

US EPA ARCHIVE DOCUMENT

ATTACHMENT 2

POLYCHLORINATED BIPHENYLS (PCBs)
GROUNDWATER MODEL ASSESSMENT



POLYCHLORINATED BIPHENYLS (PCBs) GROUNDWATER MODEL ASSESSMENT

Introduction

After reviewing the hydrogeologic setting and proposed design of the Chemical Waste Unit (CWU) and to simplify the groundwater model, it was determined that contaminant transport would conservatively be modeled vertically through the 3-foot recompacted clay liner (1×10^{-7} cm/sec), excluding the other components of the liner system and the underlying geology. The PCBs concentration (set at 500 parts per million (ppm)) in leachate would be applied directly to the top of the 3-foot recompacted clay liner. A one dimensional POLLUTE model assessing the clay liner as a possible migration pathway was created for the proposed CWU.

Groundwater Model Input

The following information documents the assumptions and values used for the model. The model represents the anticipated site conditions. The assumptions and values are based on the actual design and Construction Quality Assurance Plan proposed for the CWU and the information obtained from numerous site investigations. When site specific information was not available, appropriate and conservative values from literature were used. The input parameters for the groundwater model are identified in Table 1.

All model input must have consistent units. Each of the model input parameters are discussed briefly in the following paragraphs. Documentation for model input parameters is included within this Attachment.

Model Length

As discussed earlier, one (1) layer will be modeled at the site, the 3-foot recompacted clay liner (1.0×10^{-7} cm/sec). The model length is the thickness of the clay liner (3 feet or 0.9144 meters). Although the model has been set up assuming an infinite bottom boundary, the model was evaluated at the base of the recompacted clay liner.

Initial Leachate Concentration

The initial leachate concentration input used was 500 ppm. Clinton Landfill, Inc. has agreed to not accept PCB wastes at concentrations greater than 500 ppm. The actual expected concentration of PCBs in the leachate at the proposed CWU will be much lower (likely more than 100,000 times less) than 500 ppm but to be conservative the maximum concentration was used in the model. The model results represent the PCB concentration in ppm.

Number of Layers

As discussed above, one layer will be modeled at the site, the 3-foot recompacted clay liner. POLLUTE also allows a layer to be subdivided so that the predicted concentration distribution within a layer can be evaluated. The recompacted clay liner was divided into 3 sublayers (a sublayer for every foot of the clay liner (1, 2, and 3 feet or 0.3048, 0.6096, and 0.9144 meters).



**TABLE 1
POLLUTE MODEL INPUT PARAMETER VALUES**

Parameter	Value	Notes	Data
Model Length (L)(m)	0.9144	Total Length of Recompacked Clay Liner	1,2
Initial Concentration (ppm or mg/L)	500	Maximum Concentration of PCB Wastes	2,3
Number of Layers	1	Total Number of Modeled Layers	1,2
Modeling Period (years)	1,000		2
TALBOT PARAMETERS			
TAU	7	Talbot Parameters for the Numerical Inversion of the Laplace Transform	2
Sigma	0		2
RNU	2		2
N	20		2
Recompacted Clay Liner			
Sublayers	3	Model Parameter	2
Thickness (b) (m)	0.9144	Design Specification	1,2
Effective Porosity (n)	0.24	Average Effective Porosity from Laboratory Results for the Clinton Landfill No. 2 Recompacked Clay Liner	1,2
Partitioning Coefficient (K_d) (Kg/m ³)	6.413	Calculated from $k_{oc} \times f_{oc}$	1,2
Degradation (λ)	0.0	No Degradation Modeled	2,3
Density (ρ) (Kg/m ³)	1,970	Average Density from Laboratory Results for the Clinton Landfill No. 2 Recompacked Clay Liner	1,2
Vertical Darcy Velocity (m/yr)	0.042	Assuming 1 foot of Leachate Head on the Recompacked Clay Liner	2,3
Horizontal Darcy Velocity (m/yr)	0.0	Assuming Vertical Flow in Liner (only)	1,2
Effective Diffusion Coefficient (D^*) (m ² /yr)	0.017	Based on Average Free-Solution Diffusion Coefficient for PCBs	2,3
Mechanical Dispersion Coefficient (D_m) (m ² /yr)	0.017	$D_m = D^*$ (Dispersion was set equal to diffusion due to the low seepage rate out of the liner, dispersion will be dominated by diffusion)	2,3
Coeff. of Hydrodynamic Dispersion (D) (m ² /yr)	0.017	$D = D^*$ (Due to the low seepage rate, dispersion will be dominated by diffusion)	2,3
Explanation of Data: 1. Value is based on actual anticipated site conditions 2. Value is required model input parameter 3. Value is conservative value which will result in higher predicted concentrations than the actual anticipated site conditions			



Advective (Darcy) Velocity

POLLUTE requires the input of a Darcy velocity, which was calculated across the 3-foot recompacted clay liner. The Darcy velocity (0.042 m/yr) was determined by multiplying the gradient $(1.33)(4 \text{ feet (1 foot of leachate head} + 3 \text{ feet of clay liner)})/3 \text{ feet (thickness to the clay liner)} = 1.33$ across the clay liner by the hydraulic conductivity of the clay liner ($1.0 \times 10^{-7} \text{ cm/sec}$).

Modeling Period

The modeling period was set at 1,000 years.

Talbot Parameters

POLLUTE uses a Laplace transform to find the solution to the advection-dispersion equation. The numerical inversion of the Laplace transform depends on the Talbot parameters. The model provides default values for the Talbot parameters or they can be selected by the user. The default Talbot parameters were used in this groundwater model.

Boundary Conditions

POLLUTE requires the specification of an upper and lower boundary condition. The top boundary condition typically represents the landfill as a potential source. When modeling the landfill as a surface boundary, the concentration of each constituent in leachate can be assumed to be constant or a specific mass can be assumed to be present. Assuming a specific mass results in a decreasing source concentration over time, which would most accurately represent the fact that leachate concentrations in landfills with leachate collection and removal systems will gradually decrease over time. However, a constant concentration was assumed as it results in conservative model results.

The lower boundary condition was specified as an infinite bottom layer. This boundary condition assumes that horizontal flow can continue to any distance, which allows for realistic analysis of conditions at the base of the recompacted clay liner.

Hydrodynamic Dispersion Coefficient

POLLUTE requires the input of a hydrodynamic dispersion coefficient for each layer. Table 1 lists the model input dispersion coefficient value for the recompacted clay liner. Dispersion will be dominated by diffusion due to the low outward Darcy velocity (0.042 m/yr). An input of $0.017 \text{ m}^2/\text{y}$ (Lesage and Jackson, 1992) was used to represent the effective diffusion coefficient in the 3-foot recompacted clay liner. This value is the average free-solution diffusion coefficient for PCBs. Documentation of the average free-solution diffusion coefficient is provided in this Attachment.

Effective Porosity and Dry Density Input

Table 1 lists the effective porosity and dry density values for the recompacted clay liner. The effective porosity value for the recompacted clay liner (0.24) was obtained from laboratory data for the Clinton No.2 Landfill recompacted clay liner, which has been provided in this Attachment. The density value was also obtained from laboratory data for the Clinton No.2 Landfill recompacted clay liner.



Partitioning Coefficient

The partitioning coefficient (K_d) ($K_d = K_{oc} \times f_{oc}$) was used to simulate retardation of constituents in the recompacted clay liner. The partitioning coefficient (K_d) is specific to each particular compound and its respective partition coefficient between organic carbon and water (K_{oc}) and the recompacted clay liner organic carbon content (f_{oc}). The partitioning coefficient (K_d) was calculated to be 6,413 mg/L or 6.413 Kg/m³ ($K_d = 5.3 \times 10^5 \text{ mg/L} \times 1.21\% \text{ or } 0.0121 = 6,413 \text{ mg/L}$). The clay from the Tiskilwa Formation will be used for construction of the recompacted clay liner. The average organic carbon content (f_{oc}) for the Tiskilwa Formation was 1.21% or 0.0121. The partition coefficient of PCBs between organic carbon and water (K_{oc}) was $5.3 \times 10^5 \text{ mg/L}$. The documentation for the organic carbon content (f_{oc}) and partition coefficient of PCBs between organic carbon and water (K_{oc}) is provided in this Attachment.

Degradation

Degradation is used to simulate degradation of constituents in the subsurface. Degradation is specific to each particular compound.

Although degradation can play a significant role in reducing the migration of numerous constituents in groundwater, it is conservatively assumed that degradation is not present.

Model Evaluation Distance

The model evaluation distance is not a model input parameter. However, this distance is needed in order to evaluate the results of the groundwater model since the model only provides results for specified distances. The model was evaluated at the three points (1, 2, and 3 feet or 0.3048, 0.6096, and 0.9144 meters) in the recompacted clay liner.

Model Results

The POLLUTE output for the PCBs groundwater model assessment is included in this Attachment. The model predicted PCB concentrations, for the entire 1,000 year simulation period, at the base of the recompacted clay liner (3 feet or 0.9144 meters) is 0.0 ppm. Predicted concentrations at 1 and 2 feet were 4.48×10^{-15} ppm and 1.10×10^{-33} ppm, respectively.

The PCBs groundwater model assessment indicates that PCBs will not migrate out of the 3 foot recompacted clay liner even after 1,000 years and therefore will not impact the local or regional groundwater resources (including the Mahomet Aquifer).

It should be noted that the model discussed above is extremely conservative (resulting in higher predicted concentrations) and did not include the additional 150 feet of in-situ clay or incorporate the three layers of 60-mil thick HDPE geomembrane or the geocomposite clay liner. Additionally, the initial leachate concentration was conservatively set at 500 ppm, the actual expected concentration of PCBs in the leachate at the proposed CWU will be much lower (likely more than 100,000 times less) than 500 ppm.



References

- Anderson, M.P. and Woessner, W.W. (1992). Applied Groundwater Modeling - Simulation of Flow and Advective Transport, Academic Press, San Diego, CA.
- Freeze, R. A. and J. A. Cherry (1979). Groundwater, Prentice-Hall, Englewood Cliffs, NJ.
- Giroud, J. P. and R. Bonaparte (1989). "Leakage through Liners Constructed with Geomembranes - Part I. Geomembrane Liners" and "Leakage through Liners Constructed with Geomembranes - Part II. Composite Liners", Geotextiles and Geomembranes, Vol. 8, Elsevier Science Publishers Ltd., England, Part I, Pages 27-67 and Part II, Pages 71-111.
- Lesage, S. and R. Jackson (1992). Groundwater Contamination and Analysis at Hazardous Waste Sites, Marcel Dekker, Inc., New York, New York.
- Rowe, R.K. (1987). "Pollutant transport through barriers", Proceedings of ASCE Specialty Conference, Geotechnical Practice for Waste Disposal '87, Ann Arbor, June, pp. 159-181.
- Rowe, R.K. (1988). "Contaminant Migrating Through Groundwater: The Role of Analysis in The Design of Barriers", Canadian Geotechnical Journal, 25(4), pp. 778-798.
- Rowe, R.K. (1991). "Contaminant Impact Assessment and the Contaminating Lifespan of Landfills", Canadian Journal of Civil Engineering, vol. 18, pp. 244-253.
- Rowe, R. K. (1991). "Leachate Detection or Hydraulic Control: Two Design Options", Geotechnical Research Center Report, GEOT-3-91, Geotechnical Research Center, The University of Western Ontario.
- Rowe, R. K. (1995). "The Role of Diffusion and the Modeling of its Impact on Groundwater Quality", Geotechnical Research Center Report, GEOT-7-95, Geotechnical Research Center, The University of Western Ontario.
- Rowe, R.K. (1998). "Geosynthetics and the Minimization of Contaminant Migration through Barrier Systems Beneath Solid Waste", Sixth International Conference on Geosynthetics.
- Rowe, R.K. and Booker, J.R. (1985). "1-D pollutant migration in soils of finite depth", Journal of Geotechnical Engineering, ASCE, Vol. 11, GT4, pp. 13-42.
- Rowe, R.K. and Booker, J.R. (1985). "Two-Dimensional Pollutant Migration in Soils of Finite Depth", Canadian Geotechnical Journal, Volume 22, No. 4, pp. 479-499.
- Rowe, R.K. and Booker, J.R. (1986) "A Finite Layer Technique for Calculating Three-Dimensional Pollutant Migration in Soil", Geotechniques 36, No. 2, pp. 205-214.
- Rowe, R.K. and Booker, J.R. (1987). "An efficient analysis of pollutant migration through soil", Chapter 2 in the book "Numerical Methods for Transient and Coupled Systems", Eds. Lewis, Hinton, Bettess and Schrefler. John Wiley & Sons Ltd., pp. 13-42.



- Rowe, R. K. and J. R. Booker (1988). "A Semi-Analytical Model for Contaminant Migration in a Regular Two or Three Dimensional Fractured Network: Conservative Contaminants", Geotechnical Research Center Report, GEOT-15-88, Geotechnical Research Center, The University of Western Ontario.
- Rowe, R.K. and Booker, J.R. (1989). "Contaminant migration through a liner underlain by fractured till and an aquifer", Geotechnical Research Center Report GEOT-12-89; Faculty of Engineering Science, U.W.O.
- Rowe, R. K., J. R. Booker and M. J. Fraser. MIGRATEv9, GAEA Environmental Engineering, Ltd., Windsor, Ontario, Canada, 1995.
- Rowe, R. K., J. R. Booker and M. J. Fraser. POLLUTEv6, GAEA Environmental Engineering, Ltd., Windsor, Ontario, Canada, 1994.
- Rowe, R. K., R. M. Quigley and J. R. Booker. (1995). Clayey Barrier Systems for Waste Disposal Facilities, E & FN Spon, London, England.
- Shackelford, C.D. (1990). "Laboratory diffusion testing for waste disposal - a review", J. of Contaminant Hydrology, Elsevier Science Publishers, Amsterdam, Netherlands.
- Talbot, A. (1979). "The accurate numerical integration of La place transforms", J. Inst. Math's. Applics., 23, pp. 97-120.
- Todd, D.K. (1980). Groundwater Hydrology. John Wiley & Sons Inc. New York.
- U.S. EPA (1992). "Action Leakage Rates for Leak Detection Systems", EPA 530-R-92-004, U.S. EPA, Washington, D.C.
- Walton (1991). Principles of Groundwater Engineering, Lewis Publishers, Inc., Michigan.
- Xu, M. and Y. Eckstein (1995). "Use of Weighted Least Squares Method in Evaluation of the Relationship Between Dispersivity and Field Scale", Ground Water, v.33, No. 6, pp. 905-908.



PCBs GROUNDWATER MODEL DOCUMENTATION


```

*****
*
*
*   P O L L U T E v 6   S I M U L A T I O N   *
*
*   RUN DATE - 29- 1-** *
*   TIME - 18:11:43 *
*
*   REVISION - 27/04/1998 *
*
*   VERSION 6.3.5 *
*
*   COPYRIGHT(c) R.K. ROWE & J.R. BOOKER 1983-1998 *
*   LICENSED USER: Envirogen *
*
*****

```

Clinton Landfill #3 (Compacted Clay Liner)

POLLUTE VERSION 6.3

THE DARCY VELOCITY (Flux) THROUGH THE LAYERS $V_a = 0.4200E-01$ m/a
(Positive for down or into the layer)

PROPERTIES OF THE MATRIX

Layer No.	No. Of Sublayer	Coefficient Hydrodynamic Dispersion m ² /a	Matrix Porosity	Distribution/ Partitioning Coefficient m ³ /kg	Dry Density kg/m ³	Layer Thickness m
1	3	0.17000E-01	0.24000	0.6413E+01	1970.0000	0.9144E+00

The TOP and BOTTOM BOUNDARY CONDITIONS
are defined by CODES Top = 2 Bottom = 4
See below for details

CODE	TOP	BOTTOM
=	Zero Flux	Zero Flux
2 =	C = Const.	C = Const2.
3 =	Finite Mass	Fixed Outflow Velocity
4 =		Infinite Bottom Layer

Initial Source Concentration C0 = 0.5000E+03 mg/L

There is no Radioactive or Biological Decay being Considered

The Parameters used to Invert the Laplace Transform are

TAU = 0.700E+01 N = 20 SIG = 0.000E+00 RNU = 0.200E+01

CALCULATED CONCENTRATIONS AT SELECTED DEPTHS AND TIMES

TIME yr	DEPTH m	CONCENTRATION mg/L
=====		
0.2500E+03	0.0000E+00	0.5000E+03
	0.3048E+00	0.0000E+00
	0.6096E+00	0.0000E+00
	0.9144E+00	0.0000E+00
0.5000E+03	0.0000E+00	0.5000E+03
	0.3048E+00	0.2403E-22
	0.6096E+00	0.6019E-49
	0.9144E+00	0.0000E+00
0.7500E+03	0.0000E+00	0.5000E+03
	0.3048E+00	0.9697E-17
	0.6096E+00	0.5309E-39
	0.9144E+00	0.0000E+00
0.1000E+04	0.0000E+00	0.5000E+03
	0.3048E+00	0.4481E-14
	0.6096E+00	0.1102E-32
	0.9144E+00	0.0000E+00

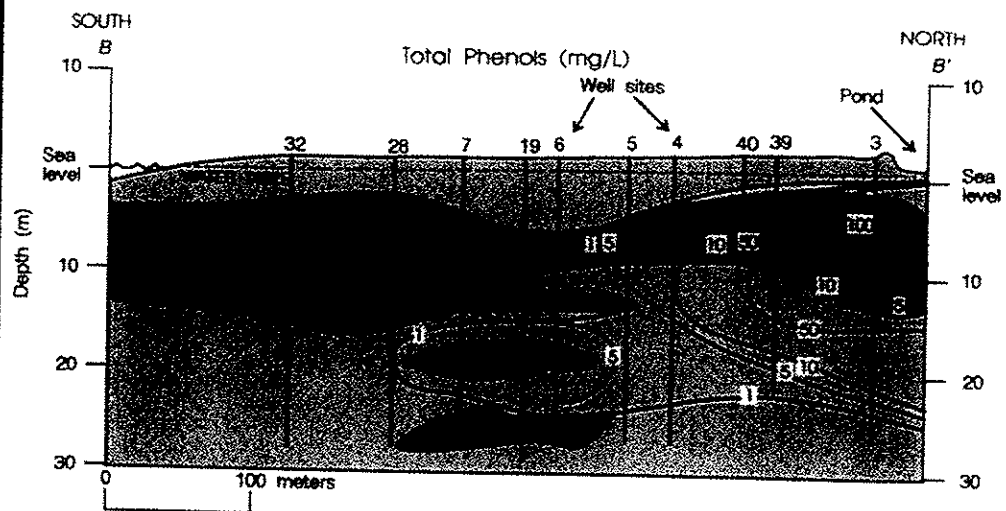
← BASE OF 3 FOOT RECOMPACTED CLAY LINER

N O T I C E

ALTHOUGH THIS PROGRAM HAS BEEN TESTED AND EXPERIENCE WOULD INDICATE THAT IT IS ACCURATE WITHIN THE LIMITS GIVEN BY THE ASSUMPTIONS OF THE THEORY USED, WE MAKE NO WARRANTY AS TO WORKABILITY OF THIS SOFTWARE OR ANY OTHER LICENSED MATERIAL. NO WARRANTIES EITHER EXPRESSED OR IMPLIED (INCLUDING WARRANTIES OF FITNESS) SHALL APPLY. NO RESPONSIBILITY IS ASSUMED FOR ANY ERRORS, MISTAKES OR MISREPRESENTATIONS THAT MAY OCCUR FROM THE USE OF THIS COMPUTER PROGRAM. THE USER ACCEPTS FULL RESPONSIBILITY FOR ASSESSING THE VALIDITY AND APPLICABILITY OF THE RESULTS OBTAINED WITH THIS PROGRAM FOR ANY SPECIFIC CASE.

```
*****
*
*
*
*   P O L L U T E   S I M U L A T I O N   *
*
*   ANALYSIS           COMPLETED          *
*
*   TIME      -      18:11:43             *
*   EXECUTION TIME      0: 0              *
*
*
*****
```


Groundwater Contamination and Analysis at Hazardous Waste Sites



edited by

Suzanne Lesage

Richard E. Jackson

e groundwater has
r chlorine isomers.
or PCB sorbed on
relative concentra-
solution processes

not been reported
ad aquatic systems
degrading PCBs.
present in shallow
onditions, nutrient
it cannot be deter-

However, aerobic
certainly proceed
es, owing to lower
organic substrates.
n the environment
patterns in PCB-
v .. These GC
chlorine isomers
somers compared
o the river. These
obic sediments of
lver Lake had sig-
omers and higher
e original Aroclor
ations in the sedi-
ion reactions can
ters, thereby ulti-

anaerobic biode-
olated from PCB-
ated in a reduced
olor 1242. There
ficant dechlorina-
nd 16 weeks. No
erved, suggesting
these experiments
ns occurred much
ved in laboratory

not been reported
populations and

environmental conditions for degradation can occur in groundwater systems. If anaerobic biodegradation reactions do occur in groundwater, the loss of the higher chlorine isomers and associated formation of the lower chlorine isomers would be similar to that expected due to attenuation of dissolved PCB concentrations by sorption. This may make identification of anaerobic biodegradation in the subsurface difficult.

C. Matrix Diffusion

In fractured rock environments there is a greater potential for extensive migration of dissolved PCBs in groundwater because groundwater velocities are frequently much higher than in porous media, and the potential for sorption of PCBs on the materials composing the fracture surfaces is generally lower than in porous media. Nevertheless, fractured sedimentary rocks such as sandstones, shales, and carbonates frequently do have a substantial matrix porosity (5–20%), and the diffusion of PCBs into the matrix will reduce concentrations in the groundwater flowing through the fractures [19–21]. The attenuation of PCBs by matrix diffusion will be substantially enhanced as a result of sorption on the matrix solids.

The degree of attenuation of PCB migration through fractured media will depend on the diffusivity of the PCB compounds and the degree of sorption. Lower chlorine PCB isomers will have slightly higher diffusivities than higher chlorine isomers because of their lower molecular weights. Aqueous diffusion coefficients for PCB isomers were estimated from the measured diffusion coefficient of benzene and the relative molecular weight of benzene and the PCB isomers [22,23]. The estimated aqueous diffusion coefficient for Cl_1 isomers is approximately 50% greater than that for Cl_8 isomers (see Table 2). The higher the diffusion coefficient, the higher the potential

Table 2 Estimated Free-Solution Diffusion Coefficients for PCB Isomers

PCB isomer	Free-solution diffusion coefficient at 20°C (m^2/sec)
Cl_1	6.4×10^{-10}
Cl_2	6.2×10^{-10}
Cl_3	5.8×10^{-10}
Cl_4	5.5×10^{-10}
Cl_5	5.3×10^{-10}
Cl_6	5.1×10^{-10}
Cl_7	4.8×10^{-10}
Cl_8	4.7×10^{-10}
Chloride	2.0×10^{-9}

$$\text{AVERAGE} = 5.48 \times 10^{-10} \text{ m}^2/\text{SEC}$$

$$= 0.017 \text{ m}^2/\text{YEAR}$$

SOILS DATA FOR INITIAL FILL AREA LINER

CLINTON LANDFILL, INC. #2

BORING/LOCATION		SAMPLE ELEV., ft-msl	MATERIAL	DRY DENSITY		VOID RATIO	POROSITY	VERT. K cm/s
				lb/cu.ft	g/c.c.			
EAST HALF								
East Berm	N5201/E7590	686.8	CL	129.1	2.068	0.3059	0.234	1.01E-08
South Berm	N5059/E7286	683.3	CL	125.6	2.012	0.3420	0.255	4.43E-09
Floor	N5151/E7306	680.0	CL	130.1	2.084	0.3144	0.239	1.22E-08
Floor	N5261/E7456	675.0	CL	127.4	2.041	0.3325	0.250	4.97E-08
Floor	N5300/E7496	680.0	CL	117.9	1.889	0.4300	0.301	1.12E-08
Floor	N5355/E7276	681.0	CL	122.7	1.965	0.3737	0.272	6.36E-09
Sidewall	N5410/E7476	693.0	CL	106.2	1.701	0.5869	0.370	1.39E-09
<u>Points below 3 foot thick earth liner:</u>								
B-1	N5092/E7212	671.0	CL	123.9	1.985	0.3606	0.265	1.82E-08
B-2	N5100/E7284	670.4	CL	124.1	1.988	0.3580	0.264	1.5E-08
B-4	N5100/E7473	668.6	MH	83.8	1.342	0.9960	0.499	1.65E-06
B-5	N5231/E7210	669.9	CL	128.4	2.057	0.3123	0.238	7E-09
B-6	N5230/E7269	670.0	CL	124.8	1.999	0.3503	0.259	9.52E-09
B-7	N5233/E7368	670.9	CL	128	2.050	0.3172	0.241	1.05E-08
B-9	N5229/E7580	691.0	CL	120.4	1.929	0.3997	0.286	5.28E-08
B-10	N5360/E7224	671.4	CL	125.7	2.014	0.3408	0.254	1.08E-08
B-11	N5360/E7282	670.8	CL	126.2	2.022	0.3359	0.251	9.21E-09
B-12	N5355/E7361	671.0	CL	123.2	1.974	0.3685	0.269	7.53E-09
B-13	N5360/E7476	671.0	CL	123.2	1.974	0.3680	0.269	4.41E-07
B-14	N5421/E7376	690.5	CL	116	1.858	0.4536	0.312	1.35E-08
WEST HALF								
th Berm		697 - 699	CL	128.4	2.057	0.333	0.250	1.5E-08
uth Berm		687 - 689	CL	128.9	2.065	0.326	0.246	1.7E-08
Floor	40' W of E berm, 150' S of N berm	676.8 - 678.8	CL	122.8	1.967	0.468	0.319	8.4E-09
Floor	150' S of N berm, 150' E of W berm	680 - 682	CL	122.8	1.967	0.484	0.326	1.1E-08
West Berm	150' S of N berm	693 - 695	CL	135.0	2.163	0.418	0.295	1.1E-08
North Berm	100 E of W berm	697 - 699	CL	122.8	1.967	0.495	0.331	2.7E-08
West Berm	120' S of N berm	680 - 682	CL	111.2	1.781	NM	-	3.7E-08
West Berm	230' S of N berm	682 - 684	CL	118.2	1.893	NM	-	6E-09
<u>Points below 3 foot thick earth liner:</u>								
Boring T-2	N5319/E7009	675.2	CL	121.4	1.945	NM	NM	1.6E-08
Boring T-3	N5321/E7092	675.6	CL	126.4	2.025	NM	NM	3.1E-08
Boring T-4	N5224/E6926	680.5	CL	111.1	1.780	NM	NM	3.4E-08
Boring T-7	N5090/E6927	677	CL	119.7	1.917	NM	NM	3.7E-08

Notes:

1. NM = not measured; Elev. ft-msl = elevation in feet above mean sea level
2. lb/cu.ft = pounds per cubic feet; g/c.c. = grams per cubic centimeter; cm/s = centimeters per second
3. Vert. K = vertical hydraulic conductivity (permeability)
4. Reference: Liner Certification Reports for Initial Fill Area by SKS.

Vertical Permeability of Clay Liner and In-Situ Soil Below Liner - Initial Fill Area Clinton Landfill #2

Type of Liner	Location	Boring	Northing ft	Easting ft	Sample Elevation ft	Dry Density lb/cu.ft	Porosity	Vertical Permeability Kv, cm/s	Kv, cm/s	
									Recompa- cted Clay Liner	In-Situ Σ Below Liner
East Half										
In-situ	Floor	B-1	5092	7212	671.0	123.9	0.265	1.82E-08		1.82E-08
In-situ	Floor	B-2	5100	7284	670.4	124.1	0.264	1.5E-08		1.5E-08
In-situ	Floor	B-4	5100	7473	668.6	83.8	0.499	1.65E-06		
In-situ	Floor	B-5	5231	7210	669.9	128.4	0.238	7.0E-09		7.0E-09
In-situ	Floor	B-6	5230	7269	670.0	124.8	0.259	9.52E-09		9.52E-09
In-situ	Floor	B-7	5233	7368	670.9	128.0	0.241	1.05E-08		1.05E-08
In-situ	Sidewall	B-9	5229	7580	691.0	120.4	0.286	5.28E-08		5.28E-08
In-situ	Floor	B-10	5360	7224	671.4	125.7	0.254	1.08E-08		1.08E-08
In-situ	Floor	B-11	5360	7282	670.8	126.2	0.251	9.21E-09		9.21E-09
In-situ	Floor	B-12	5355	7361	671.0	123.2	0.269	7.53E-09		7.53E-09
In-situ	Floor	B-13	5360	7476	671.0	123.2	0.269	4.41E-07		
In-situ	Sidewall	B-14	5421	7376	690.5	116.0	0.312	1.35E-08		1.35E-08
Recompacted	East Berm		5201	7590	686.8	129.1	0.234	1.01E-08		
Recompacted	South Berm		5059	7286	683.3	125.6	0.255	4.43E-09		
Recompacted	Floor		5151	7306	680.0	130.1	0.239	1.22E-08	1.22E-08	
Recompacted	Floor		5261	7456	675.0	127.4	0.250	4.97E-08	4.97E-08	
Recompacted	Floor		5300	7496	680.0	117.9	0.301	1.12E-08	1.12E-08	
Recompacted	Floor		5355	7276	681.0	122.7	0.272	6.36E-09	6.36E-09	
Recompacted	Sidewall		5410	7476	693.0	106.2	0.370	1.39E-09		
West Half										
Recompacted	North Berm				698.0	128.4	0.250	1.5E-08		
Recompacted	South Berm				688.0	128.9	0.246	1.7E-08		
Recompacted	Floor		40' W of E berm, 150' S of N berm		677.8	122.8	0.319	8.4E-09	8.4E-09	
Recompacted	Floor		150' S of N berm, 150' E of W berm		681.0	122.8	0.326	1.1E-08	1.1E-08	
Recompacted	West Berm		150' S of N berm		694.0	135.0	0.295	1.1E-08		
Recompacted	North Berm		100 E of W berm		698.0	122.8	0.331	2.7E-08		
Recompacted	West Berm	ST1	120' S of N berm		681.0	111.2	NM	3.7E-08		
Recompacted	West Berm	ST2	230' S of N berm		683.0	118.2	NM	6.0E-09		
In-situ	Floor	T-2	5319	7009	675.2	121.4	NM	1.6E-08		1.6E-08
In-situ	Floor	T-3	5321	7092	675.6	126.4	NM	3.1E-08		3.1E-08
In-situ	Floor	T-4	5224	6926	680.5	111.1	NM	3.4E-08		3.4E-08
In-situ	Floor	T-7	5090	6927	677.0	119.7	NM	3.7E-08		3.7E-08
Mean Kv =								8.36E-08	1.65E-08	1.94E-08

Results

Mean Kv of Recompacted Soil Liner (Floor Only) = 1.65E-08 cm/s

Mean Kv of In-Situ Soil Below Recompacted Clay Liner = 1.94E-08 cm/s

Mean Porosity of Recompacted Clay Liner = 0.288

Mean Porosity of In-Situ Soil Below Clay Liner = 0.264

Mean Dry Density of Recompacted Clay Liner = 122.9 lb/cu.ft

Mean Dry Density of In-Situ Soil Below Clay Liner = 122.8 lb/cu.ft

* ESTIMATED EFFECTIVE

POROSITY =

$0.29 - 0.29 \times 0.159$

= 0.24

Notes:

1. B-1 = Boring #B-1; ST1 = Shelby Tube #ST-1; T-1 = Boring #T-1
2. If in-situ soil Kv values are greater than 1.0E-07, the corresponding Kv and dry density values were not used in calculating mean, because recompacted liner was installed at those locations.
3. NM = Not measured
4. cm/s = centimeters per second; lb/cu.ft = pounds per cubic feet
5. Reference: Liner Certification Reports for Initial Fill Area by SKS.

* MAXIMUM EFFECTIVE POROSITY =

$0.33 - 0.33 \times 0.159$

= 0.28

* MINIMUM EFFECTIVE POROSITY

$0.24 - 0.24 \times 0.159$

= 0.20

TABLE 812.314-3
TOTAL ORGANIC CONTENT
FROM SELECTED BORINGS
Clinton Landfill No. 3

Boring ID Number	Geologic Unit	Sample Elev. (ft-MSL)	Date Collected	Total Organic Content (%)
EX-14	Berry Clay	659	12/16/2002	8.8
	Radnor Till	642	12/16/2002	1.4
EX-13	Roxanna Silt	671	12/20/2002	10.0
	Radnor Till	655.5	12/20/2002	1.8
	Berry Clay	661.5	12/20/2002	1.8
	Organic Soil	633	12/23/2002	3.0
EX-12	Berry Clay	660	12/27/2002	2.0
EX-17	Berry Clay	657	1/3/2003	2.0
EX-15	Tiskilwa Formation	673	1/7/2003	0.94
	Berry Clay	662	1/7/2003	1.8
	Radnor Till	657	1/7/2003	0.93
	Radnor Till	646	1/7/2003	0.53
EX-24	Radnor Till	643	1/9/2003	0.54
	Sangamon Soil	636	1/9/2003	3.7
EX-16	Tiskilwa Formation	672.5	1/15/2003	1.00
	Tiskilwa Formation	669.5	1/15/2003	1.1
	Roxanna Silt	668	1/15/2003	7.9
	Organic Soil	634	1/15/2003	7.3
EX-18	Radnor Till	649	1/20/2003	1.2
	Radnor Till	641	1/20/2003	0.56
	Radnor Till	638	1/20/2003	0.49
	Organic Soil	632	1/20/2003	1.9
EX-19	Roxanna Silt	663	1/22/2003	2.5
	Radnor Till	646	1/22/2003	0.57
EX-20	Roxanna Silt	665	1/31/2003	2.20
	Radnor Till	645	2/3/2003	1.1
EX-21	Tiskilwa Formation	690	1/29/2003	1.4
	Tiskilwa Formation	675	1/29/2003	1.6
	Roxanna Silt	670	1/30/2003	5.7
	Radnor Till	652	1/30/2003	1.0
	Radnor Till	643	1/30/2003	0.93
EX-22	Berry Clay	660	1/28/2003	1.4
	Radnor Till	655	1/28/2003	2.5
	Radnor Till	648	1/28/2003	1.3
	Radnor Till	644	1/28/2003	1.4

Notes:

1. Samples collected from 5-foot split-spoon sampler.
2. ft-MSL = feet above Mean Sea Level datum; value was estimated by PDC Technical Services, Inc. field personnel based on ground surface elevation data provided by SKS Engineers, Inc. at the time of sample collection.
3. Total Organic Content = by American Society for Testing and Materials (ASTM) Standard D2974-00 "Standard Test Methods for Moisture, Ash, and Organic Matter of Peat and Other Organic Soils".
4. % = Percent

TISKILWA
FORMATION
AVERAGE
TOTAL ORGANIC
CONTENT
1.21%

APPENDIX A (continued)

Chemical name	CAS No.	Water solubility (mg l ⁻¹)	Reference	Vapor pressure (mm Hg)	Reference	Henry's law constant (atm. m ³ mol ⁻¹)	Reference	K _{oc} (ml g ⁻¹)	Reference	K _{ow}	Reference
PCB's:											
Aroclor 1016 [®]	12674-11-2	4.20 · 10 ⁻¹	H	4.00 · 10 ⁻⁴	I			5.01 · 10 ⁴	DD	2.40 · 10 ⁴	H
Aroclor 1221 [®]	11104-28-2	1.50 · 10	I	6.70 · 10 ⁻³	I	3.24 · 10 ⁻⁴	DD	2.75 · 10 ³	DD	1.23 · 10 ⁴	H
Aroclor 1230 [®]	11141-16-5	1.45	I	4.06 · 10 ⁻³	I	8.64 · 10 ⁻⁴	DD	6.76 · 10 ³	DD	1.54 · 10 ³	I
Aroclor 1240 [®]	53469-21-9	2.40 · 10 ⁻¹	G	4.10 · 10 ⁻⁴	G	5.60 · 10 ⁻⁴	G	5.13 · 10 ³	DD	1.29 · 10 ⁴	I
Aroclor 1248 [®]	12672-25-6	5.40 · 10 ⁻²	G	4.90 · 10 ⁻⁴	G	3.50 · 10 ⁻³	G	4.37 · 10 ⁴	DD	5.62 · 10 ⁵	I
Aroclor 1254 [®]	11097-59-1	1.20 · 10 ⁻²	G	7.70 · 10 ⁻⁵	G	2.70 · 10 ⁻³	G	4.25 · 10 ⁴	E	1.07 · 10 ⁶	I
Aroclor 1260 [®]	11096-82-5	2.70 · 10 ⁻³	G	4.10 · 10 ⁻⁵	G	7.10 · 10 ⁻³	G	2.63 · 10 ⁵	DD	1.38 · 10 ⁷	I
Polychlorinated biphenyls (PCB's)	1336-36-3	3.10 · 10 ⁻²	A	7.70 · 10 ⁻⁵	A	1.07 · 10 ⁻³	A	5.30 · 10 ⁵	A	1.10 · 10 ⁶	A
Ethers:											
Bis(2-chloroethyl)ether	111-44-4	1.02 · 10 ⁴	A	7.10 · 10 ⁻¹	A	1.31 · 10 ⁻⁵	A	1.39 · 10	A	3.16 · 10	A
Bis(2-chloroisopropyl)ether	108-60-1	1.70 · 10 ³	A	8.50 · 10 ⁻¹	A	1.13 · 10 ⁻⁴	A	6.10 · 10	A	1.28 · 10 ²	A
Bis(2-chloroethoxy)methane	111-91-1	8.10 · 10 ⁴	I	< 1.0 · 10 ⁻¹	I	3.78 · 10 ⁻⁷	DD	1.15 · 10 ³	DD	1.82 · 10	DD
2-Chloroethyl vinyl ether	110-75-8	1.50 · 10 ⁴	H	2.87 · 10	H	2.50 · 10 ⁻⁴	Q	6.61	DD	1.90 · 10	I
4-Chlorophenyl phenyl ether	7005-72-3	3.30	H	2.70 · 10 ⁻³	I	2.19 · 10 ⁻⁴	X	3.98 · 10 ³	DD	1.20 · 10 ⁴	H

CAS = Chemical Abstract Service; solubility and vapor pressure values are given for a temperature range of 20-30°C. K_{oc} refers to the organic carbon partition coefficient. K_{ow} refers to the octanol-water partition coefficient.

References: A = U.S.E.P.A. (1986b) measured values; B = U.S.E.P.A. (1986b) calculated values; C = Shifrin (1986) measured values; D = Shifrin (1986) calculated values; E = Kenaga and Goring (1978); F = Alford-Stevens (1986); G = Lyman et al. (1982); H = Callahan et al. (1979) measured values; I = Callahan et al. (1979) estimated values; J = Windholz et al. (1983); K = Dean (1979); L = U.S.C.G. (1978); M = Leo et al. (1971); N = Hunt et al. (1982); O = Mackay et al. (1982); P = Jury et al. (1984); Q = Schwille (1988); R = NIOSH (1987); S = Riddick and Bunger (1970); T = J.L. Wilcox and Conrad (1984); U = A.S.T.M. (1985); V = Chou and Griffin (1983); W = Cole-Parmel Catalog (1989/1990); X = calculated values; Y = Ashland Product Catalog (1986); Z = M.R. Anderson (1988); AA = A.P.I. (1989); BB = NIPER (National Institute for Petroleum and Energy Research), Bartlesville, Oklahoma (pers. commun., 1989); CC = Chevron[®], San Francisco, California (pers. commun., 1988); DD = Montgomery and Welkom (1990).

From Mercer, J.W. and R.M. Cohen, A review of immiscible fluids in the subsurface: Properties, models, characterization and remediation, *J. Cont. Hyd.*, 6, 107-163, 1990.