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PROPOSED
TOTAL MAXIMUM DAILY LOAD (TMDL)

For

Dissolved Oxygen & Nutrients

In

Cockroach Bay
(WBID 1778)

March 2013

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United States Environmental Protection Agency

Region 4 serving the southeast

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SUMMARY SHEET for WBID 1778

Total Maximum Daily Load (TMDL)

2009 303(d) Listed Waterbodies for TMDLs addressed in this report:

WBID	Segment Name	Class and Waterbody Type	Major River Basin	HUC	County	State
1778	Cockroach Bay	Class II	Tampa Bay	03100206	Hillsborough	Florida

TMDL Endpoints/Targets:

Dissolved Oxygen and Nutrients

TMDL Technical Approach:

The TMDL allocations were determined by analyzing the effects of TN, TP, and BOD concentrations and loadings on DO concentrations in the waterbody. A watershed model and estuary model were used to predict delivery of pollutant loads to the waterbody and to evaluate the in-stream impacts of the pollutant loads.

TMDL Waste Load and Load Allocation

Constituent	Current Condition		TMDL Condition		Percent Reduction		
	WLA (kg/yr)	LA (kg/yr)	WLA (kg/yr)	LA (kg/yr)	WLA	LA	MS4
Total Nitrogen	--	6,651	--	2,220	--	67%	67%
Total Phosphorus	--	1,723	--	90	--	95%	95%
Biochemical Oxygen Demand	--	9,261	--	5,727	--	38%	38%

Endangered Species Present (Yes or Blank): Yes

USEPA Lead TMDL (USEPA or Blank): USEPA

TMDL Considers Point Source, Non-point Source, or Both: Both

Major NPDES Discharges to surface waters addressed in USEPA TMDL:

Permit ID	Permittee(s)	County	Permit Type
FLS000006	Hillsborough County, FDOT (District VII)	Hillsborough	Phase I MS4

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1.0 INTRODUCTION

Section 303(d) of the Clean Water Act requires each state to list those waters within its boundaries for which technology based effluent limitations are not stringent enough to protect any water quality standard applicable to such waters. Listed waters are prioritized with respect to designated use classifications and the severity of pollution. In accordance with this prioritization, states are required to develop Total Maximum Daily Loads (TMDLs) for those water bodies that are not meeting water quality standards. The TMDL process establishes the allowable loadings of pollutants or other quantifiable parameters for a waterbody based on the relationship between pollution sources and in-stream water quality conditions, so that states can establish water quality based controls to reduce pollution from both point and nonpoint sources and restore and maintain the quality of their water resources (USEPA 1991).

The Florida Department of Environmental Protection (FDEP) developed a statewide, watershed-based approach to water resource management. Under the watershed management approach, water resources are managed on the basis of natural boundaries, such as river basins, rather than political boundaries. The watershed management approach is the framework FDEP uses for implementing TMDLs. The state's 52 basins are divided into five groups and water quality is assessed in each group on a rotating five-year cycle. FDEP also established five water management districts (WMD) responsible for managing ground and surface water supplies in the counties encompassing the districts.

For the purpose of planning and management, the WMDs divided the district into planning units defined as either an individual primary tributary basin or a group of adjacent primary tributary basins with similar characteristics. These planning units contain smaller, hydrological based units called drainage basins, which are further divided by FDEP into "water segments". A water segment usually contains only one unique waterbody type (stream, lake, canal, etc.) and is about five square miles. Unique numbers or waterbody identification (WBID) numbers are assigned to each water segment. This TMDL addresses WBID 1778, which is a Group 1 waterbody located in the Coastal Middle Tampa Bay Tributary Planning Unit and is managed by the Southwest Florida Water Management District (SWFWMD). WBID 1778 is impaired for dissolved oxygen (DO) and nutrients.

2.0 PROBLEM DEFINITION

To determine the status of surface water quality in Florida, three categories of data – chemistry data, biological data, and fish consumption advisories – were evaluated to determine potential impairments. The level of impairment is defined in the Identification of Impaired Waters Rule (IWR), Section 62-303 of the Florida Administrative Code (FAC). The IWR is FDEP's methodology for determining whether waters should be included on the state's planning list and verified list. Potential impairments are determined by assessing whether a waterbody meets the criteria for inclusion on the planning list. Once a waterbody is on the planning list, additional data and information will be collected and examined to determine if the water should be included on the verified list.

The TMDL addressed in this document is being established pursuant to commitments made by the United States Environmental Protection Agency (USEPA) in the 1998 Consent Decree in the Florida TMDL lawsuit (Florida Wildlife Federation, et al. v. Carol Browner, et al., Civil Action

No. 4: 98CV356-WS, 1998). That Consent Decree established a schedule for TMDL development for waters listed on Florida's USEPA approved 1998 section 303(d) list. The 2009 section 303(d) list identified numerous WBIDs in the Coastal Middle Tampa Bay Tributary Basin as not meeting water quality standards. After assessing all readily available water quality data, USEPA is responsible for developing a TMDL for dissolved oxygen and nutrients WBID 1778, depicted in Figure 2.1.

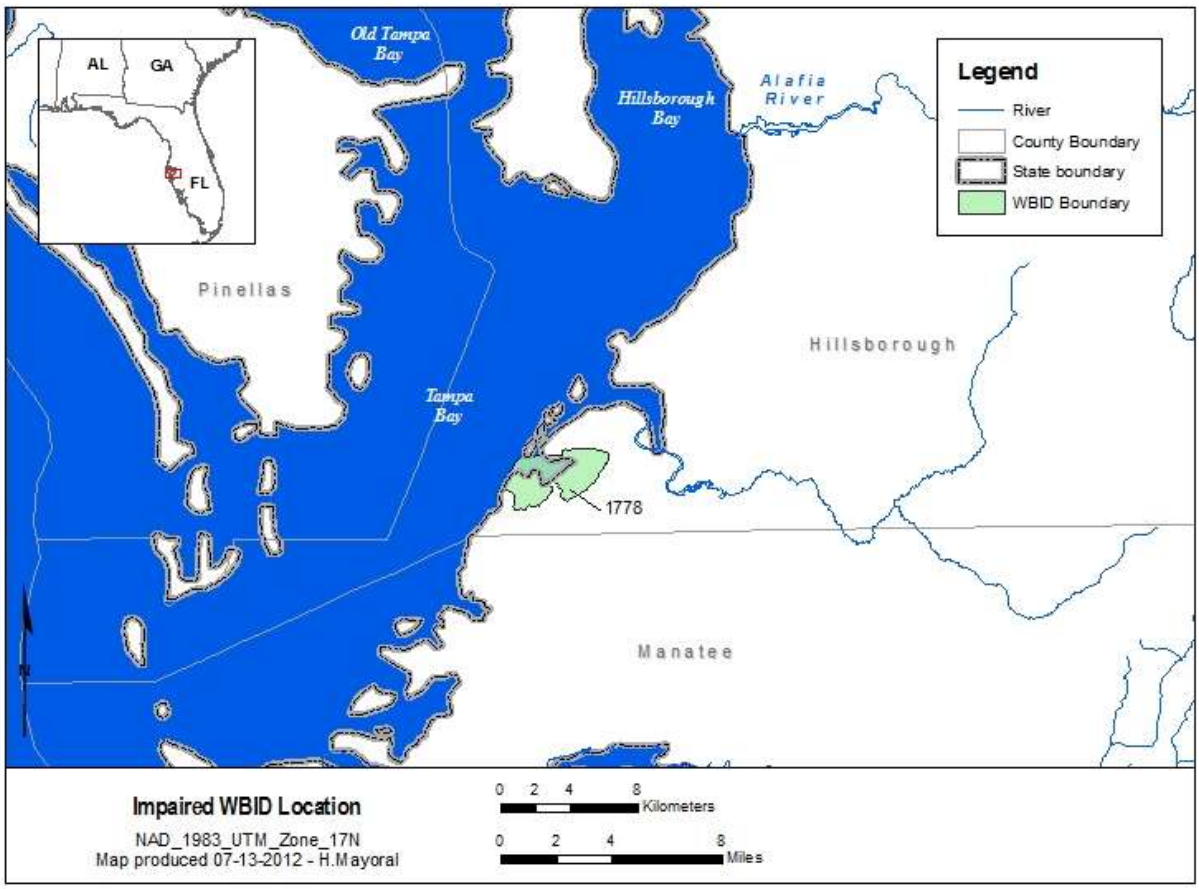

Figure 2.1 Location of the impaired WBID in Cockroach Bay

3.0 WATERSHED DESCRIPTION

3.1 Hydrologic characteristics

Tampa Bay is the largest open-water estuary in Florida, encompassing nearly 400 square miles and bordering three counties—Hillsborough, Manatee, and Pinellas (FDEP). At 2,200 square miles, its watershed is more than five times larger than the bay itself (FDEP 2003). Tampa Bay proper, which includes Old, Middle, and Lower Tampa Bays and Hillsborough Bay, extends approximately 35 miles inland from the Gulf of Mexico and is 5 to 10 miles wide along most of its length.

Cockroach Bay is located in southwest Hillsborough County, just south of where the Little Manatee River discharges to Tampa Bay. A majority of the WBID boundary encompasses a portion of Cockroach Bay Aquatic Preserve, 4,500-acres of estuaries, inlet bays, submerged

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lands, and tidally influenced creeks and rivers (FDEP 2012, FDNR 1987). It is the most pristine of four aquatic preserves within the Tampa Bay region, though submerged lands are leased to the state from the Hillsborough Port Authority (FDEP 2011). Historically, surrounding areas were farmed or quarried for shell (FDEP 2012), and there remain areas of mosquito impoundments (FDNR 1987). The preserve has recently undergone large scale habitat restoration, converting canals and ditches to braided reaches, along with recreated wetlands.

3.2 Climate

Tampa Bay is located on the west coast of Central Florida and experiences a subtropical climate with hot, humid summers and mild, short winters. Average high temperatures in the summer are in the low-90s (°F), and average low temperatures in the winter are in the 50s (°F). An average of 47 inches of rain every year is received in this part of Central Florida, of which a greater percentage falls during the wet season from June through September (SERCC 2012).

3.3 Land Use

A majority of the land use in WBID 1778, Cockroach Bay, is classified as combined wetlands (Figure 3.1 and Table 3.1). Forested and non-forested wetlands combined accounted for 42 percent of the total land use. Developed land uses account for a very small percentage, approximately four percent of the total land use. Pasture and row crops located in the western portion of the WBID account for a combined 23 percent of the total land use. The WBID encompasses the mouth of Cockroach Bay, and open water accounts for an additional 24 percent of the total land use.

The actual drainage area for the Cockroach Bay varies from the WBID boundary (Figure 3.2 and Table 3.2). The United States Geological Survey National Hydrography Dataset (NHD) was used to delineate the drainage area. Acreage between the drainage area and the WBID boundary did not vary considerably, and changes to the distribution of land uses were primarily seen in the amount of agriculture, developed land use, and combined forest land uses. The most pronounced change was observed in row crops, which increased by nearly 100 percent from 524 acres to over 1,000 acres. There were also increases in pasture land use from the contributing drainage area, which increased from 279 acres to over 980 acres.

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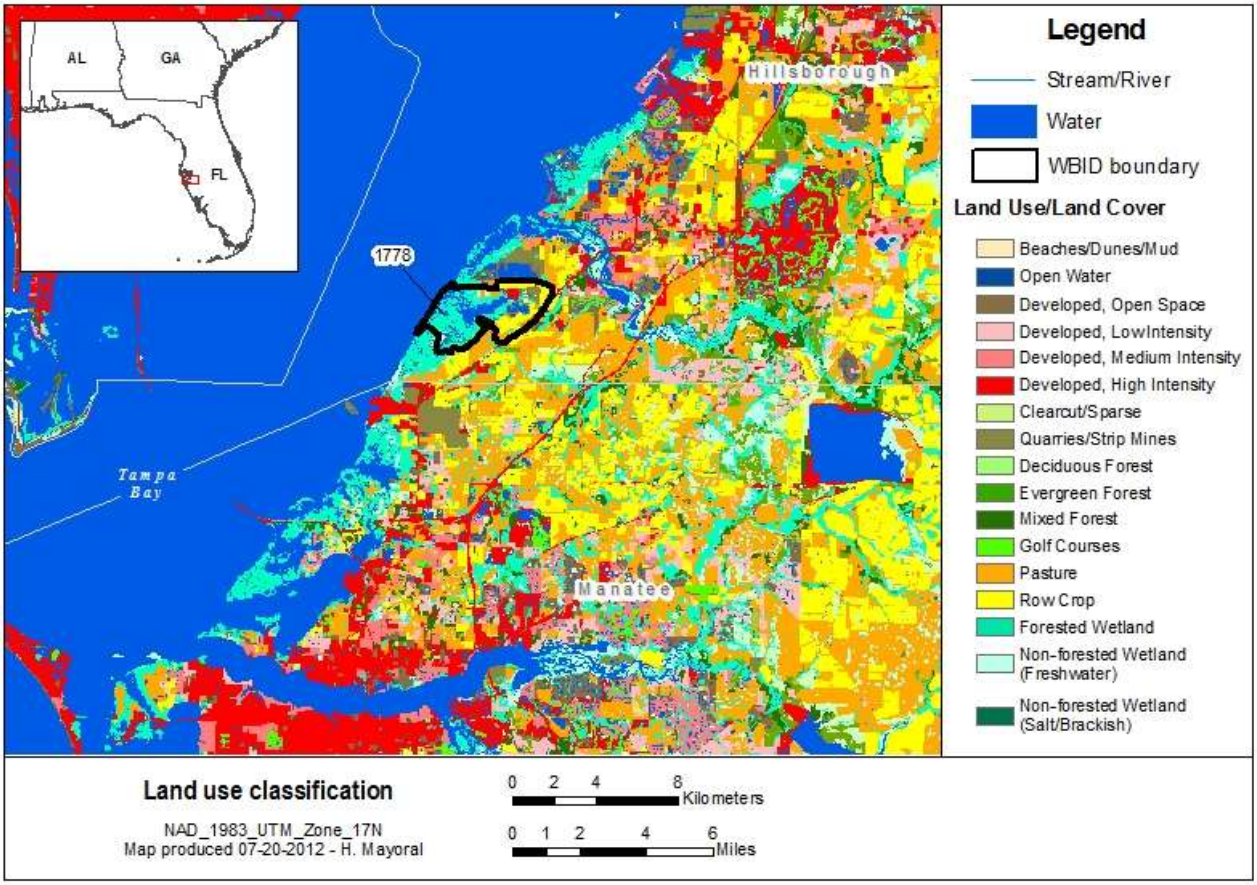

Figure 3.1 Land use for the impaired WBID in Cockroach Bay

Table 3.1 Land use distribution for WBID 1778 in Cockroach Bay

Land Use Classification	WBID 1778	
	Acres	%
Evergreen Forest	49	1%
Deciduous Forest	0	0%
Mixed Forest	6	0%
Forested Wetland	1,163	34%
Non-Forested Wetland (Freshwater)	132	4%
Non-Forested Wetland (Salt/Brackish)	116	3%
Open Water	816	24%
Pasture	279	8%

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Land Use Classification	WBID 1778	
	Acres	%
Row Crop	524	15%
Clear cut Sparse	0	0%
Quarries Strip mines	174	5%
Utility Swaths	0	0%
Developed, Open Space	25	1%
Developed, Low intensity	4	0%
Developed, Medium intensity	5	0%
Developed, High intensity	105	3%
Beaches/Dunes/Mud	0	0%
Golf Courses	0	0%
Totals	3,398	100%

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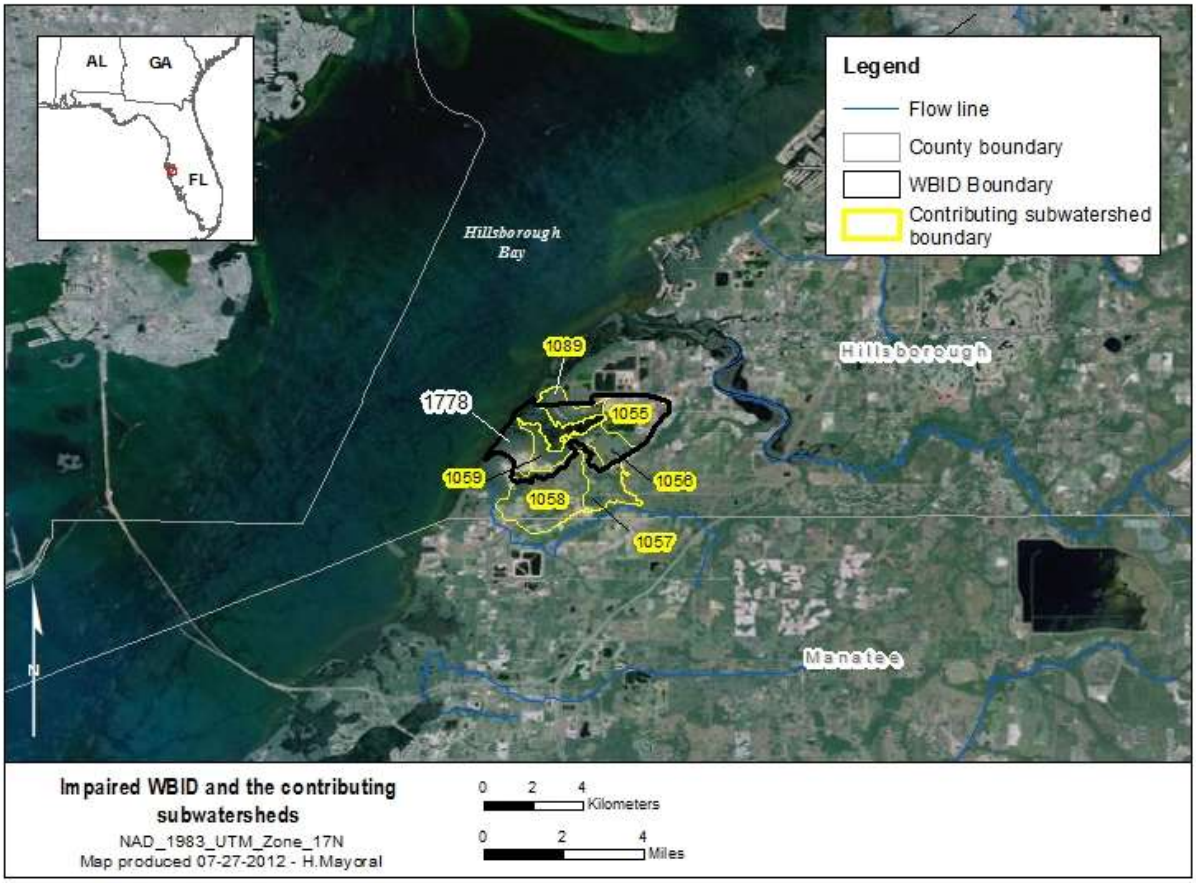

Figure 3.2 Aerial photograph illustrating contributing subwatershed and impaired WBID boundaries

Table 3.2 Land use distribution for contributing subwatersheds in Cockroach Bay

Land Use Classification	Contributing subwatersheds	
	Acres	%
Evergreen Forest	123	3%
Deciduous Forest	0	0%
Mixed Forest	17	0%
Forested Wetland	1,241	27%
Non-Forested Wetland (Freshwater)	150	3%
Non-Forested Wetland (Salt/Brackish)	99	2%
Open Water	287	6%

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Land Use Classification	Contributing subwatersheds	
	Acres	%
Pasture	987	22%
Row Crop	1,077	24%
Clear cut Sparse	46	1%
Quarries Strip mines	174	4%
Utility Swaths	0	0%
Developed, Open Space	41	1%
Developed, Low intensity	4	0%
Developed, Medium intensity	3	0%
Developed, High intensity	283	6%
Beaches/Dunes/Mud	0	0%
Golf Courses	0	0%
Totals	4,532	100%

4.0 WATER QUALITY STANDARDS/TMDL TARGETS

The TMDL reduction scenarios will be done to ensure that the discharge of nutrients is limited as needed to prevent violations of Florida's dissolved oxygen water quality standard of an average daily concentration of 5 mg/L and to ensure balanced flora and fauna. The EPA will establish the TMDL to be consistent with a natural condition if the dissolved oxygen standard cannot be achieved, as Florida's water quality standards provide that FDEP (see section 4.4. Natural Conditions) shall not strive to abate natural conditions.

The waterbody addressed in this report is a Class II marine water having a designated use for Shellfish Propagation and/or Harvesting and Fish Consumption; Recreation, Propagation and Maintenance of a Healthy, Well-Balanced Population of Fish and Wildlife. Designated use classifications are described in Florida's water quality standards in Section 62-302.400, FAC. Water quality criteria for protection of all classes of waters are established in Section 62-302.530, FAC. Individual criteria should be considered in conjunction with other provisions in water quality standards, including Section 62-302.500 FAC, which established minimum criteria

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that apply to all waters unless alternative criteria are specified Section 62-302.530, FAC. In addition, unless otherwise stated, all criteria express the maximum not to be exceeded at any time. The specific criteria addressed in this TMDL document are provided in the following sections.

4.1 Nutrients Criteria

In 1979, FDEP adopted a narrative criterion for nutrients which is applicable to Class I/ II/III waters. FDEP recently adopted numeric nutrient criteria (NNC) for many Class I/II/III waters in the state, including streams, which numerically interprets part of the state narrative criterion for nutrients. FDEP submitted its NNC to EPA for review pursuant to section 303(c) of the CWA. On November 30, 2012, EPA approved those criteria as consistent with the requirements of the CWA. The state criteria, however, are not yet effective for state law purposes.

Also, in November 2010, EPA promulgated numeric nutrient criteria for Class I/II/III inland waters in Florida, including streams. On February 18, 2012, the streams criteria were remanded back to EPA by the U.S. District Court for the Northern District of Florida for further explanation. On November 30, 2012, EPA re-proposed its stream NNC for those flowing waters not covered by Florida's NNC rule.

Therefore, for streams in Florida, the applicable nutrient water quality standard for CWA purposes remains the Class I/II/III narrative criterion.

Florida's recently adopted numeric nutrient criteria interprets the narrative water quality criterion for nutrients in paragraph 62-302.530(47)(b), F.A.C. See section 62-302.531(2). The Florida rule provides that the narrative water quality criteria for nutrients in paragraph 62-302.530(47)(a), F.A.C., continues to apply to all Class I/II/III waters. See section 62-302.531(1).

Florida's recently adopted rule applies to springs, streams, lakes, and some estuary segments. Florida's rule provides that numeric nutrient criteria are expressed as a geometric mean, and concentrations are not to be exceeded more than once in any three calendar year period. Section 62-302.531 and 62-302.532, F.A.C.

Should FDEP's numeric nutrient criteria become applicable water quality standards for CWA purposes before this TMDL is established, EPA will consider the nutrient target necessary to attain section 62-302.532, F.A.C. EPA will compare that target with the target necessary to attain paragraph 62-302.530(47)(a), F.A.C., to determine which target is more stringent.

4.1.1 Narrative Nutrient Criteria

Florida's narrative nutrient criteria provide:

> The discharge of nutrients shall continue to be limited as needed to prevent violations of other standards contained in this chapter. Man induced nutrient enrichment (total nitrogen and total phosphorus) shall be considered degradation in relation to the provisions of Sections 62-302.300, 62-302.700, and 62-4.242 F.A.C. See paragraph 62-302.530(47)(a), F.A.C.
>
> In no case shall nutrient concentrations of a body of water be altered so as to cause an imbalance in natural populations of aquatic flora or fauna. See paragraph 62-302.530(47)(b), F.A.C.

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Chlorophyll and DO levels are often used to indicate whether nutrients are present in excessive amounts. The target for this TMDL is based on levels of nutrients necessary to prevent violations of Florida's DO criterion pursuant to paragraph 62-302.530(47)(a), F.A.C., as set out more fully below.

4.1.2 Development of Numeric Nutrient Criteria

While not yet effective as water quality criteria, the FDEP's numeric nutrient criteria represent the state's most recent interpretation of the second part of Florida's narrative criteria, set out at paragraph 62-302.530(47)(b), F.A.C. See section 62-302.531(2). The first part of the narrative criteria, at paragraph 62-302.530(47)(a), F.A.C., also remains applicable to streams in Florida.

Florida's interpretation of its narrative nutrient criteria applies to streams, including (WBID 1578). For streams that do not have a site specific criteria, the interpretation provides for biological information to be considered together with nutrient thresholds to determine whether a waterbody is attaining See paragraph 62-302.531(2)(c), F.A.C. The rule provides that the nutrient criteria are attained in a stream segment where information on chlorophyll a levels, algal mats or blooms, nuisance macrophyte growth, and changes in algal species composition indicates there are no imbalances in flora and either the average score of at least two temporally independent SCIs performed at representative locations and times is 40 or higher, with neither of the two most recent SCI scores less than 35, or the nutrient thresholds set forth in Table 1 below are achieved. See paragraph 62-302.531(2)(c).

Florida's interpretation provides that nutrient levels should be expressed as a geometric mean, and concentrations are not to be exceeded more than once in any three calendar year period. Section 62-302.200 (25)(e), F.A.C.

Table 1 Inland numeric nutrient criteria

Nutrient Watershed Region	Total Phosphorus Nutrient Threshold	Total Nitrogen Nutrient Threshold
Panhandle West	0.06 mg/L	0.67 mg/L
Panhandle East	0.18 mg/L	1.03 mg/L
North Central	0.30 mg/L	1.87 mg/L
Peninsular	0.12 mg/L	1.54 mg/L
West Central	0.49 mg/L	1.65 mg/L
South Florida	No numeric nutrient threshold. The narrative criterion in paragraph 62-302.530(47)(b), F.A.C., applies.	No numeric nutrient threshold. The narrative criterion in paragraph 62-302.530(47)(b), F.A.C., applies.

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4.2 Dissolved Oxygen Criteria

Numeric criteria for DO are expressed in terms of minimum and daily average concentrations.

The water quality criterion for Class II waters is as follows:

Shall not be less than 5.0 mg/L in a 24-hour period and shall never be less than 4.0 mg/L. Normal daily and seasonal fluctuations above these levels shall be maintained. [FAC 62-302.530 (30)]

4.3 Biochemical Oxygen Demand Criteria

Biochemical Oxygen Demand (BOD) shall not be increased to exceed values which would cause dissolved oxygen to be depressed below the limit established for each class and, in no case, shall it be great enough to produce nuisance conditions. [FAC 62-302.530 (11)]

4.4 Natural Conditions

In addition to the standards for nutrients, DO, and BOD described above, Florida's standards include provisions that address waterbodies which do not meet the standards due to natural background conditions.

Florida's water quality standards provide a definition of natural background:

"Natural Background" shall mean the condition of waters in the absence of man-induced alterations based on the best scientific information available to the Department. The establishment of natural background for an altered waterbody may be based upon a similar unaltered waterbody or on historical pre-alteration data. [FAC 62-302.200(19)]

Florida's water quality standards also provide that:

Pollution which causes or contributes to new violations of water quality standards or to continuation of existing violations is harmful to the waters of this State and shall not be allowed. Waters having water quality below the criteria established for them shall be protected and enhanced. However, the Department shall not strive to abate natural conditions. [FAC 62-302.300(15)]

5.0 WATER QUALITY ASSESSMENT

The WBID addressed in this report was listed as not attaining its designated use on Florida's 2009 303(d) list for dissolved oxygen and nutrients. To determine impairment, an assessment of available data was conducted. The source for current ambient monitoring data was the Impaired Waters Rule (IWR) data Run 44, using data ranging January 1, 2002 to December 31, 2010. The IWR database contains data from various sources within the state of Florida, including the WMDs and counties.

5.1 Water Quality Data

A complete list of water quality monitoring station locations in WBID 1778 are listed in Table 5.1, with their locations illustrated in Figure 5.1. An analysis of water quality data is documented in

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Table 5.2. Water quality data for the WBID can be found below in Figure 5.2 through Figure 5.6, with the data from all water quality stations compiled in each figure.

5.1.1 Dissolved Oxygen

There are several factors that affect the concentration of dissolved oxygen (DO) in a waterbody. Natural DO levels are a function of water temperature, water depth and velocity, salinity and relative contributions from groundwater. Oxygen can be introduced by wind, diffusion, photosynthesis, and additions of higher DO water (e.g. from tributaries). DO concentrations can be lowered by processes that use up oxygen from the water, such as respiration and decomposition, and can be lowered through additions of water with lower DO (e.g. swamp or groundwater). Decomposition of organic matter, such as dead plants and animals, also consumes DO. The dissolved oxygen minimum concentration was 0.50 mg/L, and the maximum concentration was 30.39 mg/L. The mean concentration was 5.65 mg/L.

5.1.2 Biochemical Oxygen Demand

BOD is a measure of the amount of oxygen used by bacteria as they stabilize organic matter. The process can be accelerated when there is an overabundance of nutrients, increasing the aerobic bacterial activity in a waterbody. In turn, the levels of DO can become depleted, eliminating oxygen essential for biotic survival, and potentially causing extensive fish kills. Additionally, BOD is used as an indicator to determine the presence and magnitude of organic pollution from sources such as septic tank leakage, fertilizer runoff, and wastewater effluent. The mean BOD concentration for WBID 1778 was 2.96 mg/L. The maximum BOD concentration was 9.0 mg/L and the minimum concentration was 0.20 mg/L.

5.1.3 Nutrients

Excessive nutrients in a waterbody can lead to overgrowth of algae and other aquatic plants such as phytoplankton, periphyton and macrophytes. This process can deplete oxygen in the water, adversely affecting aquatic life and potentially restricting recreational uses such as fishing and boating. For the nutrient assessment the monitoring data for total nitrogen, total phosphorus and chlorophyll-a are presented. The current standards for nutrients in estuarine bodies are narrative criteria, while numeric standards have been developed for freshwater bodies. The purpose of the nutrient assessment is to present the range, variability and average conditions for the WBID.

5.1.3.1 Total Nitrogen

Total Nitrogen (TN) is comprised of nitrate (NO3), nitrite (NO2), organic nitrogen and ammonia nitrogen (NH4). Though nitrogen is a necessary nutrient required for the growth of most plants and animals, not all forms are readily used or metabolized. Increased levels of organic nitrogen can occur from the decomposition of aquatic life or from sewage, while inorganic forms are generally from erosion and fertilizers. Nitrates are components of industrial fertilizers, yet can also be naturally present in soil, and are converted to nitrite by microorganisms in the environment. Surface runoff from agricultural lands can increase the natural presence of nitrates in the environment and can lead to eutrophication. Usually, the eutrophication process is observed as a change in the structure of the algal community and includes severe algal blooms that may cover large areas for extended periods. Large algal blooms are generally followed by depletion in DO concentrations as a result of algal decomposition. The total nitrogen minimum

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concentration was 0.33 mg/L, and the maximum concentration was 4.14 mg/L. The mean total nitrogen concentration in WBID 1778 was 0.99 mg/L.

5.1.3.2 Total Phosphorus

In natural waters, total phosphorus exists in either soluble or particulate forms. Dissolved phosphorus includes inorganic and organic forms, while particulate phosphorus is made up of living and dead plankton, and adsorbed, amorphous, and precipitated forms. Inorganic forms of phosphorus include orthophosphate and polyphosphates, though polyphosphates are unstable and convert to orthophosphate over time. Orthophosphate is both stable and reactive, making it the form most used by plants. Excessive phosphorus can lead to overgrowth of algae and aquatic plants, the decomposition of which depletes oxygen in the water. The total phosphorus minimum concentration was 0.08 mg/L, and the maximum concentration was 0.55 mg/L. The mean total phosphorus concentration in WBID 1778 was 0.26 mg/L.

5.1.3.3 Chlorophyll-a

Chlorophyll is the green pigment in plants that allows them to create energy from light. In a water sample, chlorophyll is indicative of the presence of algae, and chlorophyll-*a* is a measure of the active portion of total chlorophyll. Corrected chlorophyll refers to chlorophyll-*a* measurements that are corrected for the presence of pheophytin, a natural degradation product of chlorophyll that can interfere with analysis because it has an absorption peak in the same spectral region. It is used as a proxy indicator of water quality because of its predictable response to nutrient availability. Increases in nutrients can potentially lead to blooms in phytoplankton biomass, affecting water quality and ecosystem health. The corrected chlorophyll a maximum concentration was 163.40 µg/L, and the mean concentration was 14.58 µg/L.

Table 5.1 Water quality stations located in WBID 1778

WBID	Station Number
1778	21FLHILL136
	21FLTPA 274048208229581

Table 5.2 Water quality data for WBID 1778

Parameter	Stats	WBID 1778
BOD, 5 Day, 20°C (mg/L)	# of obs	101
	min	0.20
	max	9.00
	mean	2.96
	Geomean	2.50

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Parameter	Stats	WBID 1778
DO, Analysis by Probe (mg/L)	# of obs	134
	min	0.50
	max	30.39
	mean	5.65
	Geomean	4.88
Nitrogen, Total (mg/L as N)	# of obs	129
	min	0.33
	max	4.14
	mean	0.99
	Geomean	0.90
Phosphorus, Total (mg/L as P)	# of obs	133
	min	0.08
	max	0.55
	mean	0.26
	Geomean	0.25
Chlorophyll-A-corrected (µg/L)	# of obs	61
	min	1.00
	max	163.40
	mean	14.58
	Geomean	8.09

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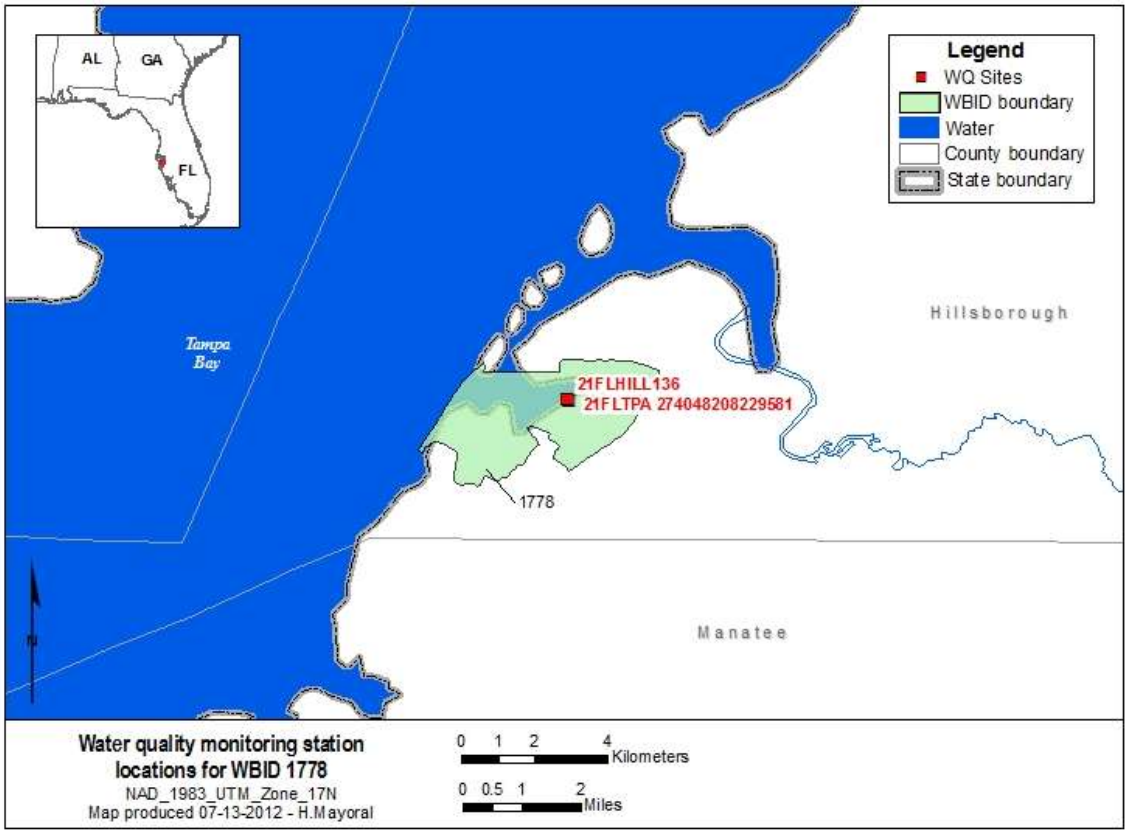

Figure 5.1 Water quality monitoring station locations for impaired WBID 1778 in the Tampa Bay basin

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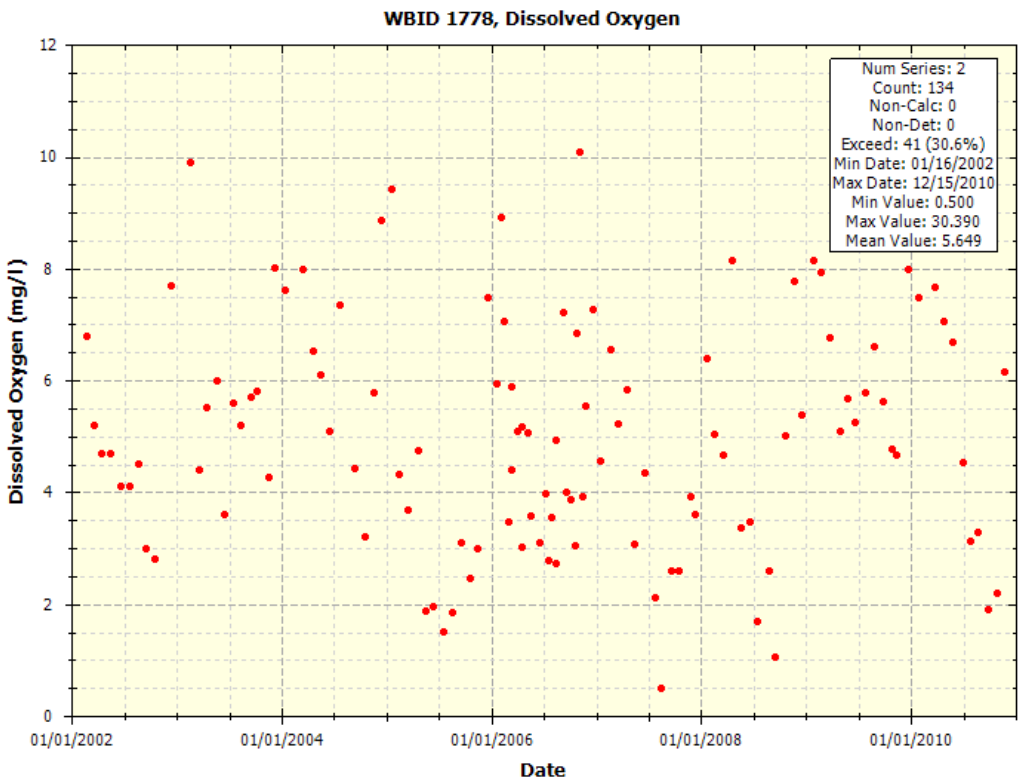

Figure 5.2 Dissolved oxygen concentrations for WBID 1778

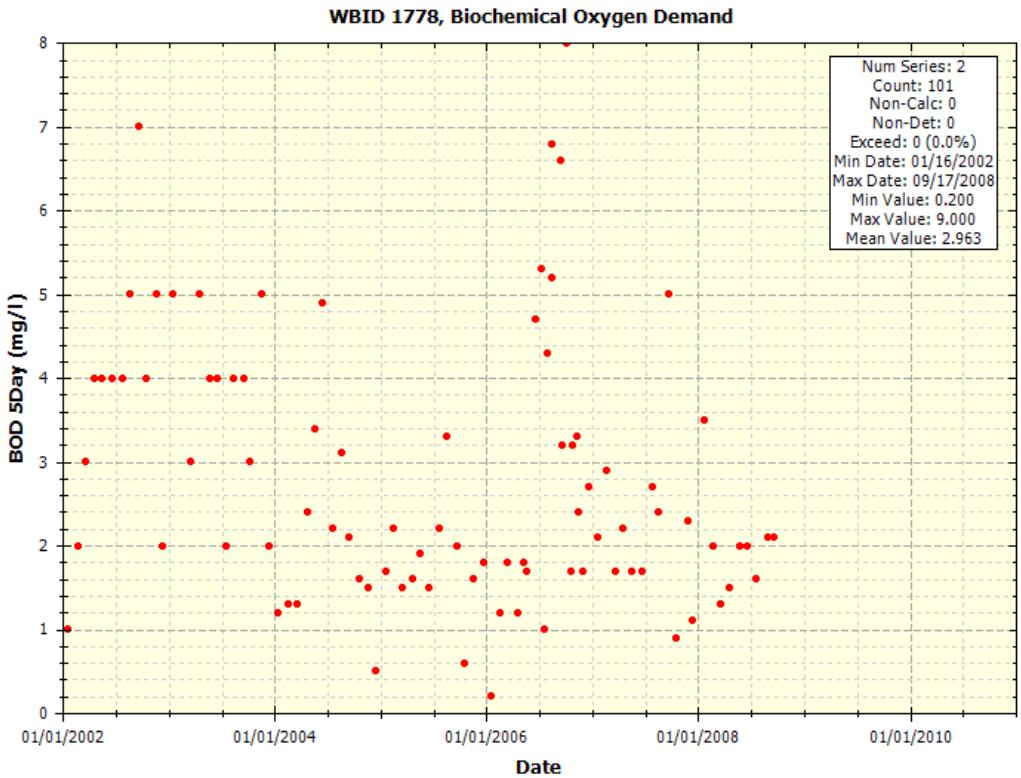

Figure 5.3 Biochemical oxygen demand concentrations for WBID 1778

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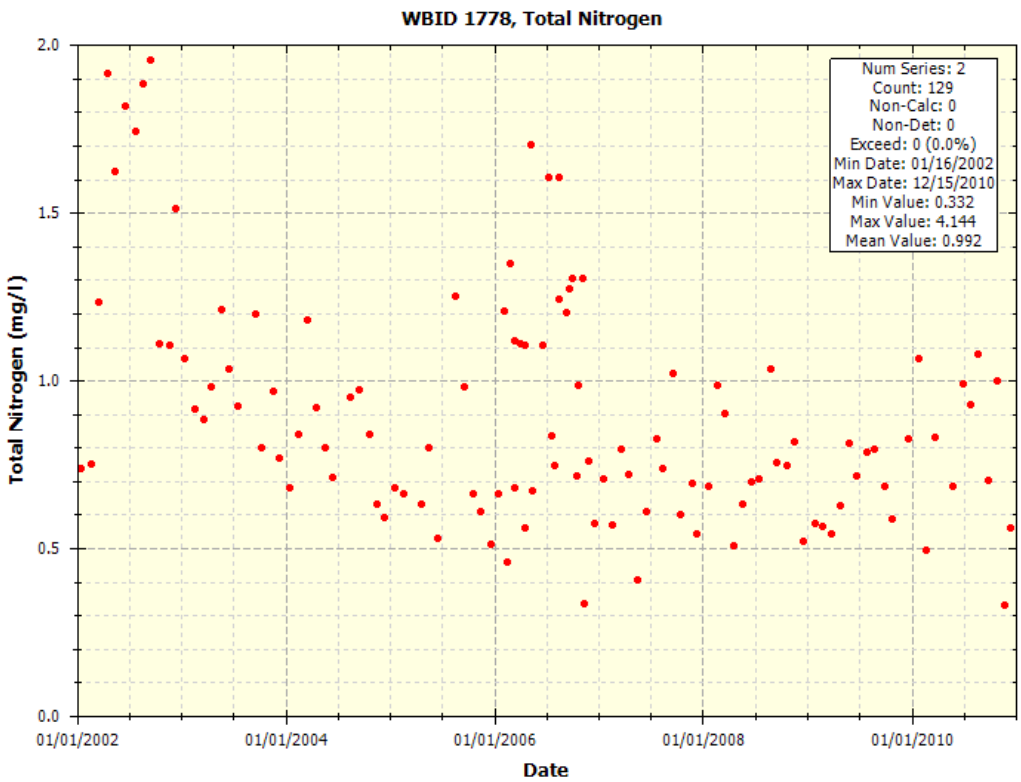

Figure 5.4 Total nitrogen concentrations for WBID 1778

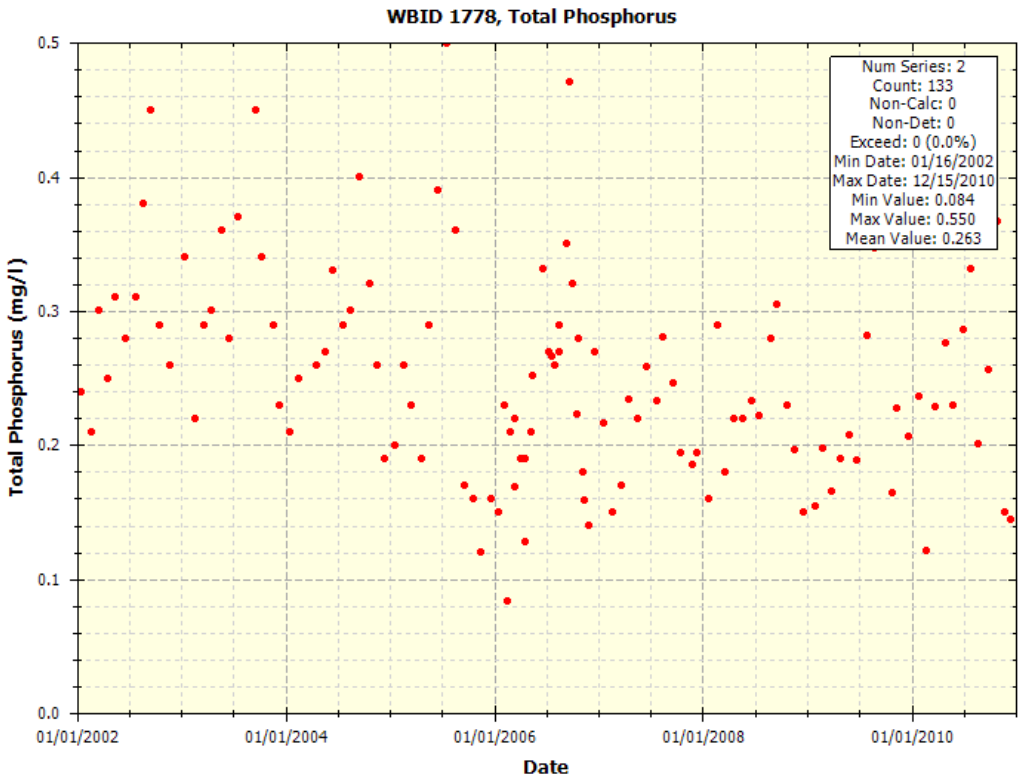

Figure 5.5 Total phosphorus concentrations for WBID 1778

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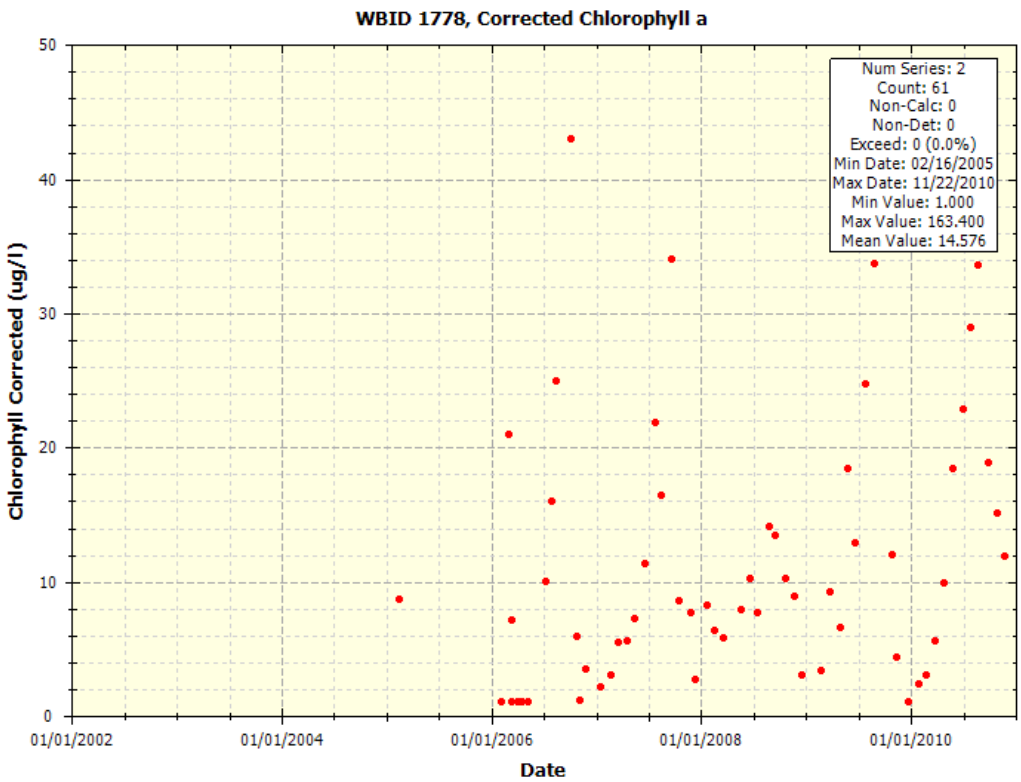

Figure 5.6 Corrected chlorophyll a concentrations for WBID 1778

6.0 SOURCE AND LOAD ASSESSMENT

An important part of the TMDL analysis is the identification of source categories, source subcategories, or individual sources of pollutants in the watershed and the amount of loading contributed by each of these sources. Sources are broadly classified as either point or nonpoint sources. Nutrients can enter surface waters from both point and nonpoint sources.

6.1 Point Sources

A point source is defined as a discernable, confined, and discrete conveyance from which pollutants are or may be discharged to surface waters. Point source discharges of industrial wastewater and treated sanitary wastewater must be authorized by National Pollutant Discharge Elimination System (NPDES) permits. NPDES permitted discharges include continuous discharges such as wastewater treatment facilities as well as some stormwater driven sources such as municipal separate stormwater sewer systems (MS4s), certain industrial facilities, and construction sites over one acre.

6.1.1 Wastewater/Industrial Permitted Facilities

A TMDL wasteload allocation (WLA) is given to wastewater and industrial NPDES permitted facilities discharging to surface waters within an impaired watershed. There are no NPDES-permitted facilities in WBID 1778, therefore no WLA for the impaired WBID was calculated.

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6.1.2 Stormwater Permitted Facilities/MS4s

MS4s are point sources also regulated by the NPDES program. According to 40 CFR 122.26(b)(8), an MS4 is "a conveyance or system of conveyances (including roads with drainage systems, municipal streets, catch basins, curbs, gutters, ditches, man-made channels, or storm drains):

(i) Owned or operated by a State, city, town, borough, county, parish, district, association, or other public body (created by or pursuant to State law)...including special districts under State law such as a sewer district, flood control district or drainage district, or similar entity, or an Indian tribe or an authorized Indian tribal organization, or a designated and approved management agency under section 208 of the Clean Water Act that discharges into waters of the United States.

(ii) Designed or used for collecting or conveying storm water;

(iii) Which is not a combined sewer; and

(iv) Which is not part of a Publicly Owned Treatment Works

MS4s may discharge nutrients and other pollutants to waterbodies in response to storm events. In 1990, USEPA developed rules establishing Phase I of the NPDES stormwater program, designed to prevent harmful pollutants from being washed by stormwater runoff into MS4s (or from being dumped directly into the MS4) and then discharged from the MS4 into local waterbodies. Phase I of the program required operators of "medium" and "large" MS4s (those generally serving populations of 100,000 or greater) to implement a stormwater management program as a means to control polluted discharges from MS4s. Approved stormwater management programs for medium and large MS4s are required to address a variety of water quality related issues including roadway runoff management, municipal owned operations, hazardous waste treatment, etc.

Phase II of the rule extends coverage of the NPDES stormwater program to certain "small" MS4s. Small MS4s are defined as any MS4 that is not a medium or large MS4 covered by Phase I of the NPDES stormwater program. Only a select subset of small MS4s, referred to as "regulated small MS4s", requires an NPDES stormwater permit. Regulated small MS4s are defined as all small MS4s located in "urbanized areas" as defined by the Bureau of the Census, and those small MS4s located outside of "urbanized areas" that are designated by NPDES permitting authorities.

In October 2000, USEPA authorized FDEP to implement the NPDES stormwater program in all areas of Florida except Indian tribal lands. FDEP's authority to administer the NPDES program is set forth in Section 403.0885, Florida Statutes (FS). The three major components of NPDES stormwater regulations are:

- MS4 permits that are issued to entities that own and operate master stormwater systems, primarily local governments. Permittees are required to implement comprehensive stormwater management programs designed to reduce the discharge of pollutants from the MS4 to the maximum extent practicable.
- Stormwater associated with industrial activities, which is regulated primarily by a multisector general permit that covers various types of industrial facilities. Regulated

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industrial facilities must obtain NPDES stormwater permit coverage and implement appropriate pollution prevention techniques to reduce contamination of stormwater.

- Construction activity general permits for projects that ultimately disturb one or more acres of land and which require the implementation of stormwater pollution prevention plans to provide for erosion and sediment control during construction.

Stormwater discharges conveyed through the storm sewer system covered by the permit are subject to the WLA of the TMDL. Any newly designated MS4s will also be required to achieve the percent reduction allocation presented in this TMDL. There is one MS4 permit associated with the impaired WBID, for Hillsborough County as a Phase I C MS4 permit, which falls under permit number FLS000006. The permit also falls under the District VII Florida Department of Transportation.

6.2 Nonpoint Sources

Nonpoint sources of pollution are diffuse sources that cannot be identified as entering a waterbody through a discrete conveyance at a single location. For nutrients, these sources include runoff of agricultural fields, golf courses, and lawns, septic tanks, and residential developments outside of MS4 areas. Nonpoint source pollution generally involves a buildup of pollutants on the land surface that wash off during rain events and as such, represent contributions from diffuse sources, rather than from a defined outlet. Potential nonpoint sources are commonly identified, and their loads estimated, based on land cover data. Most methods calculate nonpoint source loadings as the product of the water quality concentration and runoff water volume associated with certain land use practices. The mean concentration of pollutants in the runoff from a storm event is known as the event mean concentration. Figure 3.1 provides a map of the land use, while Table 3.1 lists the land use distribution in the WBID.

The following sections are organized by land use. Each section provides a description of the land use, the typical sources of nutrient loading (if applicable), and typical total nitrogen and total phosphorus event mean concentrations.

6.2.1 Urban Areas

Urban areas include land uses such as residential, industrial, extractive and commercial. Land uses in this category typically have somewhat high total nitrogen event mean concentrations and average total phosphorus event mean concentrations. Nutrient loading from MS4 and non-MS4 urban areas is attributable to multiple sources including stormwater runoff, leaks and overflows from sanitary sewer systems, illicit discharges of sanitary waste, runoff from improper disposal of waste materials, leaking septic systems, and domestic animals.

In 1982, Florida became the first state in the country to implement statewide regulations to address the issue of nonpoint source pollution by requiring new development and redevelopment to treat stormwater before it is discharged. The Stormwater Rule, as outlined in Chapter 403 FS, was established as a technology-based program that relies upon the implementation of Best Management Practices (BMPs) that are designed to achieve a specific level of treatment (i.e., performance standards) as set forth in Chapter 62-40, FAC.

Florida's stormwater program is unique in having a performance standard for older stormwater systems that were built before the implementation of the Stormwater Rule in 1982. This rule

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states: "the pollutant loading from older stormwater management systems shall be reduced as needed to restore or maintain the beneficial uses of water." [FAC 62-40-.432(2)(c)]

Nonstructural and structural BMPs are an integral part of the State's stormwater programs. Nonstructural BMPs, often referred to as "source controls", are those that can be used to prevent the generation of nonpoint source pollutants or to limit their transport off-site. Typical nonstructural BMPs include public education, land use management, preservation of wetlands and floodplains, and minimization of impervious surfaces. Technology-based structural BMPs are used to mitigate the increased stormwater peak discharge rate, volume, and pollutant loadings that accompany urbanization.

Urban, residential, and commercial developments are often a significant nonpoint source of nutrients and oxygen-demanding substances. In WBID 1778, seven percent of the contributing land use is developed. Approximately six percent of the total land acreage from the contributing drainage area is high intensity development.

Onsite Sewage Treatment and Disposal Systems (Septic Tanks)

As stated above, leaking septic tanks or onsite sewage treatment and disposal systems (OSTDs) can contribute to nutrient loading in urban areas. Water from OSTDs is typically released to the ground through on-site, subsurface drain fields or boreholes that allow the water from the tank to percolate (usually into the surficial aquifers) and either transpire to the atmosphere through surface vegetation or add to the flow of shallow ground water. When properly sited, designed, constructed, maintained, and operated, OSTDs are a safe means of disposing of domestic waste. The effluent from a well-functioning OSTD receives natural biological treatment in the soil and is comparable to secondarily treated wastewater from a sewage treatment plant. When not functioning properly, OSTDs can be a source of nutrients, pathogens, and other pollutants to both ground water and surface water.

The State of Florida Department of Health publishes data on new septic tank installations and the number of septic tank repair permits issued for each county in Florida. County estimates have been made for the number of septic systems installed in Hillsborough County since the 1970 census and the total number of repair permits issued for the ten years between 1999-2000 and 2009-2010 (FDOH 2009). Hillsborough County has had approximately 107,198 septic tanks installed since 1970, and over 15,437 repair permits issued since the year 2000 (data do not reflect septic tanks removed from service). Leaking septic systems could be a relevant source of organic and nutrient loading in the watershed.

6.2.2 Pastures

Pastures include cropland and improved and unimproved pasturelands, such as non-tilled grasses woodland pastures, feeding operations, nurseries and vineyards; as well as specialty farms. Agricultural activities, including runoff of fertilizers or animal wastes from pasture and cropland and direct animal access to streams, can generate nutrient loading to streams. The highest total nitrogen and total phosphorus event mean concentrations are often associated with agricultural land uses. Pasture land use is one of the largest land uses, accounting for 22 percent of the total land acreage contributing to WBID 1778, and is therefore a likely source of excessive nutrients.

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6.2.3 Clear cut/Sparse

The clear cut/sparse land use classification includes recent clear cuts, areas of sparse vegetation or herbaceous dry prairie, shrub and brushland, other early successional areas, and mixed rangeland. Event mean concentrations for clear cut/sparse can be relatively low for total nitrogen and total phosphorus. There is a very small amount of clear cut/sparse land uses in the contributing watershed to WBID 1778 (less than one percent).

6.2.4 Forests

Upland forests include flatwoods, oak, various types of hardwoods, conifers and tree plantations. Wildlife located within forested areas deposit their feces onto land surfaces where it can be transported to nearby streams during storm events. Generally, the pollutant load from wildlife is assumed to represent background concentrations. Event mean concentrations for upland forests are low for both total nitrogen and total phosphorus. Combined forested land use accounts for three percent of the contributing land use to WBID 1778.

6.2.5 Water and Wetlands

Water and Wetlands often have very low nutrient loadings, although decaying organic matter in wetlands can contribute to high organic nutrient concentrations. Open water accounts for only 6 percent of total land use contributing to WBID 1778. Both forested and non-forested wetlands combined account for 31 percent of total land use contributing to WBID 1778.

6.2.6 Quarries/Strip mines

This land use classification includes quarries, strip mines, exposed rock and soil, fill areas, reclaimed lands, and holding ponds. Event mean concentrations for some barren lands tend to be higher in total nitrogen. Four percent of the land acreage contributing to WBID 1778 is comprised of quarries/strip mine land uses.

7.0 ANALYTICAL APPROACH

In the development of a TMDL there needs to be a method for relating current loadings to the observed water quality problem. This relationship could be: statistical (regression for a cause and effect relationship), empirical (based on observations not necessarily from the waterbody in question) or mechanistic (physically and/or stochastically based) that inherently relates cause and effect using physical and biological relationships.

Mechanistic models were used in the development of the Cockroach Bay TMDL to relate the physical and biological relationships. A dynamic watershed model was used to predict the quantity of water and pollutants associated with runoff from rain events. The watershed model was linked to a hydrodynamic model that simulated tidal influences in the river. Both models were linked to a water quality simulation model that integrated the loadings and flow from the watershed model with flow from the hydrodynamic model to predict the water quality in the receiving waterbodies.

The period of simulation that was considered in the development of this TMDL is January 1, 2002 to December 31, 2009. The models were used to predict time series for BOD, TN, TP, and DO. The models were calibrated to current conditions and were then used to predict improvements in water quality as function of reductions in loadings.

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7.1 Mechanistic Models

7.1.1 Loading Simulation Program C++ (LSPC)

LSPC is the Loading Simulation Program in C++, a watershed modeling system that includes streamlined Hydrologic Simulation Program Fortran (HSPF) algorithms for simulating hydrology, sediment, and general water quality overland as well as a simplified stream fate and transport model. LSPC is derived from the Mining Data Analysis System (MDAS), which was originally developed by USEPA Region 3 (under contract with Tetra Tech) and has been widely used for TMDLs. In 2003, the USEPA Region 4 contracted with Tetra Tech to refine, streamline, and produce user documentation for the model for public distribution. LSPC was developed to serve as the primary watershed model for the USEPA TMDL Modeling Toolbox. LSPC was used to simulate runoff (flow, biological oxygen demand, total nitrogen, total phosphorus and dissolved oxygen) from the land surface using a daily timestep for current and natural conditions. LSPC provided tributary flows and temperature to the EFDC estuary models and tributary water quality concentrations to WASP7 estuary models.

An LSPC model was utilized to estimate the nutrient loads within and discharged from the Cockroach Bay watershed. The LSPC model utilized the data inputs, including land use and weather data, from the larger Tampa Bay Watershed model (USEPA 2012a and USEPA 2012b).

In order to evaluate the contributing sources to a waterbody and to represent the spatial variability of these sources within the watershed model, the contributing drainage area was represented by a series of sub-watersheds for each of the models. The sub-watersheds for the Tampa Bay Watershed model were developed using the 12-digit hydrologic unit code (HUC12) watershed data layer and the Geological Survey (USGS) National Hydrograph Dataset (NHD). The sub-watersheds were re-delineated at a smaller scale for the Cockroach Bay Watershed model, once again using the NHD catchments as well as the USGS National Elevation Dataset Digital Elevation Model (Figure 7.1). This figure also shows the location of the impaired WBID within the modeled watershed.

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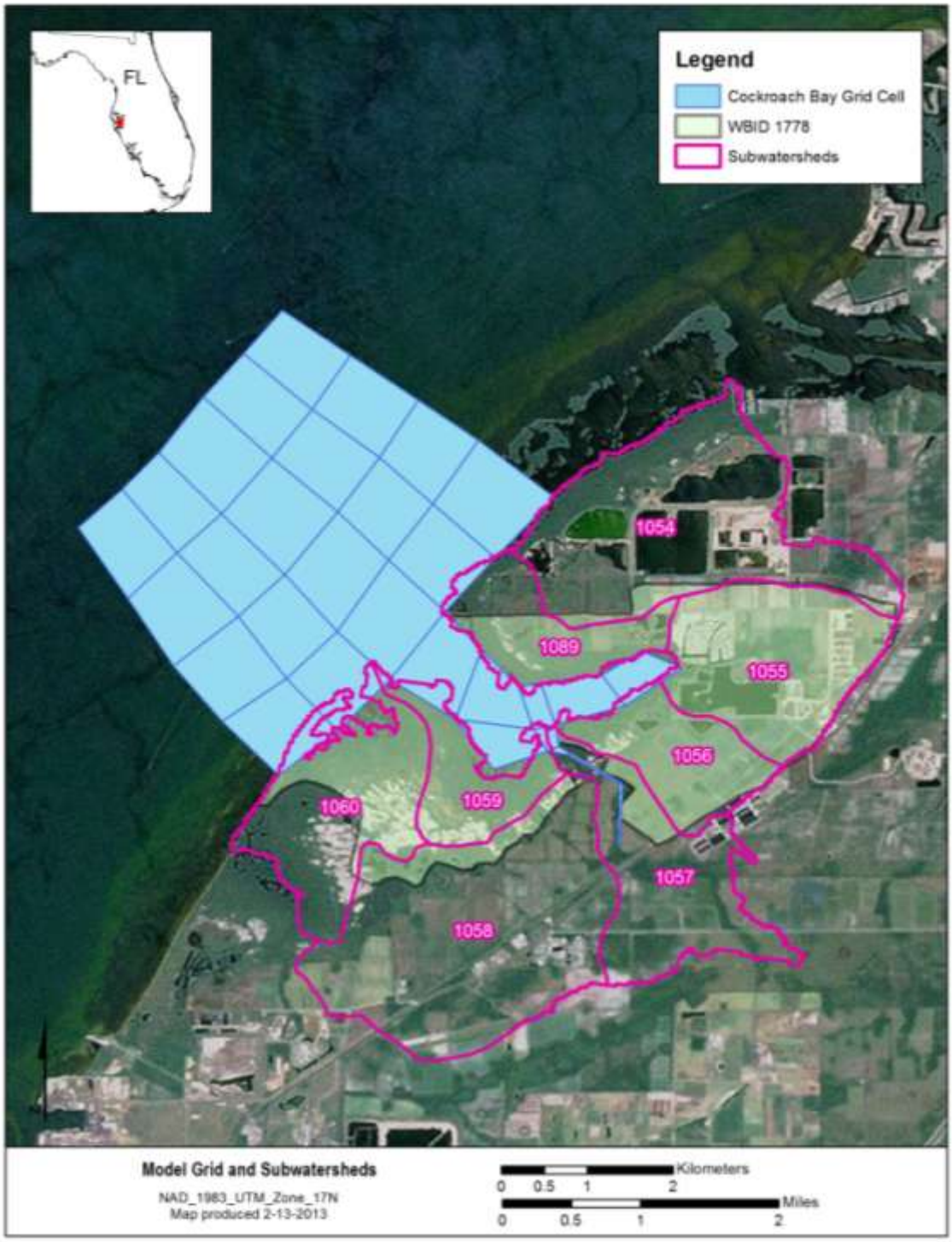

Figure 7.1 Location of Cockroach Bay LSPC subwatersheds

The LSPC model has a representative reach defined for each sub-watershed, and the main channel stem within each sub-watershed was used as the representative reach. The characteristics for each reach include the length and slope of the reach, the channel geometry and the connectivity between the sub-watersheds. Length and slope data for each reach was obtained using the USGS DEM and NHD data.

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The attributes supplied for each reach were used to develop a function table (FTABLE) that describes the hydrology of the stream reach by defining the functional relationship between water depth, surface area, water volume, and outflow in the segment. The assumption of a fixed depth, area, volume, outflow relationship rules out cases where the flow reverses direction or where one reach influences another upstream of it in a time-dependent way. LSPC does not model the tidal flow in the low-lying estuaries, and therefore the main Tampa Bay Watershed model was calibrated to non-tidally influenced USGS gages. The Cockroach Bay Watershed model was linked to the EFDC and WASP models to simulate the areas of the estuary that were tidally influenced.

The watershed model uses land use data as the basis for representing hydrology and nonpoint source loadings. The FDEP Level III Florida Land Use, specifically the Southwest Florida Water Management District (SWFWMD) 2006 dataset, was used to determine the land use representation. The National Landuse Coverage Dataset (NLCD) was used to develop the impervious land use representations.

The SWFWMD coverage utilized a variety of land use classes which were grouped and re-classified into 18 land use categories: beaches/dune/mud, open water, utility swaths, developed open space, developed low intensity, developed medium intensity, developed high intensity, clear-cut/sparse, quarries/strip mines, deciduous forest, evergreen forest, mixed forest, golf courses, pasture, row crop, forested wetland, non-forested wetland (salt/brackish), and non-forested wetland (freshwater). The LSPC model requires division of land uses in each sub-watershed into separate pervious and impervious land units. The 2006 NLCD percent impervious coverage was used to determine the percent impervious area associated with each land use category. Any impervious areas associated with utility swaths, developed open space, and developed low intensity, were grouped together and placed into a new land use category named *low intensity development impervious*. Impervious areas associated with medium intensity development and high intensity development were kept separate and placed into two new categories for *medium intensity development impervious* and *high intensity development impervious*, respectively. Finally, any impervious area not already accounted for in the three developed impervious categories, were grouped together into a fourth new category for all remaining impervious land use.

Soil data for the Florida watersheds was obtained from the Soil Survey Geographic Database (SSURGO). The database was produced and distributed by the Natural Resources Conservation Service (NRCS) - National Cartography and Geospatial Center (NCGC). The SSURGO data was used to determine the total area that each hydrologic soil group covered within each sub-watershed. The sub-watersheds were represented by the hydrologic soil group that had the highest percentage of coverage within the boundaries of the sub-watershed. There were four hydrologic soil groups which varied in their infiltration rates and water storage capacity.

In the watershed models, nonpoint source loadings and hydrological conditions are dependent on weather conditions. Hourly data from weather stations within the boundaries of, or in close proximity to, the sub-watersheds were applied to the watershed model. A weather data forcing file was generated in ASCII format (*.air) for each meteorological station used in the hydrological evaluations in LSPC. Each meteorological station file contained atmospheric data used in modeling the hydrological processes. These data included precipitation, air temperature, dew point temperature, wind speed, cloud cover, evaporation, and solar radiation. These data are

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used directly, or calculated from the observed data. The Tampa Bay Watershed model weather stations contained data through 2009.

The hydrodynamic calibration parameters from the larger Tampa Bay Watershed model were used to populate the Cockroach Bay watershed model. The Tampa Bay Watershed model was calibrated to continuous flow USGS gages. No continuous measured flow data was located in the Cockroach Bay watershed, so no calibration updates were done for flow in Cockroach Bay and the Tampa Bay Watershed model parameterization was used. Additionally, the water quality parameters from the larger Tampa Bay Watershed model were used to populate the Cockroach Bay Watershed model. The Tampa Bay Watershed model was calibrated to several water quality stations whose data was taken from IWR38. The Cockroach Bay watershed was calibrated to water quality data from IWR44, specifically to station 21FLHILL584 and 21FLGW 22086, which both contained data records for the parameters of interest (Figure 7.2). LSPC water quality calibration results are presented in Figure 7.3 through Figure 7.7.

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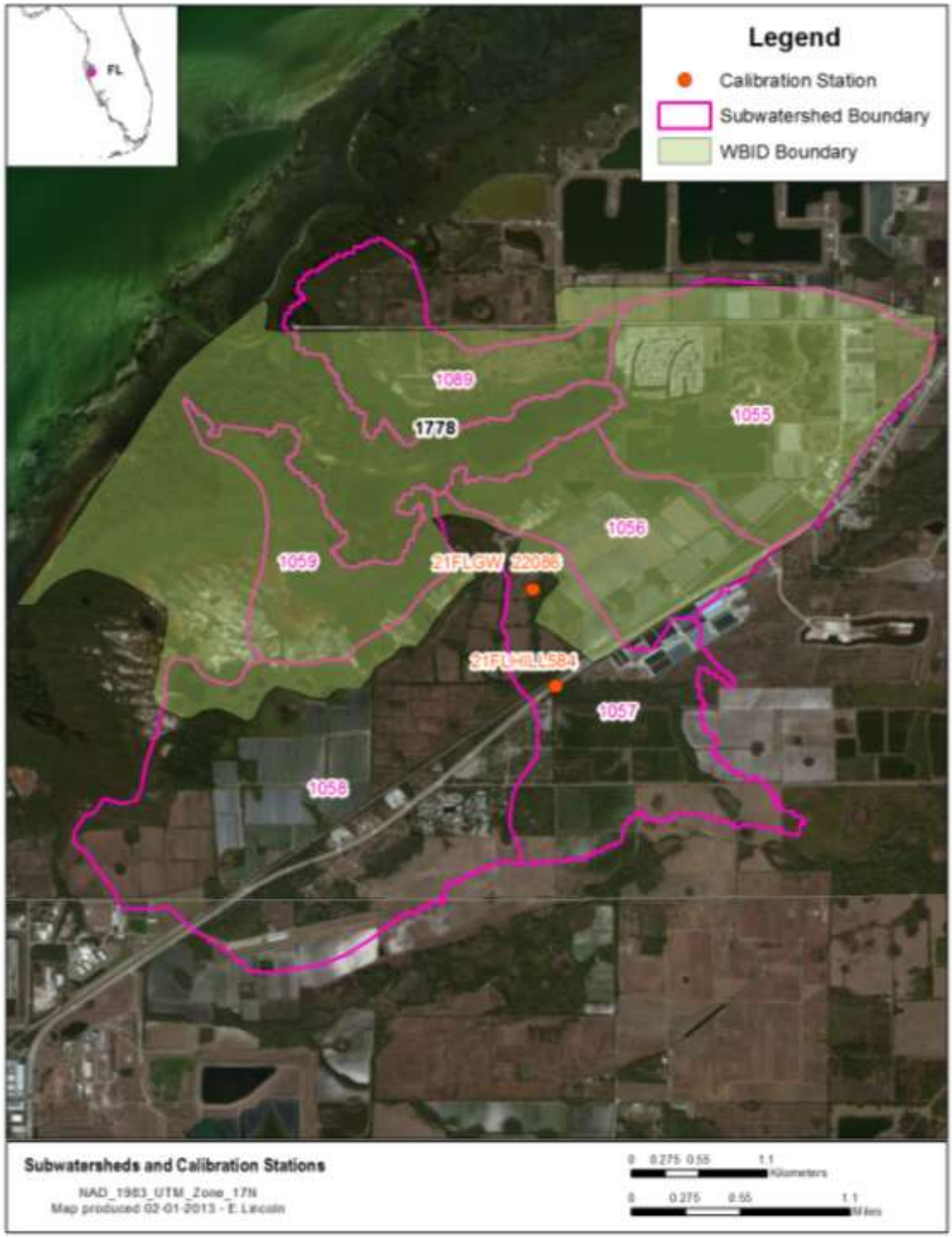

Figure 7.2 Water quality calibration station locations for the Cockroach Bay Watershed model

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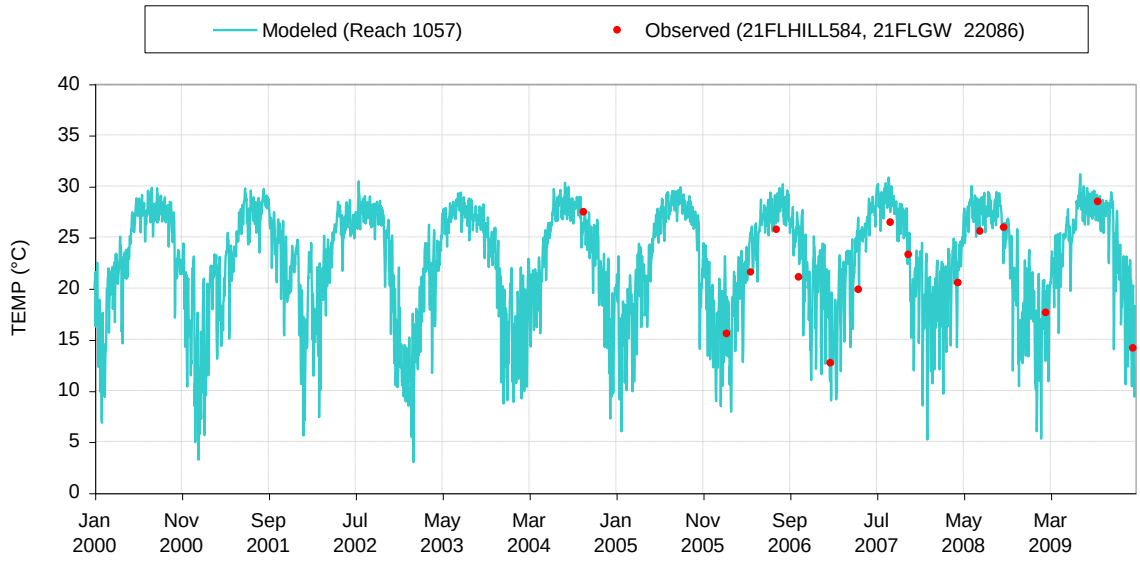

Figure 7.3 Modeled vs. Observed TEMP (°C) at 21FLHILL584 and 21FLGW 22086

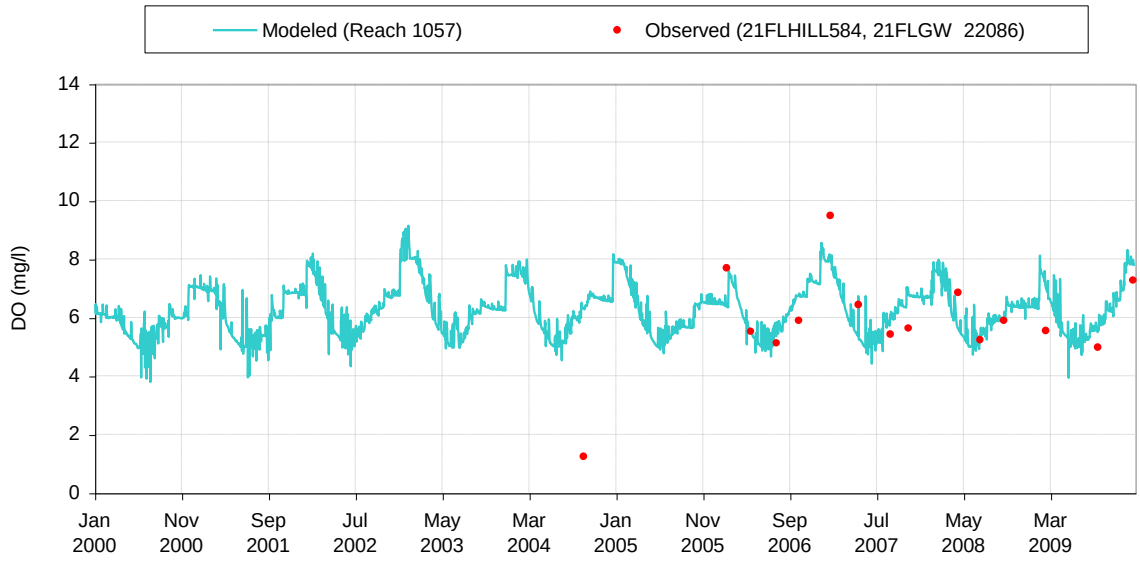

Figure 7.4 Modeled vs. Observed DO (mg/l) at 21FLHILL584 and 21FLGW 22086

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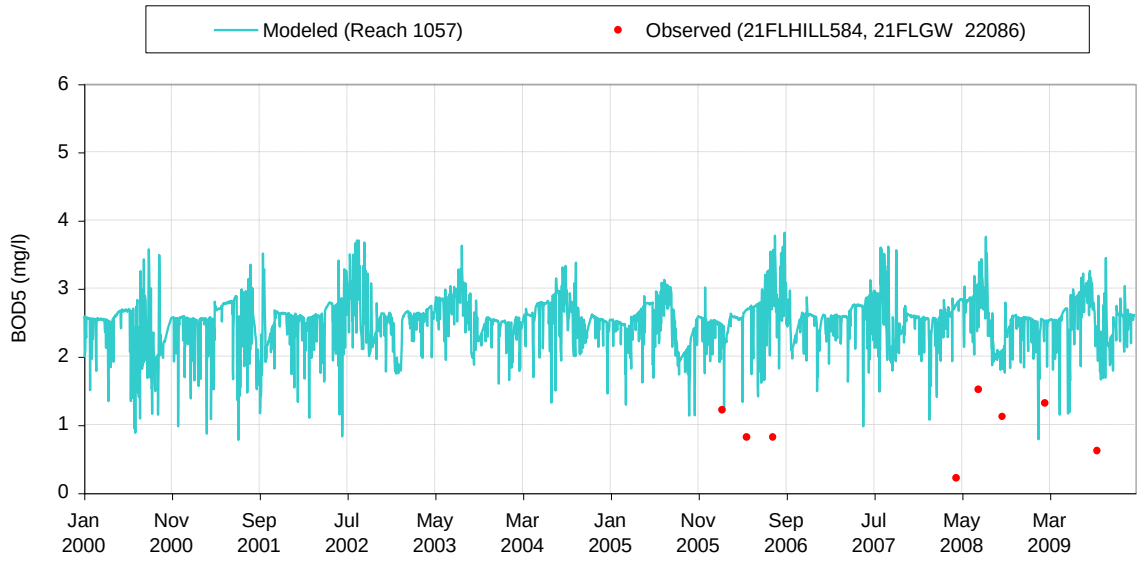

Figure 7.5 Modeled vs. Observed BOD5 (mg/l) at 21FLHILL584 and 21FLGW 22086

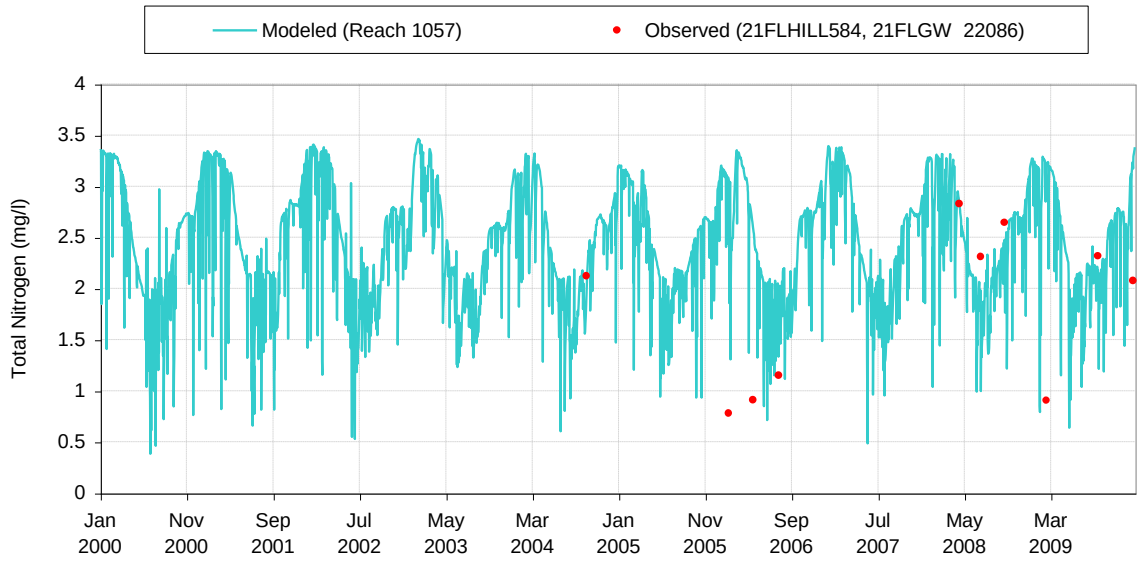

Figure 7.6 Modeled vs. Observed Total Nitrogen (mg/l) at 21FLHILL584 and 21FLGW 22086

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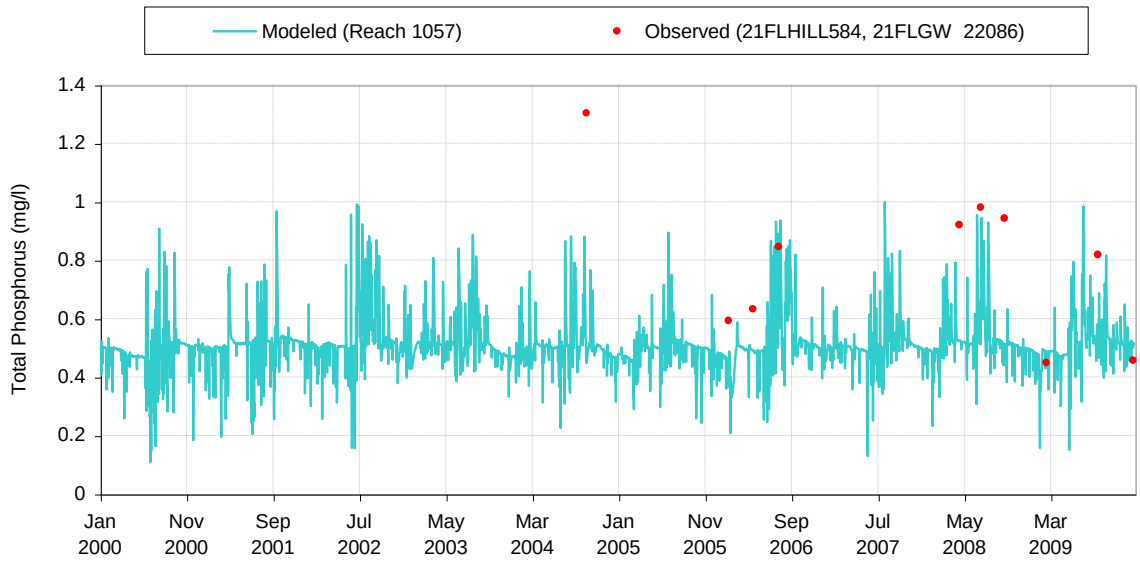

Figure 7.7 Modeled vs. Observed Total Phosphorus (mg/l) at 21FLHILL584 and 21FLGW 22086

7.1.2 Environmental Fluids Dynamic Code (EFDC)

The EFDC model is a part of the USEPA TMDL Modeling Toolbox due to its application in many TMDL-type projects. As such, the code has been peer reviewed and tested and has been freely distributed and supported by Tetra Tech. EFDC was developed by Dr. John Hamrick (Hamrick 1992) and is currently supported by Tetra Tech for USEPA Office of Research and Development (ORD), USEPA Region 4, and USEPA Headquarters. The models, tools, and databases in the TMDL Modeling Toolbox are continually updated and upgraded through TMDL development in Region 4. EFDC is a multifunctional, surface-water modeling system, which includes hydrodynamic, sediment contaminant, and eutrophication components. The EFDC model is capable of 1, 2, and 3-dimensional spatial resolution. The model employs a curvilinear-orthogonal horizontal grid and a sigma or terrain following vertical grid.

The EFDC hydrodynamic model can run independently of a water quality model. The EFDC model simulates the hydrodynamic and constituent transport and then writes a hydrodynamic linkage file for a water quality model such as the Water Quality Analysis Program (WASP7) model. This model linkage, from EFDC hydrodynamics to WASP water quality, has been applied on many USEPA Region 4 projects in support of TMDLs and has been well tested (Wool et al. 2003).

The EFDC model was utilized to simulate three-dimensional circulation dynamics of hydrodynamic state variables (water surface elevation, salinity, and temperature) in the Cockroach Bay estuary. The Cockroach Bay model utilized the Tampa Bay EFDC model that was created for the Florida Numeric Nutrient Criteria (NNC), which was resized to meet the modeling needs of Cockroach Bay. The Tampa Bay EFDC model was created using NOAA bathymetry data.

The EFDC model predicts water surface elevation, salinity, and temperature, in response to a set of multiple factors: wind speed and direction, freshwater discharge, tidal water level fluctuation,

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rainfall, surface heat flux, and temperature and salinity associated with boundary fluxes. Hourly measurements of atmospheric pressure, dry and wet bulb atmospheric temperatures, rainfall rate, wind speed and direction, and fractional cloud cover were obtained from data collected at the Tampa WBAN station 12842 for 2002 through 2009. Solar short wave radiation was calculated using the CE-Qual-W2 method.

The Cockroach Bay Estuary model used hourly water surface elevation time series data from the National Oceanic and Atmospheric Administration (NOAA) tidal stations to simulate tides at the open boundary. Observed temperature data at water quality stations were used to simulate the temperature at the open boundaries, and average salinity in the Tampa Bay Lagoon was used to simulate salinity. The Tampa Bay NNC estuary model was calibrated to measured NOAA tidal stations, and the Tampa Bay model was used to simulate the open boundary conditions in the Cockroach Bay Estuary model. The upstream inland boundary grid cell received LSPC simulated watershed discharges.

The Cockroach Bay EFDC grid consisted of 60 cells, specifically 30 cells in the horizontal direction and was two layers in the vertical direction (Figure 7.8). The grid was developed using bathymetry data from NOAA. Bathymetry was unavailable for the inland, tidally influenced streams, and channel slope from the USGS digital elevation model was used to estimate slope within the channel. The Cockroach Bay grid extended from the Tampa Bay into Cockroach Bay.

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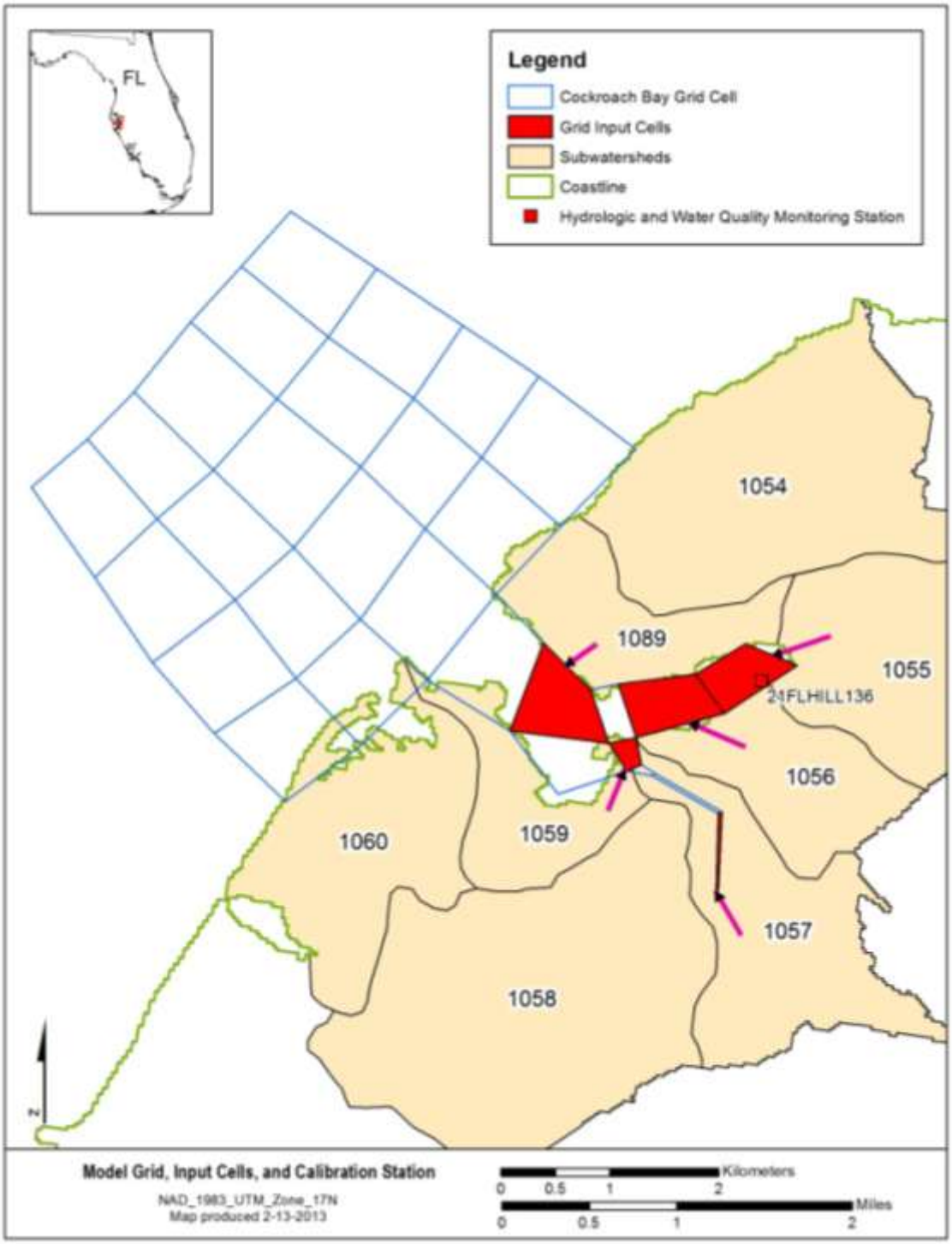

Figure 7.8 LSPC subwatershed boundaries, model grid, and calibration station for Cockroach Bay

Because there were no NOAA tidal stations located within the Cockroach Bay estuary, water surface elevation within the modeled cells could not be directly calibrated. Salinity measurements from IWR44 data were used to review the Cockroach Bay estuary EFDC

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calibration. Following model review, the salinity and temperature parameters were adjusted accordingly (Figure 7.9).

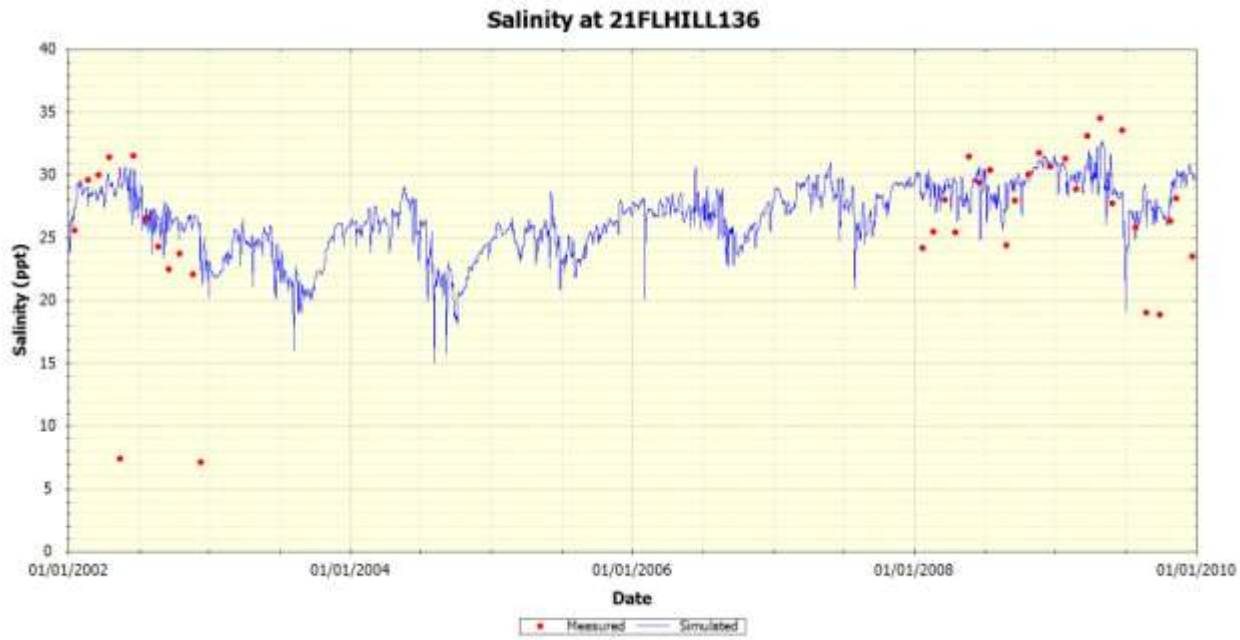

Figure 7.9 Measured verse modeled salinity (ppt) in Cockroach Bay at station 21FLHILL136

7.1.3 Water Quality Analysis Simulation Program (WASP7)

The Water Quality Analysis Simulation Program Version 7.4.1 (WASP7) is an enhanced Windows version of the USEPA Water Quality Analysis Simulation Program (WASP) (Di Toro et al., 1983; Connolly and Winfield, 1984; Ambrose, R.B. et al., 1988), with upgrades to the user's interface and the model's capabilities. The major upgrades to WASP have been the addition of multiple BOD components, addition of sediment diagenesis routines, and addition of periphyton routines. The hydrodynamic file generated by EFDC is compatible with WASP7 and it transfers segment volumes, velocities, temperature and salinity, as well as flows between segments. The time step is set in WASP7 based on the hydrodynamic simulation.

WASP7 helps users interpret and predict water quality responses to natural phenomena and man-made pollution for various pollution management decisions. WASP7 is a dynamic compartment-modeling program for aquatic systems, including both the water column and the underlying benthos. The time-varying processes of advection, dispersion, point and diffuse mass loading and boundary exchange are represented in the basic program. The purpose of the WASP7 water quality modeling was to reproduce the three-dimensional transport and chemical and biological interactions of major components of water quality in the Cockroach Bay estuary. WASP7 modeled total nitrogen (TN) and its speciation, total phosphorus (TP) and its speciation, chlorophyll-a, dissolved oxygen, and carbonaceous biochemical oxygen demand (CBOD). The model predicts these parameters in response to a set of hydrological, meteorological, atmospheric, and chemical and biological factors: loads from point and nonpoint sources, benthic ammonia and phosphate fluxes, sediment oxygen demand (SOD), solar radiation, air temperature, reaeration, offshore and inland boundary conditions.

The Cockroach Bay WASP7 model utilized the same grid cells that were developed for the Cockroach Bay EFDC model. The hydrodynamic simulation from the Cockroach Bay EFDC

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model was input into the WASP7 model. Open boundary water quality conditions used measured water quality data from Tampa Bay. Water quality loading from the LSPC model was used to simulate loads coming from rivers and streams into the estuary.

Because the LPSC model simulated TN, TP, and BOD and the WASP model simulated TN and its speciation, TP and its speciation, and CBOD, the water quality concentrations from LSPC were adjusted for WASP simulation prior to being input into the WASP model. TN was speciated into nitrate-nitrite (NOX), ammonia (NH4), and organic nitrogen (ON), and TP was speciated into orthophosphate (PO4) and organic phosphorus (OP). Water quality data in the Cockroach Bay watershed was reviewed to determine the ratio of NOX, NH4, and ON in TN, and the ratio of PO4 and OP in TP. The in-stream BOD loads from LSPC were converted to ultimate CBOD using an f-ratio of 1.5.

Water quality parameters from the Tampa Bay NNC WASP model were used for initial parameter population for the Cockroach Bay WASP7 model. The Cockroach Bay estuary model calibration was reviewed against water quality data located in IWR44. Following review, the calibration was adjusted accordingly to provide the best existing scenario model calibration for the water quality parameters of concern. Results at select water quality stations are presented in the Current Condition scenario, Section 7.2.1.

7.2 Scenarios

Two modeling scenarios were developed and evaluated in this TMDL determination: a current condition and a natural condition scenario. Concentrations and loadings were evaluated to determine if DO concentrations in the natural condition scenario could meet the DO standard, and the impact of nutrients on the DO concentrations. The results from the scenarios were used to develop the TMDL.

7.2.1 Current Condition

The current condition scenario evaluated current hydrologic and water quality conditions in the watershed, specifically water quality concentration and loadings at the outlet of WBID 1778. The current condition annual average concentrations for the Cockroach Bay WBID are presented in Table 7.1. The current condition simulation was used to determine the base loadings for the WBID. These base loadings (Table 7.2), when compared with the TMDL scenarios, were used to determine the percent reduction in nutrient loads that will be needed to achieve water quality standards. Figure 7.10 through Figure 7.15 provide the calibrated current condition modeled parameters for WBID 1778.

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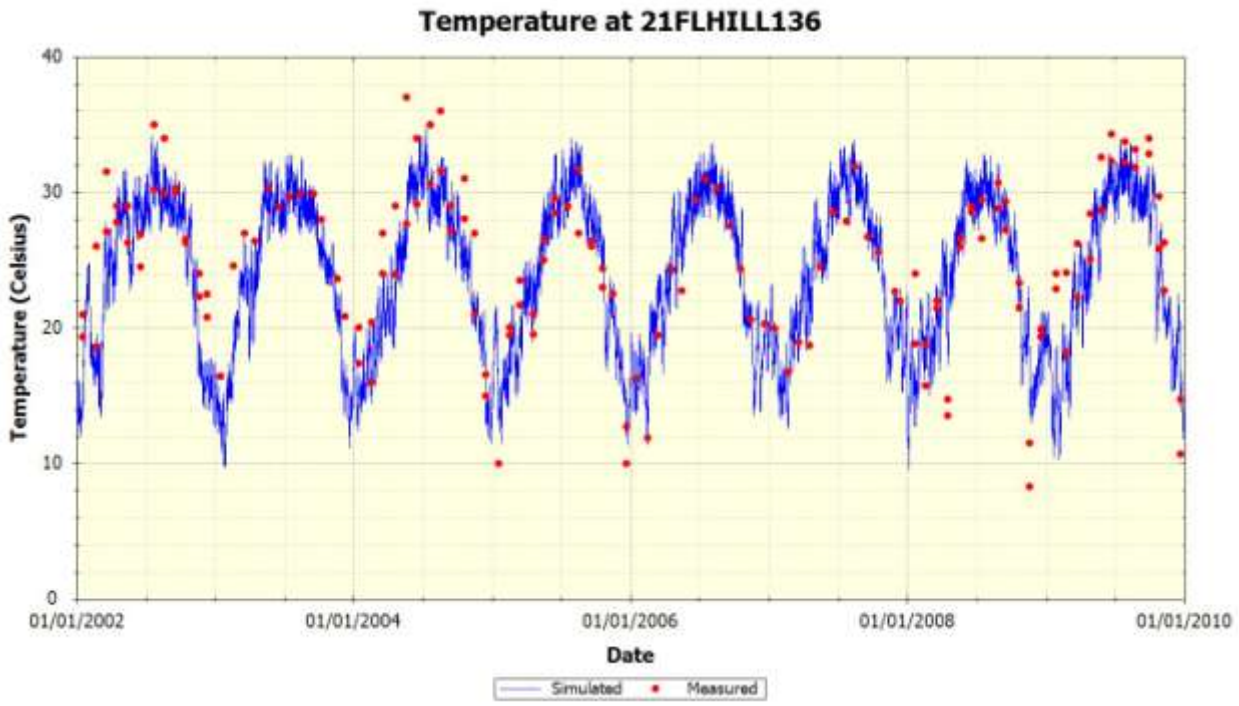

Figure 7.10 Simulated temperature verse measured temperature in Cockroach Bay at station 21FHILL136

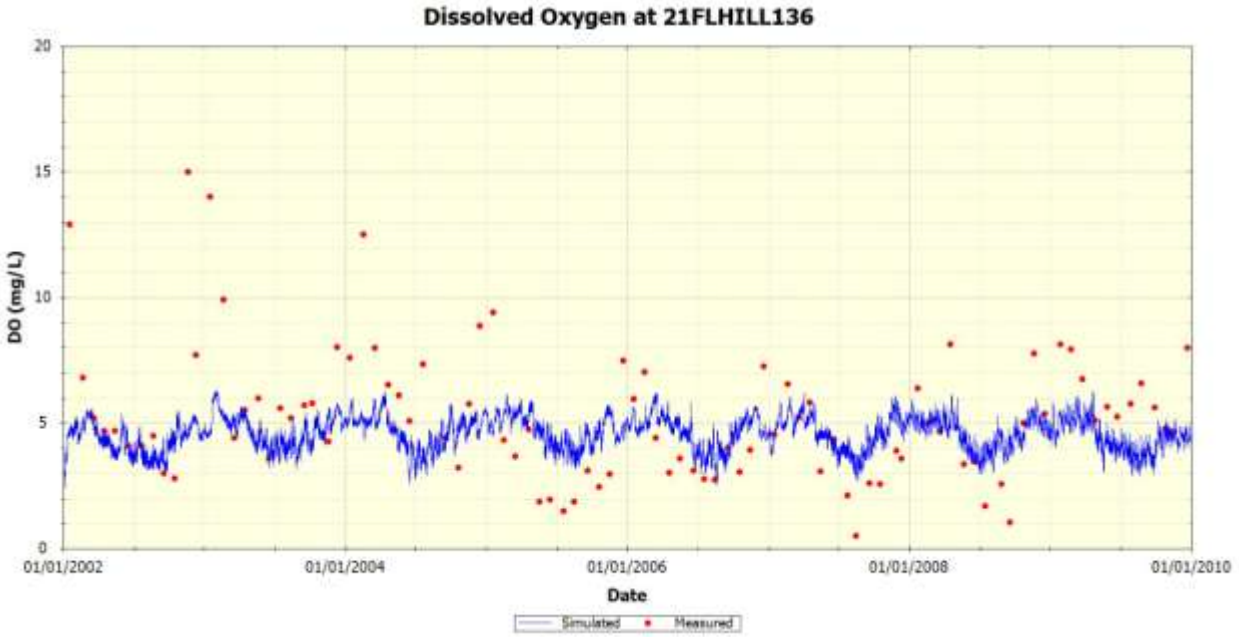

Figure 7.11 Simulated dissolved oxygen verse measured dissolved oxygen in Cockroach Bay at station 21FHILL136

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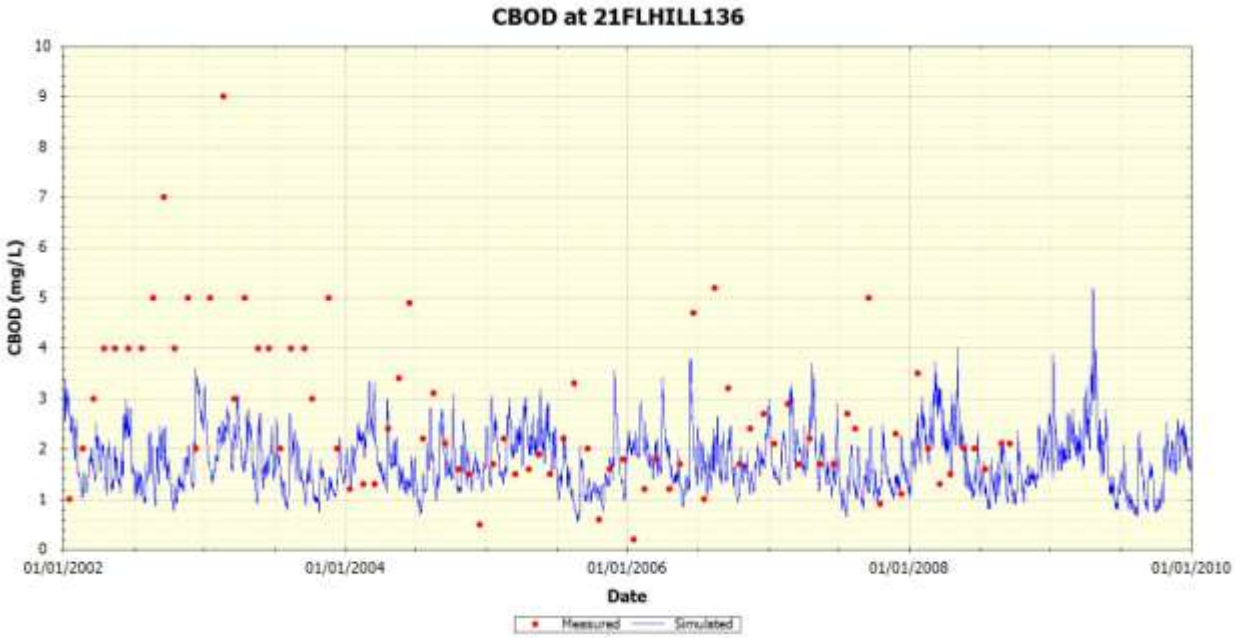

Figure 7.12 Simulated biochemical oxygen demand verse measured biochemical oxygen demand in Cockroach Bay at station 21FHILL136

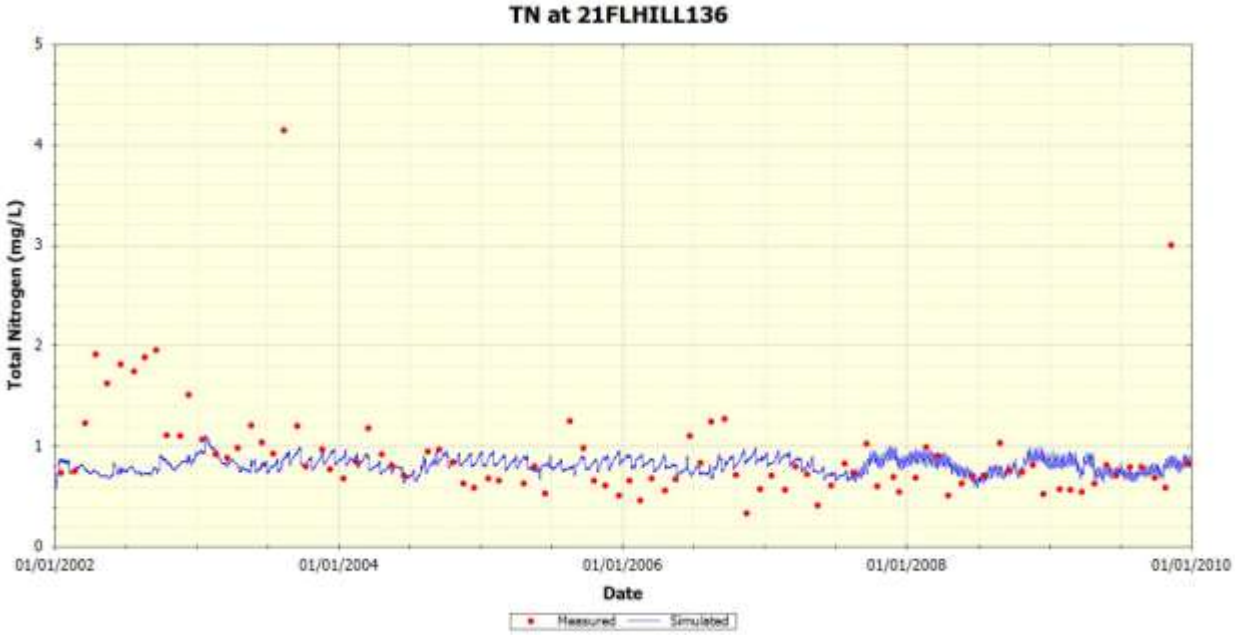

Figure 7.13 Simulated total nitrogen verse measured total nitrogen in Cockroach Bay at station 21FHILL136

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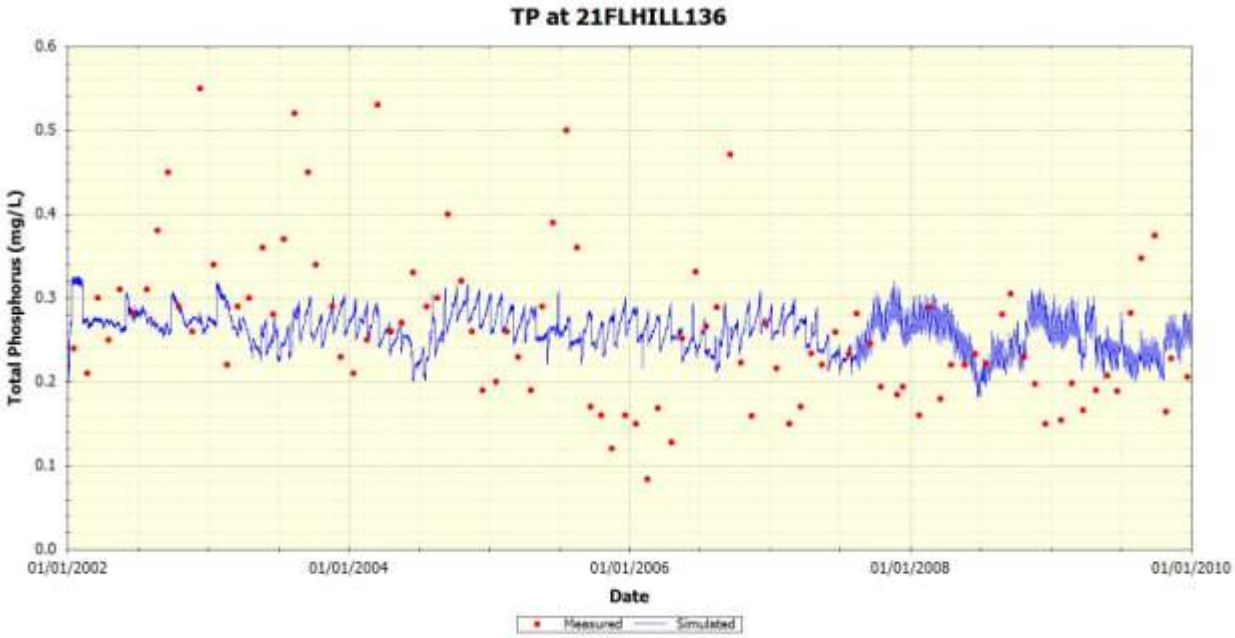

Figure 7.14 Simulated total phosphorus verse measured total phosphorus in Cockroach Bay at station 21FHILL136

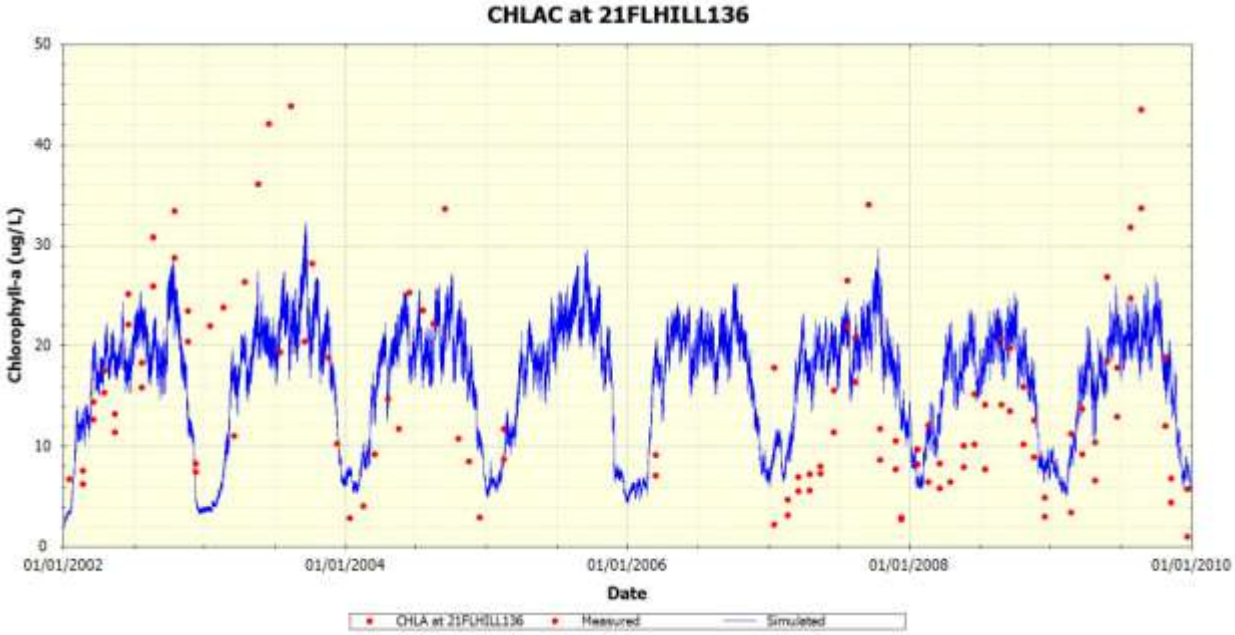

Figure 7.15 Simulated chlorophyll a verse measured chlorophyll a in Cockroach Bay at station 21FHILL136

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Table 7.1 Current condition concentrations in the impaired WBID in Cockroach Bay

Parameter	WBID 1778
Total nitrogen (mg/L)	0.81
Total phosphorus (mg/L)	0.26
cBOD (mg/L)	2.03
DO (mg/L)	4.72

Table 7.2 Current condition loadings in the impaired WBID in Cockroach Bay

Parameter	WBID 1778	
	WLA (kg/yr)	LA (kg/yr)
Total nitrogen (mg/L)	--	6,651
Total phosphorus (mg/L)	--	1,723
BOD (mg/L)	--	9,261

7.2.2 Natural Condition

The natural condition scenario was developed to estimate water quality conditions if there was no impact from anthropogenic sources. The point sources located in the model were removed for the natural condition analysis. Land uses that were associated with anthropogenic activities (urban, agriculture, transportation, barren lands and rangeland) were converted to upland forests or forested wetlands based on the current ratio of forest and wetland land uses in the model. Additionally, following the initial natural condition scenario run, sediment oxygen demand (SOD) was revised by using the following formula: $SOD_{revised}$= (Avg $Chla_{natural}$ / Avg $Chla_{existing}$) * SOD. The lower, revised SOD represents the change expected in SOD following excessive nutrient removal from the system. The natural condition water quality predictions are presented in Table 7.3 and Table 7.4.

The purpose of the natural conditions scenario was to determine whether water quality standards could be achieved without abating the naturally occurring loads from the watershed. The natural condition modeling scenario indicated that the DO standard is not achievable under natural conditions at all times, indicating that low DO is a naturally occurring phenomenon in WBID 1778. Figure 7.16 through Figure 7.20 provide the natural condition scenario modeled parameters for WBID 1778, and Figure 7.21 provides the cumulative distribution function of DO concentrations for both the modeled existing condition and natural condition results. The cumulative distribution curve shows there is an increase in DO concentrations in the natural condition scenario, specifically in DO concentration values less than 5 mg/L in the existing condition run. However, the DO concentration in the natural condition run is less than 5 mg/L approximately 1 percent of the time.

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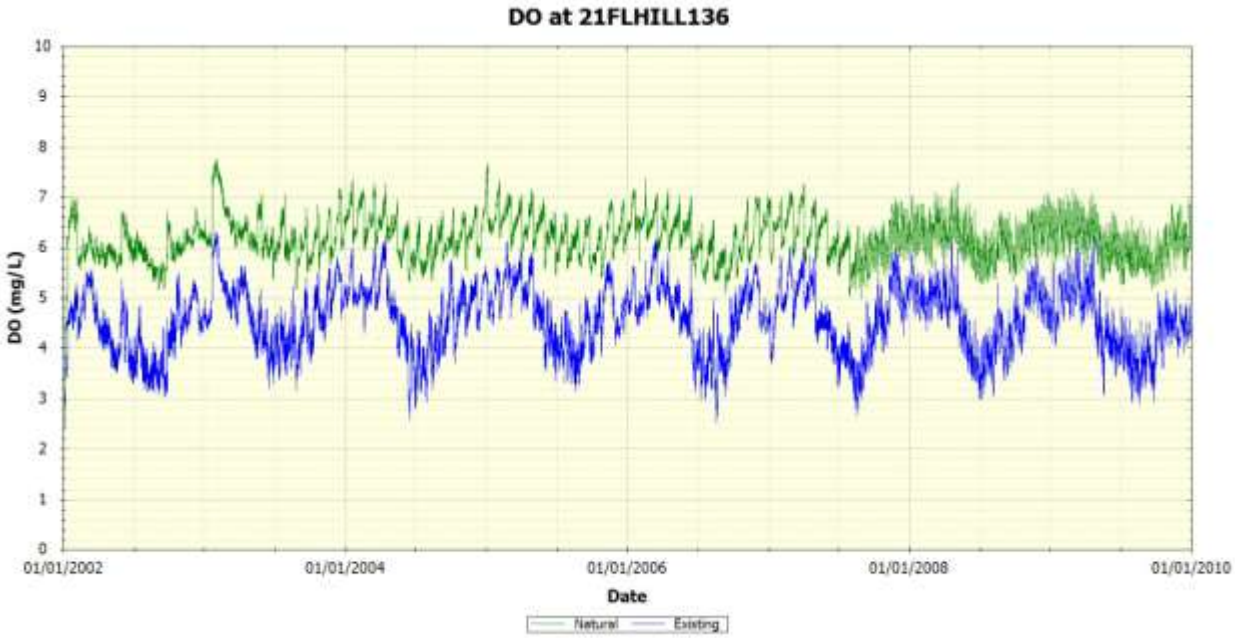

Figure 7.16 Natural condition dissolved oxygen in Cockroach Bay at station 21FHILL136

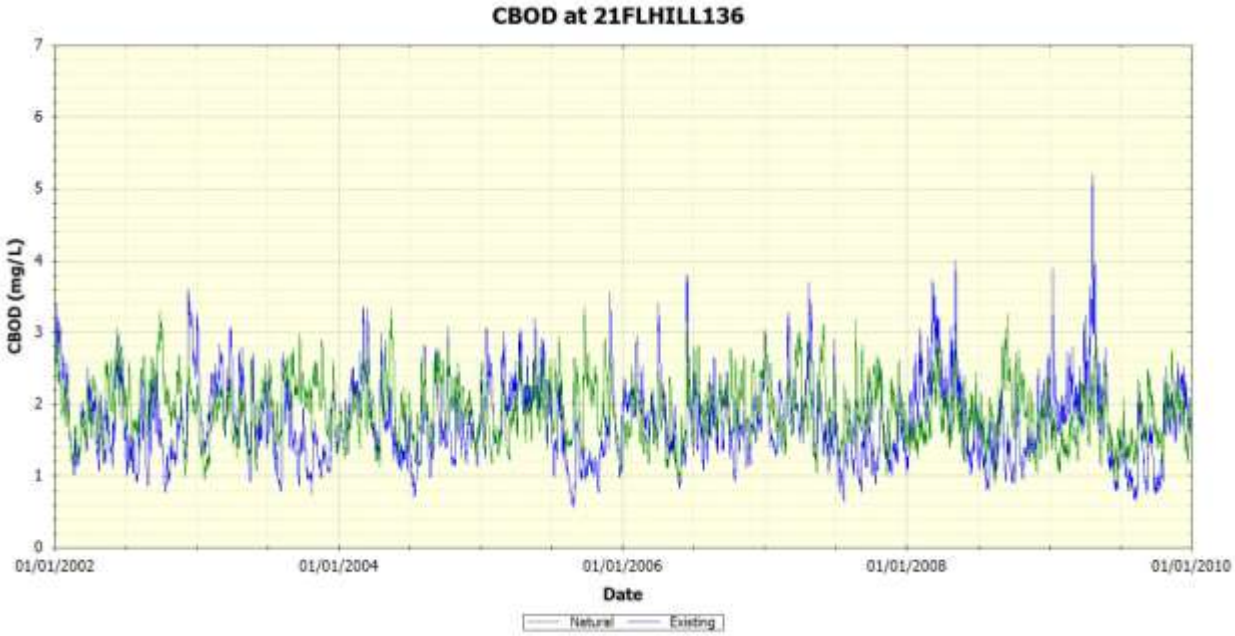

Figure 7.17 Natural condition biochemical oxygen demand in Cockroach Bay at station 21FHILL136

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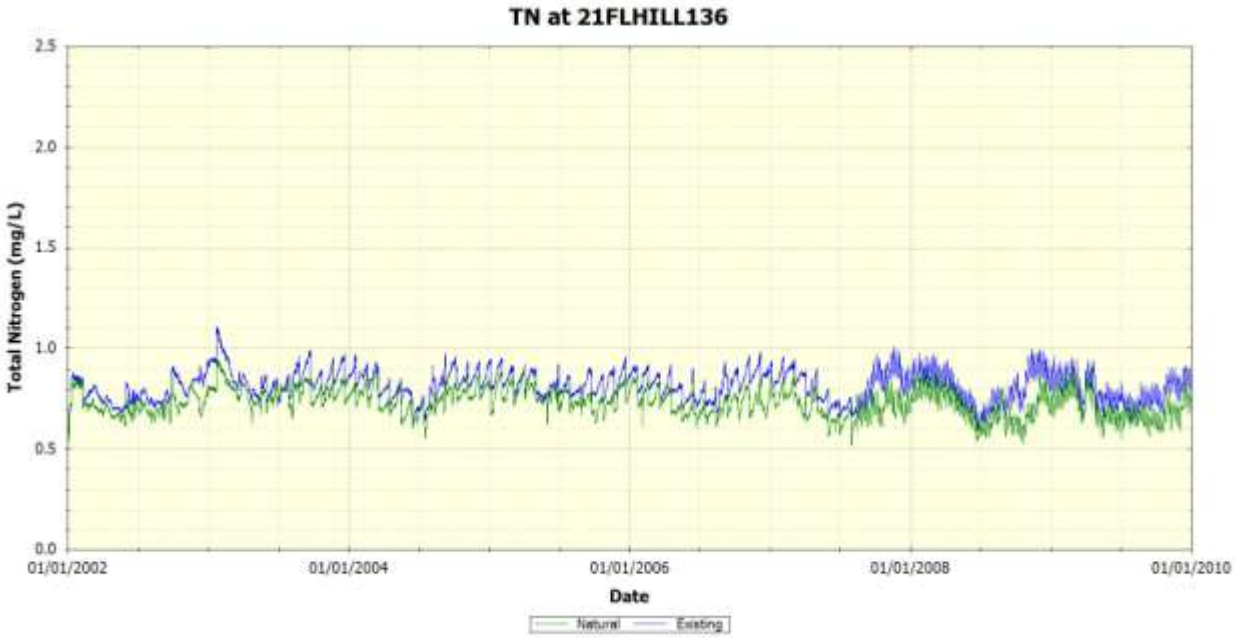

Figure 7.18 Natural condition total nitrogen in Cockroach Bay at station 21FHILL136

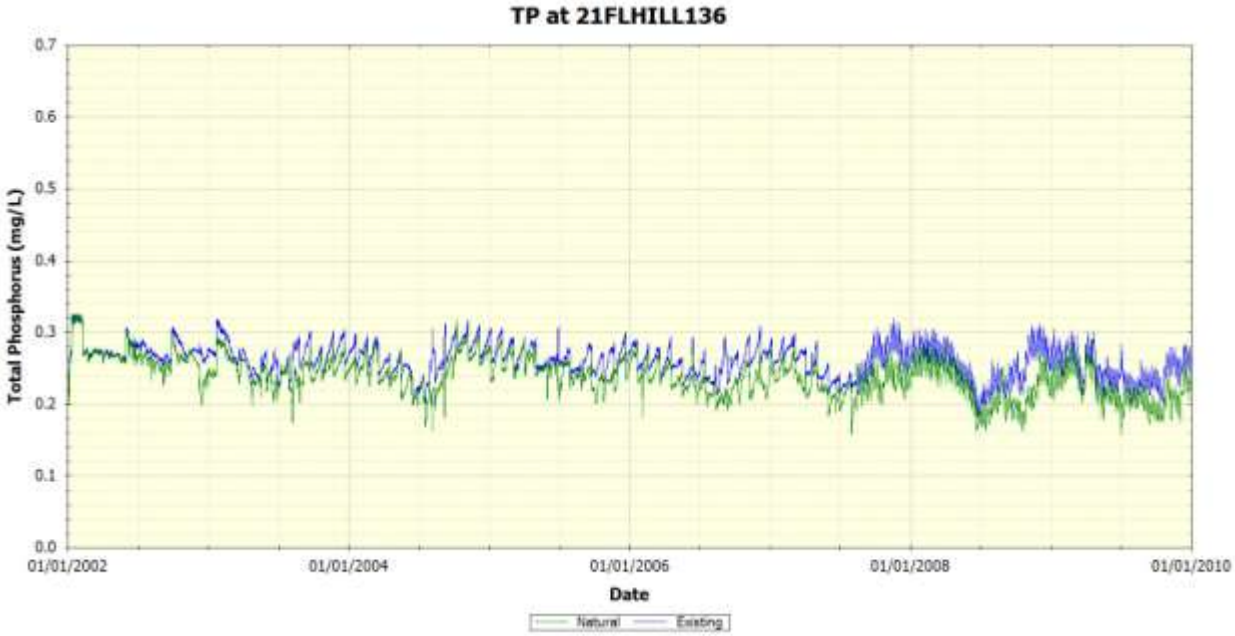

Figure 7.19 Natural condition total phosphorus in Cockroach Bay at station 21FHILL136

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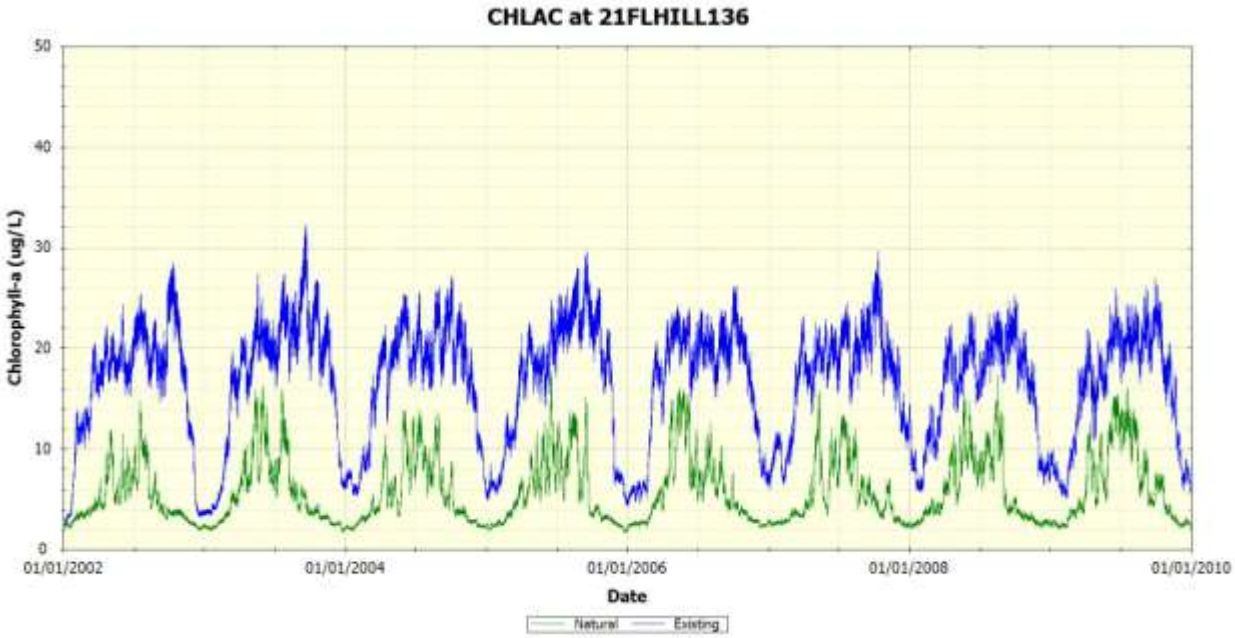

Figure 7.20 Natural condition chlorophyll a in Cockroach Bay at station 21FHILL136

Table 7.3 Natural condition concentrations in the impaired WBID in Cockroach Bay

Parameter	WBID 1778
Total nitrogen (mg/L)	0.80
Total phosphorus (mg/L)	0.24
cBOD (mg/L)	1.99
DO (mg/L)	6.23

Table 7.4 Natural condition loadings in the impaired WBID in Cockroach Bay

Parameter	WBID 1778	
	WLA (kg/yr)	LA (kg/yr)
Total nitrogen (mg/L)	--	2,220
Total phosphorus (mg/L)	--	90
BOD (mg/L)	--	5,727

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Figure 7.21 Dissolved oxygen concentration cumulative distribution function in Cockroach Bay at station 21FHILL136

8.0 TMDL DETERMINATION

The TMDL for a given pollutant and waterbody is comprised of the sum of individual wasteload allocations (WLAs) for point sources, and load allocations (LAs) for both nonpoint sources and natural background levels. In addition, the TMDL must include a margin of safety (MOS), either implicitly or explicitly, to account for the uncertainty in the relationship between pollutant loads and the quality of the receiving waterbody. Conceptually, this definition is represented by the equation:

$$\text{TMDL} = \sum \text{WLAs} + \sum \text{LAs} + \text{MOS}$$

The TMDL is the total amount of pollutant that can be assimilated by the receiving waterbody and still achieve water quality standards and the waterbody's designated use. In this TMDL development, allowable concentrations from all pollutant sources that cumulatively amount to no more than the TMDL must be set and thereby provide the basis to establish water quality-based controls. These TMDLs are expressed as annual geometric mean concentrations, since the approach used to determine the TMDL targets relied on geometric means. The TMDLs targets were determined to be the conditions needed to restore and maintain a balanced aquatic system. Furthermore, it is important to consider nutrient loading over time, since nutrients can accumulate in waterbodies.

The TMDL was determined for the concentrations and loadings at the outlet of WBID 1778, and included all loadings from upstream sources and streams. During the development of this TMDL, it was determined that the natural condition scenario (removal of all anthropogenic sources and land uses) met the DO standard the majority of the model run time from 2002 through 2009. Additionally, there was a reduction in chlorophyll a in the bayou in the natural condition run because of the reduction in nutrients. EPA believes that setting the TMDL condition to a natural condition protects both 47(a) and 47(b) of the Florida narrative nutrient

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standard. It assures that no man induced activities would have caused an imbalance of flora and fauna. Natural background levels are presumed to protect aquatic life. Florida's water quality standards do not allow the abatement of natural conditions, this TMDL represents the lowest level of nutrients that could be established.

The nutrient loadings from the natural condition scenario were therefore used to determine the TMDL in accordance with the Natural Conditions narrative rule. The allocations for WBID 1778 for total nitrogen, total phosphorus, and biochemical oxygen demand are presented in Table 8.1.

Table 8.1 TMDL Load Allocations for Cockroach Bay, WBID 1778

Constituent	Current Condition		TMDL Condition		Percent Reduction		
	WLA (kg/yr)	LA (kg/yr)	WLA (kg/yr)	LA (kg/yr)	WLA	LA	MS4
Total Nitrogen	--	6,651	--	2,220	--	67%	67%
Total Phosphorus	--	1,723	--	90	--	95%	95%
Biochemical Oxygen Demand	--	9,261	--	5,727	--	38%	38%

8.1 Critical Conditions and Seasonal Variation

EPA regulations at 40 CFR 130.7(c)(1) require TMDLs to take into account critical conditions for stream flow, loading, and water quality parameters. The critical condition is the combination of environmental factors creating the "worst case" scenario of water quality conditions in the waterbody. By achieving the water quality standards at critical conditions, it is expected that water quality standards should be achieved during all other times. Seasonal variation must also be considered to ensure that water quality standards will be met during all seasons of the year, and that the TMDLs account for any seasonal change in flow or pollutant discharges, and any applicable water quality criteria or designated uses (such as swimming) that are expressed on a seasonal basis.

The critical condition for nonpoint source concentration and wet weather point source concentrations is typically an extended dry period followed by a rainfall runoff event. During the dry weather period, nutrients build up on the land surface, and are washed off by rainfall. The critical condition for continuous point source concentrations typically occurs during periods of low stream flow when dilution is minimized. Although loading of nonpoint source pollutants contributing to a nutrient impairment may occur during a runoff event, the expression of that nutrient impairment is more likely to occur during warmer months, and at times when the waterbody is poorly flushed.

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8.2 Margin of Safety

The Margin of Safety accounts for uncertainty in the relationship between a pollutant load and the resultant condition of the waterbody. There are two methods for incorporating an MOS into TMDLs (USEPA 1991):

- Implicitly incorporate the MOS using conservative model assumptions to develop allocations
- Explicitly specify a portion of the total TMDL as the MOS and use the remainder for Allocations

This TMDL uses an implicit MOS since the TMDL targets for nutrients were set to natural background conditions.

8.3 Waste Load Allocations

Only MS4s and NPDES facilities discharging directly into lake segments (or upstream tributaries of those segments) are assigned a WLA. The WLAs, if applicable, are expressed separately for continuous discharge facilities (e.g., WWTPs) and MS4 areas, as the former discharges during all weather conditions whereas the later discharges in response to storm events.

8.3.1 Wastewater/Industrial Permitted Facilities

A TMDL wasteload allocation (WLA) is given to wastewater and industrial NPDES-permitted facilities discharging to surface waters within an impaired watershed. There are no continuous discharge NPDES-permitted point sources discharging to surface waters in WBID 1778, therefore no WLA was calculated.

8.3.2 Municipal Separate Storm Sewer System Permits

The WLA for MS4s are expressed in terms of percent reductions equivalent to the reductions required for nonpoint sources. Given the available data, it is not possible to estimate concentrations coming exclusively from the MS4 areas. Although the aggregate concentration allocations for stormwater discharges are expressed in numeric form, i.e., percent reduction, based on the information available today, it is infeasible to calculate numeric WLAs for individual stormwater outfalls because discharges from these sources can be highly intermittent, are usually characterized by very high flows occurring over relatively short time intervals, and carry a variety of pollutants whose nature and extent varies according to geography and local land use. For example, municipal sources such as those covered by this TMDL often include numerous individual outfalls spread over large areas. Water quality impacts, in turn, also depend on a wide range of factors, including the magnitude and duration of rainfall events, the time period between events, soil conditions, fraction of land that is impervious to rainfall, other land use activities, and the ratio of stormwater discharge to receiving water flow.

This TMDL assumes for the reasons stated above that it is infeasible to calculate numeric water quality-based effluent limitations for stormwater discharges. Therefore, in the absence of information presented to the permitting authority showing otherwise, this TMDL assumes that water quality-based effluent limitations for stormwater sources of nutrients derived from this TMDL can be expressed in narrative form (e.g., as best management practices), provided that: (1) the permitting authority explains in the permit fact sheet the reasons it expects the chosen BMPs to achieve the aggregate wasteload allocation for these stormwater discharges; and (2) the

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state will perform ambient water quality monitoring for nutrients for the purpose of determining whether the BMPs in fact are achieving such aggregate wasteload allocation.

All Phase 1 MS4 permits issued in Florida include a re-opener clause allowing permit revisions for implementing TMDLs once they are formally adopted by rule. Florida may designate an area as a regulated Phase II MS4 in accordance with Rule 62-620.800, FAC. Florida's Phase II MS4 Generic Permit has a "self-implementing" provision that requires MS4 permittees to update their stormwater management program as needed to meet their TMDL allocations once those TMDLs are adopted. Permitted MS4s will be responsible for reducing only the loads associated with stormwater outfalls which it owns, manages, or otherwise has responsible control. MS4s are not responsible for reducing other nonpoint source loads within its jurisdiction. All future MS4s permitted in the area are automatically prescribed a WLA equivalent to the percent reduction assigned to the LA. The MS4 service areas described in Section 6.1.2 of this report are required to meet the percent reduction prescribed in Table 8.1 through the implementation of BMPs.

8.4 Load Allocations

The load allocation for nonpoint sources was assigned a percent reduction in nutrient concentrations from the current concentrations coming into the WBID addressed in the TMDL report.

9.0 RECOMMENDATIONS/IMPLEMENTATION

The initial step in implementing a TMDL is to more specifically locate pollutant source(s) in the watershed. FDEP employs the Basin Management Action Plan (B-MAP) as the mechanism for developing strategies to accomplish the specified load reductions. Components of a B-MAP are:

- Allocations among stakeholders
- Listing of specific activities to achieve reductions
- Project initiation and completion timeliness
- Identification of funding opportunities
- Agreements
- Local ordinances
- Local water quality standards and permits
- Follow-up monitoring

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