwyer, C.E. Henke, N.E. Kemble, D.R. Mount, and R.G. Fox. 1996. "Calculation and Evaluation of Sediment Effect Concentrations for the Amphipod *Hyalella azteca* and the Midge *Chironomus riparius*." *International Association of Great Lakes Research*. Volume 22. Pages 602-623.

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U.S. EPA. 1993. *Technical Basis for Deriving Sediment Quality Criteria for Nonionic Organic Contaminants for the Protection of Benthic Organisms by Using Equilibrium Partitioning*. Office of Water. EPA-822-R-93-011. September.

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TABLE E-4

MARINE/ESTUARINE SEDIMENT TOXICITY REFERENCE VALUES

(Page 1 of 8)

Compound	Marine/Estuarine Surface Water TRV ^a	K _{oc} Value ^b	Bed Sediment TRV (dry weight)	Reference and Notes ^c
Ploychlorinateddibenzo-p-dioxins ($\mu\text{g/kg}$)				
2,3,7,8-TCDD	0.0000038	2,691,535	0.41	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Polynuclear aromatic hydrocarbons (PAH) ($\mu\text{g/kg}$)				
Total high molecular weight (HMW) PAH	Not applicable	Not applicable	870	Recommended NOEL for Florida Department of Environmental Regulation (DER) (MacDonald 1993). This TRV may be used in risk of total HMW PAHs is assessed.
Benzo(a)pyrene	Not applicable	Not applicable	230	Recommended NOEL for Florida DER (MacDonald 1993).
Benzo(a)anthracene	Not applicable	Not applicable	160	Recommended NOEL for Florida DER (MacDonald 1993).
Benzo(b)fluoranthene	0.5	836,000	418,000	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Benzo(k)fluoranthene	Not applicable	Not applicable	240	TRV is a LEL value from Persaud et al. (1993).
Chrysene	Not applicable	Not applicable	220	Recommended NOEL for Florida DER (MacDonald 1993).
Dibenz(a,h)anthracene	Not applicable	Not applicable	31	Recommended NOEL for Florida DER (MacDonald 1993).
Indeno(1,2,3-cd)pyrene	Not applicable	Not applicable	1,360	TRV was computed from OC-based marine sediment quality criterion from Washington State Department of Ecology (1991) and fractional organic carbon content of 0.04, as follows: $\text{TRV} = 34 \text{ mg/kg} * 0.04 * 1000 \mu\text{g/mg}$.

TABLE E-4

MARINE/ESTUARINE SEDIMENT TOXICITY REFERENCE VALUES

(Page 2 of 8)

Compound	Marine/Estuarine Surface Water TRV ^a	K _{oc} Value ^b	Bed Sediment TRV (dry weight)	Reference and Notes ^c
Polychlorinated biphenyls (PCB) ($\mu\text{g/kg}$)				
Aroclor 1016	Not applicable	Not applicable	22.7	TRV is an ERL value for Total PCB from Long et al. (1995).
Aroclor 1254	Not applicable	Not applicable	22.7	TRV is an ERL value for Total PCB from Long et al. (1995).
Nitroaromatics ($\mu\text{g/kg}$)				
1,3-Dinitrobenzene	66.8	20.6	55.0	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
2,4-Dinitrotoluene	370	51	754.8	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
2,6-Dinitrotoluene	370	41.9	620.1	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Nitrobenzene	66.8	119	318.0	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Pentachloronitrobenzene	10	5,890	2356	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d

TABLE E-4

MARINE/ESTUARINE SEDIMENT TOXICITY REFERENCE VALUES

(Page 3 of 8)

Compound	Marine/Estuarine Surface Water TRV ^a	K _{oc} Value ^b	Bed Sediment TRV (dry weight)	Reference and Notes ^c
Phthalate esters (μg/kg)				
Bis(2-ethylhexyl)phthalate	Not applicable	Not applicable	470	TRV was calculated using OC-based marine sediment quality criterion from Washington State Department of Ecology (1991) and fractional organic carbon content of 0.04, as follows: TRV = 47 mg/kg * 0.04 * 1000 μg/mg.
Di(n)octyl phthalate	Not applicable	Not applicable	580	TRV was calculated using OC-based marine sediment quality criterion from Washington State Department of Ecology (1991) and fractional organic carbon content of 0.04, as follows: TRV = 58 mg/kg * 0.04 * 1000 μg/mg.
Volatile organic compounds (μg/kg)				
Acetone	21,000	0.951	798.8	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Acrylonitrile	100	2.22	8.88	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Chloroform	180	53.0	381.6	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Crotonaldehyde	13	Not available	Not computed	No TRV was calculated because no K _{oc} or K _{ow} value was identified.
1,4-Dioxane	67,000	0.876	2348	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Formaldehyde	49.6	2.62	5.2	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d

TABLE E-4

MARINE/ESTUARINE SEDIMENT TOXICITY REFERENCE VALUES

(Page 4 of 8)

Compound	Marine/Estuarine Surface Water TRV ^a	K _{oc} Value ^b	Bed Sediment TRV (dry weight)	Reference and Notes ^c
Vinyl chloride	3,880	11.1	1722.7	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Other chlorinated organics (μg/kg)				
Hexachlorobenzene	Not applicable	Not applicable	15.2	TRV was calculated using OC-based marine sediment quality criterion from Washington State Department of Ecology (1991) and a fractional OC content of 0.04, as follows: TRV = 0.38 mg/kg * 0.04 * 1000 μg/mg.
Hexachlorobutadiene	Not applicable	Not applicable	156	TRV was calculated using OC-based marine sediment quality criterion from Washington State Department of Ecology (1991) and a fractional OC content of 0.04, as follows: TRV = 3.9 mg/kg * 0.04 * 1000 μg/mg.
Hexachlorocyclopentadiene	0.07	9,510	26.6	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Pentachlorobenzene	1.8	32,148	2315	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Pentachlorophenol	Not applicable	Not applicable	360	TRV is marine sediment quality criterion from Washington State Department of Ecology (1991).
Pesticides (μg/kg)				
4,4'-DDE	Not applicable	Not applicable	1.7	Recommended NOEL for p,p'-DDE for Florida DER (MacDonald 1993).
Heptachlor	0.0036	9,530	1.37	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d
Hexachlorophene	0.033	1,800,000	2376	TRV was calculated using EqP approach (EPA 1993), assuming a fractional organic content of 0.04. ^d

TABLE E-4

MARINE/ESTUARINE SEDIMENT TOXICITY REFERENCE VALUES

(Page 5 of 8)

Compound	Marine/Estuarine Surface Water TRV ^a	K _{oc} Value ^b	Bed Sediment TRV (dry weight)	Reference and Notes ^c
Inorganics (mg/kg)				
Aluminum	Not applicable	Not applicable	Not available	Screening or toxicity value not available.
Antimony	Not applicable	Not applicable	2	TRV is an ERL value (Long and Morgan 1991).
Arsenic	Not applicable	Not applicable	6	TRV is an LEL value for Province of Ontario (Persaud et al. 1993).
Barium	Not applicable	Not applicable	20	TRV is a U.S. EPA Region 5 guideline value for classification of sediments for determining the suitability of dredged material for open water disposal, as cited in Hull and Suter II (1994).
Beryllium	Not applicable	Not applicable	Not available	Screening or toxicity value not available.
Cadmium	Not applicable	Not applicable	1.0	Recommended NOEL for Florida DER (MacDonald 1993).
Chromium (hexavalent)	Not applicable	Not applicable	8.1	TRV is an ERL value for total chromium (Long et al. 1995).
Copper	Not applicable	Not applicable	28	Recommended NOEL for Florida DER (MacDonald 1993).
Total Cyanide	Not applicable	Not applicable	0.1	TRV is a U.S. EPA Region V guideline value for classification of sediments for determining the suitability of dredged material for open water disposal, as cited in Hull and Suter II (1994).

TABLE E-4

MARINE/ESTUARINE SEDIMENT TOXICITY REFERENCE VALUES

(Page 6 of 8)

Compound	Marine/Estuarine Surface Water TRV ^a	K _{oc} Value ^b	Bed Sediment TRV (dry weight)	Reference and Notes ^c
Lead	Not applicable	Not applicable	21.0	Recommended NOEL for Florida DER (MacDonald 1993).
Mercuric chloride	Not applicable	Not applicable	0.1	No toxicity data available for divalent inorganic mercury. Total mercury is used as surrogate. Recommended NOEL for Florida DER (MacDonald 1993).
Methyl mercury	Not applicable	Not applicable	0.1	No toxicity data available for methyl mercury. Total mercury is used as surrogate. Recommended NOEL for Florida DER (MacDonald 1993).
Nickel	Not applicable	Not applicable	20.9	TRV is an ERL value (Long et al. 1995).
Selenium	Not applicable	Not applicable	Not Available	Screening or toxicity value not available.
Silver	Not applicable	Not applicable	0.5	Recommended NOEL for Florida DER (MacDonald 1993).
Thallium	Not applicable	Not applicable	Not available	Screening or toxicity value not available.
Zinc	Not applicable	Not applicable	68	Recommended NOEL for Florida DER (MacDonald 1993).

TABLE E-4

MARINE/ESTUARINE SEDIMENT TOXICITY REFERENCE VALUES

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Notes:

- a Sediment TRVs are in units of micrograms per kilogram ($\mu\text{g/kg}$) and milligrams per kilograms (mg/kg) for organic and inorganic constituents, respectively.
- b Values are in units of liters per kilogram (L/kg). K_{oc} = Organic carbon normalized sorption coefficient. References and equations used to calculate values are provided in Appendix A.
- c The references refer to the study or studies from which the endpoint and concentrations were identified. Complete reference citations are provided below.
- d Sediment TRV calculated with the following equation:

$$\text{Sediment TRV} = \text{Marine/estuarine surface water TRV (Table E-2)} * K_{oc} * f_{oc,bs}$$

where,

K_{oc} = organic carbon partition coefficient, and
 $f_{oc,bs}$ = fraction of organic carbon in bed sediment, assumed to be 1 percent = 0.01.

K_{oc} values are discussed in Appendix A.

EqP	=	Equilibrium Partitioning
ERL	=	Effects Range-Low
HMW	=	High molecular weight
LEL	=	Lowest Effect Level
NOEL	=	No Observed Effect Level
TRV	=	Toxicity Reference Value

TABLE E-4

MARINE/ESTUARINE SEDIMENT TOXICITY REFERENCE VALUES

(Page 8 of 8)

REFERENCES

Default TRVs for sediments in marine and estuarine habitats were identified from several sets of toxicity values (standards, benchmarks, and guidelines) presented below. While some compound-specific marine/estuarine sediment toxicity information is available in the scientific literature, available toxicity values were not used because of the complexity in understanding the role of naturally-occurring sediment features (such as grain size, ammonia, sulfide, soil type, and organic carbon content) in toxicity to benthic invertebrates. Among these sets of value, the lowest available toxicity value for a particular compound was adopted as the TRV. In many cases, a default TRV was calculated from the corresponding freshwater TRV using EPA's equilibrium partitioning approach, assuming a 4 percent organic carbon content.

Hull, R.N. and G.W. Suter II. 1994. *Toxicological Benchmarks for Screening Contaminants of Potential Concern for Effects on Sediment-Associated Biota: 1994 Revision*. ES/ER/TM-95/R1. Environmental Sciences Division, Oak Ridge National Laboratory. Oak Ridge, Tennessee. June.

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Long, E.R., D.D. MacDonald, S.L. Smith, and F.D. Calder. 1995. "Incidence of Adverse Biological Effects Within Ranges of Chemical Concentrations in Marine and Estuarine Sediments." *Environmental Management*. Volume 19. Pages 81-97.

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Persaud, D., R. Jaaguagi, and A. Hayton. 1993. *Guidelines for the Protection and Management of Aquatic Sediment Quality in Ontario*. Ontario Ministry of the Environment. Queen's Printer of Ontario. March.

U.S. EPA. 1993. *Technical Basis for Deriving Sediment Quality Criteria for Nonionic Organic Contaminants for the Protection of Benthic Organisms by Using Equilibrium Partitioning*. Office of Water. EPA-822-R-93-011. September.

Washington State Department of Ecology. 1991. *Sediment Management Standards*. Washington Administrative Code 173-204.

TABLE E-5
TERRESTRIAL PLANT TOXICITY REFERENCE VALUES

(Page 1 of 15)

Compound	Basis for TRV				TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Concentration	Uncertainty Factor ^b		
Polychlorinateddibenzo-p-dioxins (μg/kg)						
2,3,7,8-TCDD	--	--	--	--	--	Toxicity value not identified.
Polynuclear aromatic hydrocarbons (PAH) (μg/kg)						
Total high molecular weight (HMW) PAH	Chronic NOAEL	Wheat	1,200	Not applicable	1,200	Benzo(a)pyrene toxicity used as representative toxicity of all HMW PAHs. This TRV may be used to characterize risk of total HMW PAHs to terrestrial plants.
Benzo(a)pyrene	Chronic NOAEL	Wheat	1,200	Not applicable	1,200	Sims and Overcash (1983)
Benzo(a)anthracene	Not available	--	--	--	1,200	Toxicity value not available. Benzo(a)pyrene used as surrogate.
Benzo(b)fluoranthene	Chronic NOAEL	Wheat	1,200	Not applicable	1,200	Sims and Overcash (1983).
Benzo(k)fluoranthene	Not available	--	--	--	1,200	Toxicity value not available. Benzo(a)pyrene used as surrogate.
Chrysene	Not available	--	--	--	1,200	Toxicity value not available. Benzo(a)pyrene used as surrogate.
Dibenz(a,h)anthracene	Not available	--	--	--	1,200	Toxicity value not available. Benzo(a)pyrene used as surrogate.

TABLE E-5
TERRESTRIAL PLANT TOXICITY REFERENCE VALUES

(Page 2 of 15)

Compound	Basis for TRV				TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Concentration	Uncertainty Factor ^b		
Indeno(1,2,3-cd)pyrene	Not available	--	--	--	1,200	Toxicity value not available. Benzo(a)pyrene used as surrogate.
Polychlorinated biphenyls (PCB) ($\mu\text{g/kg}$)						
Aroclor 1016	--	--	--	--	10,000	No toxicity value available. Aroclor 1254 TRV adopted as surrogate.
Aroclor 1254	Chronic NOAEL	Soybean shoot weight	10,000	Not applicable	10,000	Value for toxicity of Aroclor 1254 (Weber and Mrozek 1979).
Nitroaromatics ($\mu\text{g/kg}$)						
1,3-Dinitrobenzene	--	--	--	--	--	Toxicity value not available.
2,4-Dinitrotoluene	--	--	--	--	--	Toxicity value not available.
2,6-Dinitrotoluene	--	--	--	--	--	Toxicity value not available.
Nitrobenzene	--	--	--	--	--	Toxicity value not available.
Pentachloronitrobenzene	--	--	--	--	--	Toxicity value not available.
Phthalate esters ($\mu\text{g/kg}$)						
Bis(2-ethylhexyl)phthalate	--	--	--	--	--	Toxicity value not available.
Di(n)octyl phthalate	--	--	--	--	--	Toxicity value not available.
Volatile organic compounds ($\mu\text{g/kg}$)						

TABLE E-5
TERRESTRIAL PLANT TOXICITY REFERENCE VALUES

(Page 3 of 15)

Compound	Basis for TRV				TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Concentration	Uncertainty Factor ^b		
Acetone	--	--	--	--	--	Toxicity value not available.
Acrylonitrile	--	--	--	--	--	Toxicity value not available.
Chloroform	--	--	--	--	--	Toxicity value not available.
Crotonaldehyde	--	--	--	--	--	Toxicity value not available.
1,4-Dioxane	--	--	--	--	--	Toxicity value not available.
Formaldehyde	--	--	--	--	--	Toxicity value not available.
Vinyl chloride	--	--	--	--	--	Toxicity value not available.
Other chlorinated organics ($\mu\text{g/kg}$)						
Hexachlorobenzene	--	--	--	--	--	Toxicity value not available.
Hexachlorobutadiene	--	--	--	--	--	Toxicity value not available.
Hexachlorocyclopentadiene	Acute EC50	Lettuce growth	10,000	0.01	100	Hulzebos et al. (1993)
Pentachlorobenzene	--	--	--	--	--	Toxicity value not available.
Pentachlorophenol	Chronic LOAEL	Rice	17,300	0.1	1,730	Nagasawa et al. (1981)
Pesticides ($\mu\text{g/kg}$)						
4,4'-DDE	--	--	--	--	--	Toxicity value not available.

TABLE E-5
TERRESTRIAL PLANT TOXICITY REFERENCE VALUES

(Page 4 of 15)

Compound	Basis for TRV				TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Concentration	Uncertainty Factor ^b		
Heptachlor	Chronic NOAEL	Carrot	1,000	Not applicable	1,000	Ahrens and Kring (1968)
Hexachlorophene	--	--	--	--	--	Toxicity value not available.
Inorganics (mg/kg)						
Aluminum	Subchronic NOAEL	White clover seedling establishment	50	0.1 ^e	5	Mackay et al. (1990)
Antimony	Not specified	Not specified	5	0.1 ^e	0.5	Kabata-Pendias and Pendias (1992)
Arsenic	Chronic LOAEL	Corn yield (weight)	10	0.1	1	Woolson et al. (1971)
Barium	Chronic LOAEL	Barley shoot growth	500	0.01 ^e	5	Chaudry et al. (1977)
Beryllium	Not specified	Not specified	10	0.01 ^e	0.1	Kabata-Pendias and Pendias (1992)
Cadmium	Chronic LOAEL	Spruce seedling growth	2	0.1 ^e	0.2	Burton et al. (1984)
Chromium (hexavalent)	Subchronic EC50	Lettuce growth	1.8	0.01	0.018	Adema and Hazen (1989)
Copper	Chronic LOAEL	Barley	10	0.1	1.0	Toivonem and Hofstra (1979)

TABLE E-5
TERRESTRIAL PLANT TOXICITY REFERENCE VALUES

(Page 5 of 15)

Compound	Basis for TRV				TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Concentration	Uncertainty Factor ^b		
Cyanide, total	--	--	--	--	--	Toxicity value not available.
Lead	Chronic LOAEL	Senna	46	0.1	4.6	Krishnayya and Bedi (1986)
Mercuric chloride	Acute NOEC	Barley	34.9	0.01 ^e	0.349	Panda et al. (1992)
Methyl mercury	--	--	--	--	--	Toxicity value not available.
Nickel	Chronic NOAEL	Bush bean shoot growth	25	Not applicable	25	Wallace et al. (1977)
Selenium	Subchronic NOAEL	Alfalfa shoot weight	0.5	0.1	0.05	Wan et al. (1988)
Silver	Not specified	Not specified	2	0.01 ^e	0.02	Kabata-Pendias and Pendias (1992)
Thallium	Not specified	Not specified	1	0.01 ^e	0.01	Kabata-Pendias and Pendias (1992)
Zinc	Chronic LOAEL	Spring barley	9	0.1	0.9	Davis, Beckett, and Wollan (1978)

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Notes:

- a To evaluate exposure duration, the following general guidelines were used: Chronic duration represents exposures occurring about 10 or more days, including exposure during a critical life stage, such as germination and shoot development. Subchronic duration generally lasts 2 days through several days, however a sensitive life stage is not exposed. Acute duration generally includes exposures occurring 0 to 2 days.
- b Uncertainty factors are used to extrapolate a toxicity value to a chronic NOAEL TRV. See Chapter 5 (Section 5.4) of the SLERAP for a discussion on the use of uncertainty factors.
- c TRV was calculated by multiplying the toxicity value with the uncertainty factor.
- d The references refer to the source of the toxicity value. Complete reference citations are provided below.
- e Best scientific judgment was used to identify uncertainty factor. See Chapter 5 (Section 5.4.1.2) for a discussion on the use of best scientific judgement. Factors evaluated include test duration, ecological relevance of endpoint, and experimental design.

EC50	=	Effective concentration for 50 percent of the test organisms.
HWC	=	High molecular weight
LOAEL	=	Lowest Observed Adverse Effects Level
NOAEL	=	No Observed Adverse Effects Level
NOEC	=	No Observed Effects Concentration
TRV	=	Toxicity Reference Value

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REFERENCES

Efroymson, Will, Suter II, and Wooten (1997) provides a comprehensive review of ecologically-relevant terrestrial plant toxicity information. This source was reviewed to identify studies to develop TRVs for terrestrial plant. Based on the information presented, one or more references were obtained and reviewed to identify compound-specific toxicity values. For some compounds, the available information identified a single study meeting the requirements for a TRV, as discussed in Chapter 5 (Section 5.4) of the SLERAP. In most cases, each reference was obtained and reviewed to identify a single toxicity value to develop a TRV for each compound. In a few cases where a primary study could not be obtained, a toxicity value is based on a secondary source. As noted below, additional compendia were reviewed to identify toxicity studies to review. For compounds not discussed in Efroymson, Will, Suter II, and Wooten (1997), the scientific literature was searched, and relevant studies were obtained and reviewed. The references reviewed are listed below. The study selected for the TRV is highlighted in bold.

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Benzo(k)fluoranthene

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Aroclor 1254

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TABLE E-6

SOIL INVERTEBRATE TOXICITY REFERENCE VALUES

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Compound	TRV				TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Test Species	Concentration	Uncertainty Factor ^b		
Polychlorinateddibenzo-p-dioxins (μg/kg)						
2,3,7,8-TCDD	Chronic (85-day); no mortality reported at 5,000 μg/kg	Earthworm (<i>Allolobophora caliginosa</i>)	5,000	0.1 ^e	500	Toxicity value for 2,3,7,8-TCDD (Reinecke and Nash 1984). UF applied to concentration because mortality only endpoint available and data not subjected to statistical analysis.
Polynuclear aromatic hydrocarbons (PAH) (μg/kg)						
Total HMW PAH	Not available	--	--	--	25,000	Benzo(a) pyrene used as surrogate for HMW PAH compounds.
Benzo(a)pyrene	Chronic (28-day) NOAEL for growth	Woodlouse (<i>Porcellio scaber</i>)	25,000	Not applicable	25,000	van Straalen and Verweij (1991)
Benzo(a)anthracene	Not available	--	--	--	25,000	Toxicity value not available. TRV for benzo(a)pyrene used as surrogate.
Benzo(b)fluoranthene	Not available	--	--	--	25,000	Toxicity value not available. TRV for benzo(a)pyrene used as surrogate.
Benzo(k)fluoranthene	Not available	--	--	--	25,000	Toxicity value not available. TRV for benzo(a)pyrene used as surrogate.
Chrysene	Not available	--	--	--	25,000	Toxicity value not available. TRV for benzo(a)pyrene used as surrogate.
Dibenz(a,h)anthracene	Not available	--	--	--	25,000	Toxicity value not available. TRV for benzo(a)pyrene used as surrogate.
Indeno(1,2,3-cd)pyrene	Not available	--	--	--	25,000	Toxicity value not available. TRV for benzo(a)pyrene used as surrogate.

TABLE E-6
SOIL INVERTEBRATE TOXICITY REFERENCE VALUES

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Compound	TRV				TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Test Species	Concentration	Uncertainty Factor ^b		
Polychlorinated biphenyls (PCB) (μg/kg)						
Aroclor 1016	Acute median LC50	Earthworm (Eisenia foetida)	251,000	0.01	2,510	Rhett et al. (1989).
Aroclor 1254	Acute median LC50	Earthworm (Eisenia foetida)	251,000	0.01	2,510	Rhett et al. (1989).
Nitroaromatics (μg/kg)						
1,3-Dinitrobenzene	--	--	--	--	2,260	Toxicity value not available. Nitrobenzene used as surrogate.
2,4-Dinitrotoluene	--	--	--	--	--	Toxicity value not available.
2,6-Dinitrotoluene	--	--	--	--	--	Toxicity value not available.
Nitrobenzene	Subchronic (14-day) LC50	Earthworm (species uncertain)	226,000	0.01 ^e	2,260	Neuhauser et al. (1986).
Pentachloronitrobenzene	--	--	--	--		Toxicity value not available.
Phthalate esters (μg/kg)						
Bis(2-ethylhexyl)phthalate	--	--	--	--	--	Toxicity value not available.
Di(n)octyl phthalate	--	--	--	--	--	Toxicity value not available.
Volatile organic compounds (μg/kg)						
Acetone	--	--	--	--	--	Toxicity value not available.
Acrylonitrile	--	--	--	--	--	Toxicity value not available.

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SOIL INVERTEBRATE TOXICITY REFERENCE VALUES

(Page 3 of 12)

Compound	TRV				TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Test Species	Concentration	Uncertainty Factor ^b		
Chloroform	--	--	--	--	--	Toxicity value not available.
Crotonaldehyde	--	--	--	--	--	Toxicity value not available.
1,4-Dioxane	--	--	--	--	--	Toxicity value not available.
Formaldehyde	--	--	--	--	--	Toxicity value not available.
Vinyl chloride	--	--	--	--	--	Toxicity value not available.
Other chlorinated organics ($\mu\text{g/kg}$)						
Hexachlorobenzene	--	--	--	--	--	Toxicity value not available.
Hexachlorobutadiene	--	--	--	--	--	Toxicity value not available.
Hexachlorocyclopentadiene	--	--	--	--	--	Toxicity value not available.
Pentachlorobenzene	LC50 of unspecified duration	Earthworm (species uncertain)	115,000	0.01 ^e	1,150	van Gestel et al. (1991)
Pentachlorophenol	Chronic (21-day) NOAEL for hatching success	Earthworm (<i>Eisenia andrei</i>)	10,000	Not applicable	10,000	van Gestel et al. (1988)
Pesticides ($\mu\text{g/kg}$)						
4,4'-DDE	--	--	--	--	--	Toxicity value not available.
Heptachlor	--	--	--	--	--	Toxicity value not available.
Hexachlorophene	--	--	--	--	--	Toxicity value not available.
Inorganics (mg/kg)						

TABLE E-6

SOIL INVERTEBRATE TOXICITY REFERENCE VALUES

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Compound	TRV				TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Test Species	Concentration	Uncertainty Factor ^b		
Aluminum	--	--	--	--	--	Toxicity value not available.
Antimony	--	--	--	--	--	Toxicity value not available.
Arsenic	Chronic (56-day); reduced cocoon production reported at single concentration tested	Earthworm (<i>Eisenia fetida</i>)	25	0.01 ^e	0.25	Fischer and Koszorus (1992)
Barium	--	--	--	--	--	Toxicity value not available.
Beryllium	--	--	--	--	--	Toxicity value not available.
Cadmium	Chronic (4-month) NOAEL for cocoon production	Earthworm (<i>Dendrobaena rubida</i>)	10	Not applicable	10	Bengtsson and et al. (1986)
Chromium (hexavalent)	Chronic (60-day); survival reduced 25 percent at lowest tested concentration	Earthworm (<i>Octochaetus pattoni</i>)	2	0.1 ^e	0.2	Abbasi and Soni (1983)
Copper	Chronic (56-day) NOAEL for cocoon production	Earthworm (<i>Eisenia fetida</i>)	32.0	Not applicable	32.0	Spurgeon et al. (1994)
Cyanide, total	--	--	--	--	--	Toxicity value not available.
Lead	Chronic (4-month) NOAEL for cocoon production	Earthworm (<i>Dendrobaena rubida</i>)	100	Not applicable	100	Bengtsson et al. 1986

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SOIL INVERTEBRATE TOXICITY REFERENCE VALUES

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Compound	TRV				TRV ^c	Reference and Notes ^d
	Duration and Endpoint ^a	Test Species	Concentration	Uncertainty Factor ^b		
Mercuric chloride	Not available	--	--	--	2.5	Toxicity value not available. TRV for methyl mercury used as a surrogate.
Methyl mercury	Chronic (12-week) NOAEL for segment regeneration and survival	Earthworm (<i>Eisenia foetida</i>)	2.5	Not applicable	2.5	Beyer et al. (1985). Wet weight NOAEL of 1 mg/kg converted to corresponding dry weight NOAEL based on 60 percent moisture content. Uncertainty factor of 0.1 used because segment regeneration may not be a sensitive endpoint.
Nickel	Chronic (20-week) NOAEL for cocoon production	Earthworm (<i>Eisenia foetida</i>)	100	Not applicable	100	Malecki et al. (1982)
Selenium	Chronic; reduced cocoon production at single tested concentration	Earthworm (<i>Eisenia foetida</i>)	77	0.1 ^e	7.7	Fischer and Koszorus (1992)
Silver	--	--	--	--	--	Toxicity value not available.
Thallium	--	--	--	--	--	Toxicity value not available.
Zinc	Chronic (56-day) NOEC for cocoon production	Earthworm (<i>Eisenia foetida</i>)	199	Not applicable	199	Spurgeon et al. (1994)

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SOIL INVERTEBRATE TOXICITY REFERENCE VALUES

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Notes:

- a - duration, the following general guidelines were used: Chronic duration represents exposures occurring about 10 or more days, including exposure during a critical life stage encompassing a sensitive endpoint. Subchronic duration generally lasts 2 days through several days, however a sensitive life stage is not exposed. Acute duration generally includes exposures from 0 to 2 days.
- b Uncertainty factors are used to extrapolate a toxicity value to a chronic NOAEL TRV. See Chapter 5 (Section 5.4) of the SLERAP for a discussion on the use of uncertainty factors.
- c TRV was calculated by multiplying the toxicity value with the uncertainty factor.
- d The references refer to the source of the toxicity value. Complete reference citations are provided below.
- e Best scientific judgment used to identify uncertainty factor. See Chapter 5 (Section 5.4.1.2) for a discussion on the use of best scientific judgement. Factors evaluated include test duration, ecological relevance of measured effect, experimental design, and availability of toxicity data.

HMW	=	High molecular weight
LC50	=	Concentration lethal to 50 percent of the test organisms.
NOAEL	=	No Observed Adverse Effects Level
NOEC	=	No Observed Effects Level
UF	=	Uncertainty Factor
TRV	=	Toxicity Reference Value

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SOIL INVERTEBRATE TOXICITY REFERENCE VALUES

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REFERENCES

Efroymson, Will, and Suter II (1997) provides a comprehensive review of ecologically-relevant soil invertebrate toxicity information. This source was reviewed to identify studies to develop TRVs for invertebrates. Effects of compounds on microbial communities were not considered. Based on the information presented, one or more references were obtained and reviewed to identify compound-specific toxicity values. For some compounds, the available information identified a single study meeting the requirements for a TRV, as discussed in Section 5.4. In most cases, each reference was obtained and reviewed to identify a single toxicity value to develop a TRV for each compound. In a few cases where a primary study could not be obtained, a toxicity value is based on a secondary source. As noted below, additional compendia were reviewed to identify toxicity studies to review. For compounds not discussed in Efroymson, Will, and Suter II (1997), the scientific literature was searched, and relevant studies were obtained and reviewed. The references reviewed are listed below. The study selected for the TRV is highlighted in bold.

Polychlorinated dibenzo(p)dioxins

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Nitrobenzene

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Pentachlorobenzene

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Arsenic

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Cadmium

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SOIL INVERTEBRATE TOXICITY REFERENCE VALUES

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Copper

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Mercuric chloride

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SOIL INVERTEBRATE TOXICITY REFERENCE VALUES

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Nickel

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TABLE E-7

MAMMAL TOXICITY REFERENCE VALUES

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Compound	Basis for Toxicity Reference Value (TRV)				TRV	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Dose ^b	Uncertainty Factor ^c		
Polychlorinateddibenzo-p-dioxins (μg/kg BW-day)						
2,3,7,8-TCDD	Chronic (multigenerational) NOAEL for reproduction	Rat	0.001	Not applicable	0.001	Murray et al. (1979). TRV based on toxicity of 2,3,7,8-TCDD.
Polynuclear aromatic hydrocarbons (PAH) (μg/kg BW-day)						
Total high molecular weight (HMW) PAH	--	--	--	--	100	TRV based on benzo(a)pyrene toxicity. This TRV should be assessing the risk of Total HMW PAH.
Benzo(a)pyrene	Acute (10 days) LOAEL (reproductive effects)	Mouse	10,000	0.01	100	Mackenzie and Angevine (1981)
Benzo(a)anthracene	Single dose LOAEL (gastrointestinal effects)	Mouse	16,666	0.01	167	Bock and King (1959)
Benzo(b)fluoranthene	--	--	--	--	--	Toxicity value not available.
Benzo(k)fluoranthene	--	--	--	--	--	Toxicity value not available.
Chrysene	--	--	--	--	--	Toxicity value not available.
Dibenz(a,h)anthracene	Subchronic (15 days) LOAEL (reduced growth rate)	Rat	200	0.01 ^e	2	Haddow et al. (1937)
Indeno(1,2,3-cd)pyrene	--	--	--	--	--	Toxicity value not available.

TABLE E-7

MAMMAL TOXICITY REFERENCE VALUES

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Compound	Basis for Toxicity Reference Value (TRV)				TRV	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Dose ^b	Uncertainty Factor ^c		
Polychlorinated biphenyls (PCB) (μg/kg BW-day)						
Aroclor 1016	Subchronic (14.5 weeks) LOAEL (mortality)	Mink	20.6	0.01	0.206	Aulerich et al. (1985). TRV based on toxicity of 3,4,5-hexachlorobiphenyl.
Aroclor 1254	Subchronic (14.5 weeks) LOAEL (mortality)	Mink	20.6	0.01	0.206	Aulerich et al. (1985). TRV based on toxicity of 3,4,5-hexachlorobiphenyl.
Nitroaromatics (μg/kg BW-day)						
1,3-Dinitrobenzene	Chronic (16 weeks) NOAEL	Rat	1,051	1.0	1,051	Cody et al. (1981)
2,4-Dinitrotoluene	Chronic (24 months) NOAEL	Dog	700	1.0	700	Ellis et al. (1979)
2,6-Dinitrotoluene	Single dose LOAEL (mortality)	Dog	4,000	0.01	400	Lee et al. (1976)
Nitrobenzene	--	--	--	--	--	Toxicity value not available.
Pentachloronitrobenzene	Chronic (2 years) NOAEL	Mouse	458,333	1.0	458,333	National Toxicology Program (1987)
Phthalate esters (μg/kg BW-day)						
Bis(2-ethylhexyl)phthalate	Chronic (2 years) NOAEL	Rat	60,000	1.0	60,000	Carpenter et al. (1953)
Di(n)octyl phthalate	Chronic (105 days) NOAEL	Mouse	7,500,000	1.0	7,500,000	Heindel et al. (1989)
Volatile organic compounds (μg/kg BW-day)						
Acetone	Subchronic (90 days) NOAEL	Albino Rat, male	100,000	0.1	10,000	U.S. EPA (1986)
Acrylonitrile	Chronic (2 years) LOAEL (lesions and other organ effects)	Rat	4,600	0.1	460	Quast et al. (1980)
Chloroform	Chronic (80 weeks) NOAEL	Mouse	60,000	1.0	60,000	Roe et al. (1979)

TABLE E-7
MAMMAL TOXICITY REFERENCE VALUES

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Compound	Basis for Toxicity Reference Value (TRV)				TRV	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Dose ^b	Uncertainty Factor ^c		
Crotonaldehyde	Acute (4-hour) LD50	Rat	8,000	0.01	80	Rinehart (1967)
1,4-Dioxane	Chronic (23 months) LOAEL (lung tumors)	Guinea Pig	1,069,767	0.1	106,777	Hoch-Ligeti and Argus (1970)
Formaldehyde	Acute (single dose) LOAEL (mortality)	Rat	230,000	0.01	2,300	Tsuchiya et al. (1975)
Vinyl chloride	Chronic (2 years) NOAEL	Rat	1,700	0.1	170	Feron et al. (1981)
Other chlorinated organics ($\mu\text{g/kg BW-day}$)						
Hexachlorobenzene	Chronic (>247 days) NOAEL	Rat	1,600	1.0	1,600	Grant et al. (1977)
Hexachlorobutadiene	Chronic (2 years) NOAEL	Rat	200	1.0	200	Kociba et al. (1977)
Hexachlorocyclopentadiene	Subchronic (13 weeks) NOAEL	Rat	38,000	0.1	3,800	Abdo et al. (1984)
Pentachlorobenzene	Chronic (180 days) NOAEL	Rat	7,250	1.0	7,250	Linder et al. (1980)
Pentachlorophenol	Subchronic (62 days) NOAEL	Rat	3,000	0.1	300	Schwetz et al. (1978)
Pesticides ($\mu\text{g/kg BW-day}$)						
4,4'-DDE	Subchronic (5 weeks) NOAEL	Rat	10,000	0.1	1,000	Kornburst et al. (1986)
Heptachlor	Subchronic (60 days) LOAEL (mortality)	Rat	250	0.01	2.5	Green (1970)
Hexachlorophene	Acute LD50	Rat	560,000	0.01	5600	Meister (1994)
Inorganics (mg/kg BW-day)						
Aluminum	Chronic (>1 year) LOAEL (growth)	Rat	19.3	0.1	1.93	Ondreicka et al. (1966)

TABLE E-7
MAMMAL TOXICITY REFERENCE VALUES

(Page 4 of 15)

Compound	Basis for Toxicity Reference Value (TRV)				TRV	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Dose ^b	Uncertainty Factor ^c		
Antimony	Chronic (4 years) LOAEL (mortality)	Rat	0.66	0.1	0.066	Schroeder et al. (1970)
Arsenic	Chronic (2 years) NOAEL	Dog	1.25	1.0	1.25	Byron et al. (1967)
Barium	Chronic (16 months) NOAEL	Rat	0.51	1.0	0.51	Perry et al. (1983)
Beryllium	Chronic (>1 year) NOAEL	Rat	0.66	1.0	0.66	Schroeder and Mitchner (1975)
Cadmium	Chronic (>150 days) LOAEL (reproduction)	Mouse	2.52	0.01	0.0252	Schroeder and Mitchner (1971)
Chromium (hexavalent)	Chronic (1 year) NOAEL	Rat	3.5	1.0	3.5	MacKenzie et al. (1958)
Copper	Chronic (357 days) NOAEL	Mink	12.0	1.0	12.0	Aulerich et al. (1982)
Total Cyanide	Chronic (2 years) NOAEL	Rat	24	1.0	24	Howard and Hanzal (1955)
Lead	Chronic (>150 days) LOAEL (mortality)	Mouse	3.75	0.01	0.0375	Schroeder and Mitchner (1971)
Mercuric chloride	Chronic (6 months) NOAEL (reproduction)	Mink	1.01	1.0	1.01	Aulerich et al. (1974)
Methyl mercury	Subchronic (93 days) NOAEL	Rat	0.032	1.0	0.032	Verschuuren et al. (1976)
Nickel	Chronic (2 years) NOAEL	Rat	50	1.0	50	Ambrose et al. (1976)
Selenium	Chronic (>150 days) LOAEL (mortality)	Mouse	0.76	0.1	0.076	Schroeder and Mitchner (1971)
Silver	Chronic (125 days) LOAEL (hypoactivity)	Mouse	3.75	0.1	0.375	Rungby and Danscher (1984)

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MAMMAL TOXICITY REFERENCE VALUES

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Compound	Basis for Toxicity Reference Value (TRV)				TRV	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Dose ^b	Uncertainty Factor ^c		
Thallium	Subchronic (60 days) LOAEL (testicular function)	Rat	1.31	0.01 ^h	0.0131	Formigli et al. (1986)
Zinc	Subchronic (13 weeks) NOAEL	Mouse	104	0.1	10.4	Maita et al. (1981)

Notes:

- a The duration of exposure is defined as chronic if it represents about 10 percent or more of the test animal's lifetime expectancy. Acute exposures represent single exposure or multiple exposures occurring within about two weeks or less. Subchronic exposures are defined as multiple exposures occurring for less than 10 percent of the test animal's lifetime expectancy but more than 2 weeks.
- b Reported values, which were dose in food or diet, were converted to dose based on body weight and intake rate using Opresko, Sample, and Suter 1996.
- c Uncertainty factors are used to extrapolate a toxicity value to a chronic NOAEL TRV. See Chapter 5 (Section 5.4) for a discussion on the use of uncertainty factors. The TRV was calculated by multiplying the toxicity value by the uncertainty factor.
- d The references refer to the study or studies from which the endpoint and doses were identified. Complete reference citations are provided at the end of this table.
- e Best scientific judgement used to identify uncertainty factor. See Chapter 5 (Section 5.4.1.2) for a discussion of the use of best scientific judgement. Factors evaluated include test duration, ecological relevance of endpoint, experimental design, and availability of toxicity data.

HMW = High molecular weight
 LD50 = Lethal dose to 50 percent of the test organisms.
 LOAEL = Lowest Observed Adverse Effect Level
 NOAEL = No Observed Adverse Effect Level
 TRV = Toxicity Reference Value

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MAMMAL TOXICITY REFERENCE VALUES

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REFERENCES

Sample, Opresko, and Suter II (1996) provides a comprehensive review of ecologically-relevant mammal toxicity information. This source was reviewed to identify studies to develop TRVs for mammals. Based on the information presented, one or more references were obtained and reviewed to identify compound-specific toxicity values. For some compounds, the available information identified a single study meeting the requirements for a TRV, as discussed in Section 5.4. In most cases, each reference was obtained and reviewed to identify a single toxicity value to develop a TRV for each compound. In a few cases where a primary study could not be obtained, a toxicity value is based on a secondary source. As noted below, additional compendia were reviewed to identify toxicity studies to review. For compounds not discussed in Sample, Opresko, and Suter II (1996), the scientific literature was searched, and relevant studies were obtained and reviewed. The references reviewed are listed below. The study selected for the TRV is highlighted in bold.

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Dibenz(a,h)anthracene

Haddow, A., C.M. Scott, and J.D. Scott. 1937. "The Influence of Certain Carcinogenic and Other Hydrocarbons on Body Growth in the Rat." *Proceeding R. Soc. London. Series B*. Volume 122. Pages 477-507. As cited in IARC Monographs, 1983.

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2,6-Dinitrotoluene

- Lee, C.C., H.V. Ellis III, J.J. Kowalski, J.R. Hodgson, R.D. Short, J.C. Bhandari, T.W. Reddig, and J.L. Minor. 1976. "Mammalian Toxicity of Munitions Compounds. Phase II: Effects of Multiple Doses. Part III: 2,6-Dinitrotoluene. Progress Report No. 4." Midwest Research Institute. Project No. 3900-B. Contract No. DAMD-17-74-C-4073. As cited in ATSDR Toxicological Profile for 2,4- Dinitrotoluene and 2,6-Dinitrotoluene. December 1989.

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Compound	Basis for TRV				TRV	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Dose ^b	Uncertainty Factor ^c		
Polychlorinateddibenzo(p)dioxins (μg/kg BW-day)						
2,3,7,8-TCDD	Subchronic (10 weeks) NOAEL	Ring-necked pheasant hen	0.01	Not applicable	0.01	Nosek et al. (1992). TRV based on toxicity of 2,3,7,8-TCDD.
Polynuclear aromatic hydrocarbons (PAH) (μg/kg BW-day)						
Total high molecular weight (HMW) PAH	--	--	--	--	0.14	TRV based on toxicity of benzo(k)fluoranthene. If TRVs are not available for all individual HMW PAHs, this TRV should be used to assess potential risk of Total HMW PAH.
Benzo(a)pyrene	Acute NOAEL	Chicken embryo	100	0.01	1.0	Brunström et al. (1991).
Benzo(a)anthracene	Acute LD50	Chicken embryo	79	0.01	0.79	Brunström et al. (1991).
Benzo(b)fluoranthene	--	--	--	--	0.14	No toxicity data available for benzo(b) fluoranthene. Benzo(k)fluoranthene used as surrogate.
Benzo(k)fluoranthene	Acute LD50	Chicken embryo	14	0.01	0.14	Brunström et al. (1991).
Chrysene	Acute LOAEL	Chicken embryo	100	0.01	1.0	Brunström et al. (1991).
Dibenz(a,h)anthracene	Acute LD50	Chicken embryo	39	0.01	0.39	Brunström et al. (1991).

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Compound	Basis for TRV				TRV	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Dose ^b	Uncertainty Factor ^c		
Indeno(1,2,3-cd)pyrene	Acute LOAEL	Chicken embryo	100	0.01	1.0	Brunström et al. (1991).
Polychlorinated biphenyls (PCB) ($\mu\text{g/kg BW-day}$)						
Aroclor 1016	--	--	--	--	--	No toxicity data available. Aroclor 1254 TRV used as surrogate.
Aroclor 1254	Chronic (3 months) LOAEL (embryonic mortality)	Ring dove	720	0.1	72	Peakall et al. (1972). TRV based on toxicity of Aroclor 1254.
Nitroaromatics ($\mu\text{g/kg BW-day}$)						
1,3-Dinitrobenzene	Acute LD50	Redwing blackbird	42.2	0.01	0.422	Schafer (1972)
2,4-Dinitrotoluene	--	--	--	--	--	Toxicity value not available.
2,6-Dinitrotoluene	--	--	--	--	--	Toxicity value not available.
Nitrobenzene	--	--	--	--	--	Toxicity value not available.
Pentachloronitrobenzene	Chronic (35 weeks) NOAEL	Chicken	68,750	Not applicable	68,750	Dunn et al. (1979)
Phthalate esters ($\mu\text{g/kg BW-day}$)						
Bis(2-ethylhexyl)phthalate	Subchronic (4 weeks) NOAEL	Ring dove	1,110	0.1	111	Peakall (1974)
Di(n)octyl phthalate	--	--	--	--	--	Toxicity value not available.
Volatile organic compounds ($\mu\text{g/kg BW-day}$)						

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Compound	Basis for TRV				TRV	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Dose ^b	Uncertainty Factor ^c		
Acetone	Acute (5 days) NOAEL	Coturnix quail	5,200,000	0.01 ^h	52,000	Hill and Camardese (1986)
Acrylonitrile	--	--	--	--	--	Toxicity value not available.
Chloroform	--	--	--	--	--	Toxicity value not available.
Crotonaldehyde	--	--	--	--	--	Toxicity value not available.
1,4-Dioxane	--	--	--	--	--	Toxicity value not available.
Formaldehyde	--	--	--	--	--	Toxicity value not available.
Vinyl chloride	--	--	--	--	--	Toxicity value not available.
Other chlorinated organics ($\mu\text{g/kg BW-day}$)						
Hexachlorobenzene	Acute (5 days) NOAEL	Coturnix quail	22,500	0.01	225	Hill and Camardese (1986)
Hexachlorobutadiene	Chronic (3 months) NOAEL	Japanese quail	3185	Not applicable	3185	Schwartz et al. (1974)
Hexachlorocyclopentadiene	--	--	--	--	--	Toxicity value not available.
Pentachlorobenzene	--	--	--	--	--	Toxicity value not available.
Pentachlorophenol	Acute (5 days) NOAEL	Quail	403,000	0.01	4,030	Hill and Camardese (1986)
Pesticides ($\mu\text{g/kg BW-day}$)						
4,4'-DDE	Acute (5 days) LOAEL (mortality)	Coturnix quail	84,500	0.01	845	Hill and Camardese (1986). Test data for 1,1'-DDE used as a surrogate for 4,4'-DDE.

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BIRD TOXICITY REFERENCE VALUES

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Compound	Basis for TRV				TRV	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Dose ^b	Uncertainty Factor ^c		
Heptachlor	Acute (5 days) LOAEL (mortality)	Quail	6,500	0.01	65	Hill and Camardese (1986)
Hexachlorophene	Acute LD50	Bobwhite quail	575,000	0.01	5,750	Meister (1994)
Inorganics (mg/kg BW-day)						
Aluminum	Chronic (4 -months) NOAEL (reproduction)	Ringed Turtle Dove	110	1.0	100	Carriere et al. (1986)
Antimony	--	--	--	--	--	Toxicity value not available. Ridgeway and Karnofsky (1952) reported LD50 for doses to eggs; however, that value could not be converted to a dose based on post-hatching environmental exposure.
Arsenic	Chronic (7 months) NOAEL	Brown-headed cowbird	2.46	1.0	2.46	U.S. Fish and Wildlife Service (1969)
Barium	Subchronic (4 weeks) NOAEL	One day old chick	208.26	0.1	20.8	Johnson et al. (1960)
Beryllium	--	--	--	--	--	Toxicity value not available.
Cadmium	Chronic (90 days) NOAEL	Mallard drake	1.45	Not applicable	1.45	White and Finley (1978)
Chromium (hexavalent)	Chronic (5 months) NOAEL	Black duck	1.0	Not applicable	1.0	Haseltine et al. (1985). TRV based on trivalent chromium.
Copper	Chronic (10 weeks) NOAEL (growth)	1-day old chicks	46.97	1.0	46.97	Mehring et al. (1960)

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BIRD TOXICITY REFERENCE VALUES

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Compound	Basis for TRV				TRV	Reference and Notes ^d
	Duration and Endpoint ^a	Test Organism	Dose ^b	Uncertainty Factor ^c		
Total Cyanide	Acute LD50	American kestrel	4	0.01	0.04	Wiemeyer et al. (1986). Sodium cyanide is used as a surrogate for total cyanides.
Lead	Acute (7 days) LOAEL (altered enzyme levels)	Ringed turtle dove	25	0.001	0.025	Kendall and Scanlon (1982)
Mercuric chloride	Acute (5 days) LOAEL (mortality)	Coturnix quail	325	0.01	3.25	Hill and Camardese (1986)
Methyl mercury	Chronic (3 generations) LOAEL (mortality)	Mallard	0.064	0.1	0.0064	Heinz (1979)
Nickel	Subchronic (5 days) NOAEL	Coturnix quail	650	0.1	65	Hill and Camardese (1986)
Selenium	Chronic (78 days) NOAEL	Mallard	0.5	1.0	0.5	Heinz et al. (1987)
Silver	Subchronic (14 days) NOAEL	Mallard	1,780	0.1	178	U.S. EPA (1997)
Thallium	Acute LD50	Starling	35	0.01	0.35	Schafer (1972)
Zinc	Chronic (44 weeks) NOAEL	Leghorn hen and New Hampshire rooster	130.9	1.0	130.9	Stahl et al. (1990)

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Notes:

- a The duration of exposure is defined as chronic if it represents about 10 percent or more of the test animal's lifetime expectancy. Acute exposures represent single exposure or multiple exposures occurring within about two weeks or less. Subchronic exposures are defined as multiple exposures occurring for less than 10 percent of the test animal's lifetime expectancy but more than 2 weeks.
- b Reported value which were dose in diet or water were converted to dose based on body weight and intake rate using Opresko, Sample, and Suter (1996).
- c Uncertainty factors are used to extrapolate a reported toxicity value to a chronic NOAEL TRV. See Chapter 5 (Section 5.4) of the SLERAP for a discussion on the use of uncertainty factors. The TRV was calculated by multiplying the toxicity value by the uncertainty factor. A "not applicable" uncertainty factor is equivalent to a value equal to 1.0.
- d The references refer to the study from which the endpoint and doses were identified. Complete reference citations are provided below.
- e Best scientific judgement used to identify uncertainty factor. See Chapter 5 (Section 5.4.1.2) for a discussion on the use of best scientific judgement. Factors evaluated include test duration, ecological relevance of endpoint, experimental design, and availability of toxicity data.

HMW	=	High molecular weight
LOAEL	=	Lowest Observed Adverse Effect Level
LD50	=	Concentration lethal to 50 percent of the test organisms.
NOAEL	=	No Observed Adverse Effect Level
TRV	=	Toxicity Reference Value

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REFERENCES

Sample, Opresko, and Suter II (1996) provides a comprehensive review of bird toxicity information. This source was reviewed to identify studies to develop TRVs for birds. Based on the information presented, one or more references were obtained and reviewed to identify compound-specific toxicity values. For some compounds, the available information identified a single study meeting the requirements for a TRV, as discussed in Chapter 5 (Section 5.4) of the SLERAP. In most cases, each reference was obtained and reviewed to identify a single toxicity value to develop a TRV for each compound. As noted below, additional compendia were reviewed to identify toxicity studies to review. In a few cases where a primary study could not be obtained, a toxicity value is based on a secondary source. For compounds not discussed in Sample, Opresko, and Suter II (1996), the scientific literature was searched, and relevant studies were obtained and reviewed. The references reviewed are listed below. The study selected for the TRV is highlighted in bold.

Polychlorinated dibenzo(p)dioxins

Nosek, J.A., S.R. Craven, J.R. Sullivan, S.S. Hurley, and R.E. Peterson. 1992. "Toxicity and Reproductive Effects of 2,3,7,8-Tetrachlorodibenzo-p-dioxin in Ring-Necked Pheasant Hens." *Journal of Toxicology and Environmental Health*. Volume 35. Pages 187-198.

U.S. EPA. 1993. *Interim Report on Data and Methods for Assessment of 2,3,7,8-Tetrachlorodibenzo-p-dioxin Risks to Aquatic Life and Associated Wildlife*. EPA/600/R-93/055. Office of Research and Development. Washington, D.C. March. This report identified the two studies listed below.

Greig, J.B., G. Jones, W.H. Butler, and J.M. Barnes. 1973. "Toxic Effects of 2,3,7,8-Tetrachlorodibenzo-p-dioxins. *Food and Cosmetics Toxicology*. Volume 11. Pages 585-595.

Hudson, R., R. Tucker, and M. Haegle. 1984. *Handbook of Toxicity of Pesticides to Wildlife*. Second Ed. U.S. Fish and Wildlife, Resources Publication No. 153. Washington, D.C.

Benzo(a)pyrene

Brunström, B., D. Broman, and C. Näf. 1991. "Toxicity and EROD-Inducing Potency of 24 Polycyclic Aromatic Hydrocarbons (PAHs) in Chick Embryos." *Archives of Toxicology*. Volume 65. Pages 485-489.

Benzo(a)anthracene

Brunström, B., D. Broman, and C. Näf. 1991. "Toxicity and EROD-Inducing Potency of 24 Polycyclic Aromatic Hydrocarbons (PAHs) in Chick Embryos." *Archives of Toxicology*. Volume 65. Pages 485-489.

Benzo(k)fluoranthene

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Brunström, B., D. Broman, and C. Näf. 1991. "Toxicity and EROD-Inducing Potency of 24 Polycyclic Aromatic Hydrocarbons (PAHs) in Chick Embryos." *Archives of Toxicology*. Volume 65. Pages 485-489.

Chrysene

Brunström, B., D. Broman, and C. Näf. 1991. "Toxicity and EROD-Inducing Potency of 24 Polycyclic Aromatic Hydrocarbons (PAHs) in Chick Embryos." *Archives of Toxicology*. Volume 65. Pages 485-489.

Dibenz(a,h)anthracene

Brunström, B., D. Broman, and C. Näf. 1991. "Toxicity and EROD-Inducing Potency of 24 Polycyclic Aromatic Hydrocarbons (PAHs) in Chick Embryos." *Archives of Toxicology*. Volume 65. Pages 485-489.

Indeno(1,2,3-cd)pyrene

Brunström, B., D. Broman, and C. Näf. 1991. "Toxicity and EROD-Inducing Potency of 24 Polycyclic Aromatic Hydrocarbons (PAHs) in Chick Embryos." *Archives of Toxicology*. Volume 65. Pages 485-489.

Polychlorinated Biphenyls

Peakall, D.B., J.L. Lincer, S.E. Bloom. 1972. "Embryonic Mortality and Chromosomal Alterations Caused by Aroclor 1254 in Ring Doves." *Environmental Health Perspectives*. Volume 1. Pages 103-104.

Dahlgren, R.B., R.L. Linder, and C.W. Carlson. 1972. "Polychlorinated Biphenyls: Their Effects on Pinned Pheasants." *Environmental Health Perspectives*. Volume 1. Pages 89-101.

McLane, M.A.R., and D.L. Hughes. 1980. "Reproductive Success of Screech Owls Fed Aroclor 1248." *Archives of Environmental Contamination and Toxicology*. Volume 9. Pages 661-665.

1,3-Dinitrobenzene

Schafer, E.W. 1972. "The Acute Oral Toxicity of 369 Pesticidal, Pharmaceutical and Other Chemicals to Wild Birds." *Toxicological and Applied Pharmacology*. Volume 21. Pages 315-330.

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Pentachloronitrobenzene

Dunn, J. S., P. B. Bush, N. H. Booth, R.L. Farrell, D. M. Thomason, and D. D. Goetsch. 1979. Effect of Pentachloronitrobenzene upon Egg Production, Hatchability, and Residue Accumulation in the Tissues of White Leghorn Hens. *Toxicology and Applied Pharmacology*. Volume 48. Pages 425-433.

Bis(2-ethylhexyl)phthalate

Peakall, D.B. 1974. "Effects of Di-n-butyl and Di-2-ethylhexyl Phthalate on the Eggs of Ring Doves. *Bulletin of Environmental Contamination and Toxicology*." Volume 12. Pages 698-702.

Acetone

Hill, E.F., and M.B. Camardese. 1986. "Lethal Dietary Toxicities of Environmental Contaminants and Pesticides to Coturnix." Fish and Wildlife Service. Technical Report 2.

1,4-Dioxane

Giavini, E., C. Vismara, and L. Broccia. 1985. "Teratogenesis Study of Dioxane in Rats." *Toxicology Letters*. Volume 26. Pages 85-88. This study did not evaluate an ecologically relevant endpoint. Therefore, the data were not used to develop a TRV.

Hexachlorobenzene

Hill, E.F., and M.B. Camardese. 1986. "Lethal Dietary Toxicities of Environmental Contaminants and Pesticides to Coturnix." Fish and Wildlife Service. Technical Report 2.

Hexachlorobutadiene

Schwetz, B.A., J.M. Norris, R.J. Kociba, P.A. Keeler, R.F. Cornier, and P.J. Gehring. 1974. "Reproduction Study in Japanese Quail Fed Hexachlorobutadiene for 90 Days." *Toxicology and Applied Pharmacology*. Volume 30. Pages 255-265.

Pentachlorophenol

Hill, E.F., and M.B. Camardese. 1986. "Lethal Dietary Toxicities of Environmental Contaminants and Pesticides to Coturnix." Fish and Wildlife Service. Technical Report 2.

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4,4'-DDE

Hill, E.F., and M.B. Camardese. 1986. "Lethal Dietary Toxicities of Environmental Contaminants and Pesticides to Coturnix." Fish and Wildlife Service. Technical Report 2.

Mendenhall, V.M., E.E. Klaas, and M.A.R. McLane. 1983. "Breeding Success of Barn Owls (*Tyto alba*) Fed Low Levels of DDE and Dieldrin." *Archives of Environmental Contamination and Toxicology*. Volume 12. Pages 235-240.

Shellenberger, T.E. 1978. "A Multi-Generation Toxicity Evaluation of P-P'-DDT and Dieldrin with Japanese Quail. I. Effects on Growth and Reproduction." *Drug Chemistry and Toxicology*. Volume 1. Pages 137-146

Heptachlor

Hill, E.F., and M.B. Camardese. 1986. "Lethal Dietary Toxicities of Environmental Contaminants and Pesticides to Coturnix." Fish and Wildlife Service. Technical Report 2.

Hexachlorophene

Meister, R.J. (ed.) 1994. *Farm Chemicals Handbook '94*. Meister Publishing Company, Willoughby, Ohio. Volume 80. Page C189.

Aluminum

Carriere, D., K.L. Fischer, D.B. Peakall, and P. Anghern. 1986. "Effects of Dietary Aluminum Sulphate on Reproductive Success and Growth of Ringed Turtle Doves (*Streptopelia risoria*)."

Canadian Journal of Zoology. Volume 64. Pages 1500-1505.

Carriere, D., K. Fischer, D. Peakall, and P. Angehrn. 1986. "Effects of Dietary Aluminum in Combination with Reduced Calcium and Phosphorus on the Ring Dove (*Streptopelia risoria*)."

Water, Air, and Soil Pollution. Volume 30. Pages 757-764.

Antimony

Ridgeway, L.P. and D.A. Karnofsky. 1952. "The Effects of Metals on the Chick Embryo: Toxicity and Production of Abnormalities in Development." *Annals of New York Academy of Sciences*. Volume 55. Pages 203-215.

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Arsenic

U.S. Fish and Wildlife Service. 1969. "Publication 74." Bureau of Sport Fisheries and Wildlife. As cited in Sample, Opresko, and Suter II (1996).

Barium

Johnson, D., Jr., A.L. Mehring, Jr., and H.W. Titus. 1960. "Tolerance of Chickens for Barium." *Proceedings of the Society for Experimental Biology and Medicine*. Volume 104. Pages 436-438.

Cadmium

White, D.H., and M.T. Finley. 1978. "Uptake and Retention of Dietary Cadmium in Mallard Ducks." *Environmental Research*. Volume 17. Pages 53-59.

Chromium

Haseltine, S.D., and others. 1985. "Effects of Chromium on Reproduction and Growth of Black Ducks." As cited in U.S. Fish and Wildlife Service. 1986. *Chromium Hazards to Fish, Wildlife, and Invertebrates: A Synoptic Review*. January. Page 38.

Copper

Mehring, A.L.Jr., J.H. Brumbaugh, A.J. Sutherland, and H.W. Titus. 1960. "The Tolerance of Growing Chickens for Dietary Copper." *Poultry Science*. Volume 39. Pages 713-719.

Cyanide

Wiemeyer, S.N., E.F. Hill, J.W. Carpenter, and A.J. Krynitsky. 1986. "Acute Oral Toxicity of Sodium Cyanide in Birds." *Journal of Wildlife Diseases*. Volume 22. Pages 538-46.

Lead

Kendall, R.J., and P.F. Scanlon. 1982. "The Toxicology of Ingested Lead Acetate in Ringed Turtle Doves *Streptopelia risoria*." *Environmental Pollution*. Volume 27. Pages 255-262.

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Pattee, O.H. 1984. "Eggshell Thickness and Reproduction in American Kestrels Exposed to Chronic Dietary Lead." *Archives of Environmental Contamination and Toxicology*. Volume 13. Pages 29-34.

Mercuric chloride

Hill, E.F., and M.B. Camardese. 1986. "Lethal Dietary Toxicities of Environmental Contaminants and Pesticides to Coturnix." Fish and Wildlife Service. Technical Report 2.

Hill, E. F. and C. S. Schaffner. 1976. "Sexual Maturation and Productivity of Japanese Quail Fed Graded Concentrations of Mercuric Chloride." *Poultry Science*. Volume 55. Pages 1449-1459.

Methyl mercury

Heinz, G.H. 1979. "Methylmercury: Reproductive and Behavioral Effects on Three Generations of Mallard Ducks." *Journal of Wildlife Management*. Volume 43. Pages 394-401.

Spann, J.W., G.H. Heinz, M.B. Camardese, E.F. Hill, J.F. Moore, and H.C. Murray. 1986. "Differences in Mortality Among Bobwhite Fed Methylmercury Chloride Dissolved in Various Carriers." *Environmental Toxicology and Chemistry*. Volume 5. Pages 721-724.

Nickel

Hill, E.F., and M.B. Camardese. 1986. "Lethal Dietary Toxicities of Environmental Contaminants and Pesticides to Coturnix." Fish and Wildlife Service. Technical Report 2.

Cain, B.W., and E.A. Pafford. 1981. "Effects of Dietary Nickel on Survival and Growth of Mallard Ducklings." *Archives of Environmental Contamination and Toxicology*. Volume 10. Pages 737-745.

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Heinz, G., and others. 1987. "Research at Patuxent Wildlife Research Center." As cited in Sample, Opresko, and Suter II (1996).

Heinz, G.H., D.J. Hoffman, A.J. Krynnitsky, and D.M.G. Weller. 1987. "Reproduction in Mallards Fed Selenium." *Environmental Toxicology and Chemistry*. Volume 6. Page 423-433.

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Silver

U.S. EPA. 1997. **Aquatic Toxicity Information Retrieval Database (AQUIRE)**. Office of Research and Development, National Health and Environmental Effects Research Laboratory, Mid-Continent Ecology Division. January.

Thallium

Schafer, E.W. 1972. "The Acute Oral Toxicity of 369 Pesticidal, Pharmaceutical and Other Chemicals to Wild Birds." *Toxicological and Applied Pharmacology*. Volume 21. Pages 315-330.

Zinc

Stahl, J.L., J.L. Greger, and M.E. Cook. 1990. "Breeding-Hen and Progeny Performance When Hens Are Fed Excessive Dietary Zinc." *Poultry Science*. Volume 69. Pages 259-263.