US ERA ARCHIVE DOCUMENT

**SEPA** 

# Research and Development

HEALTH AND ENVIRONMENTAL EFFECTS PROFILE FOR CARBARYL

Carurell # 160

# Prepared for

OFFICE OF SOLID WASTE AND EMERGENCY RESPONSE

# Prepared by

Environmental Criteria and Assessment Office Cincinnati OH 45268

DRAFT: DO NOT CITE OR QUOTE

#### NOTICE

This document is a preliminary draft. It has not been formally released by the U.S. Environmental Protection Agency and should not at this stage be construed to represent Agency policy. It is being circulated for comments on its technical accuracy and policy implications.

#### DISCLAIMER

This report is an external draft for review purposes only and does not constitute Agency policy. Mention of trade names or commercial products does not constitute endorsement or recommendation for use.

ŗ

## TABLE OF CONTENTS

																											-	<u> Page</u>
1.	INTRODU	UCTION.		.• ,		٠	•					٠	.•		•			•	•	.• .			•	•	, . <b>*</b>			1
•																									•	*.		1
	1.1.	STRUCTI	URE A	IND	CA	S	NUI	JRF	:K	•-	:	•	:	•		•	•	٠	•	٠	•	.•	•	•	٠	•	,	ή
	1.2.	PHYSIC	AL AN	ID I	CHE	MI	CAL	_ F	RU	PE	RI	It	S		•	٠	•	•		•			٠	•	٠	•	•	2
	1.3.	PRODUC	TION	DA	TA.		•	•	•	٠	•	, <b>•</b>	٠	•	•	•	٠	•	•	•	•	٠	٠	•	٠	•	•	2
Ł	1.4.	USE DA	TA.	•		•	٠	•		٠	٠	٠	•	٠	•	٠	•	•	•	٠	٠	٠	•	٠	.•		•	2
2.	ENVIRO	NMENTAL	FATE	E A	ND	TR	AN	SPO	ORT	P	RC	CE	SS	SES	<b>S</b> .	•		•	٠	.•	٠		•	.*	•	,	•	3
	2.1.	AIR			. ,								,•		•	•	•		•		•	•	•	•		•	•	3
	2.2.	WATER			_		_			_									٠		•	٠	•	•		•		3
	2.3.	SOIL .		•	•		•	•		٠	•	٠	٠	•	٠	٠	•	•	•		,•	•	٠	•	•	•	•	8
3.	EXPOSU	RE						•		•	٠	•	•	•	.•	<b>.</b>	•		•		•	•	•	•	, ,	•	•	13
	0 1	WATER.															_	_								•	•	13
	3.1.	FOOD .	• •	•	•	• •	•	•	•	•	•	•	•	•	Ī	Ī	•	-	-	•								13
	3.2.	INHALA	TION	•	•	• •	•	•	•	•	٠	•	•	•	•	•	•	•	Ĭ	•		_						13
	3.3.			•	•	•	•	•	•	•	٠	•	٠	•	•	•	•	•	•	Ī	•	-						13
	3.4.	DERMAL																										
4.	PHARMA	COKINET	TICS.	•	٠	• /		•	•		•	.•	•	•	•	•	•	•	•	٠	•	•		•	•	•	•	15
	4.1.	ABSORF	PTION	١.									•							٠,			•	•	•			15
	4.2.	DICTOI	RUTT	NN						_							٠.							•	•	•	•	15
	4.3.	METAR	OL TSM	١.	_	_												٠					•	•	•	٠	÷	17
	4.4.	EXCRE	TION.			•	• .			•	•	•	•			•	•	•	•	•		•	•	•	•	•	•	21
5.	EFFEC"	TS			•	•				•	,.	•	•		•			•	•	•	, i	• ,	•	•	•	•	•	23
	5.1.	CARCI	NUCER	J T C	TTV	,					_		_									•		•			•	23
	5.1. 5.2.	MUTAG			111	•	•	•	•																			26
	5.3.	TERAT	OCEN!	r r r	τv	•	•	•	•											, ,						,•	•	37
		OTHER	DEDI		1101	riv	F	FFI	FFC	T																		45
	5.4.	CHRON	TC AL	NOD NO	CHI	S C N	DU	MT	ר ז	ΓNΥ		T	TY.	•	•							_						48
	5.5.	OTHER	TO WI	NU CVA	NT	TM	IE N	DM	ATI	i OA	1	, .	•		•	•						_				٠		51
	5.6.	UIHEK	KELI	E V M	IN I	Ιπ	ir U	KM	n i i	LOI	•	•	•	• ,	•	•	•	•			•	-	•		_			
6.	AQUAT	IC TOXI	CITY	• , •	•	•	•	•	• 1	• •	•	•	•	•	•	•	•	•	•	•	•	•	•	•	٠	•		
	6.1.	ACUTE				•			•			•		•	•		•	•	• .	•	•	•	•	٠			٠	54
	6.2.	CHRON	ITC.			_						•	٠		•	•	•		•	•	•		•	٠	.•	٠		20
	6.3.	PI ANT	FFF	<b>ECT</b>	S.									•		•	•	•	•	•	•	•	٠	٠	٠	٠	•	DU
	6.4.	DECTO	ME										_	_	_	_	4					•			•			וסו
	6.5.	OTHER	REL	EV#	INT	I	NF O	RM	AT	I 0	N		•	•	•	•	•	•	•	•	٠	•	•	•	•	•	•	61
7.	EXIST	ING GUI	DELI	NES	S A	ND	ST	AN	DA	RD	S	•	•	•	•	•	•	•	•	•	•	•	<b>.</b>	٠	•	•	. •	63
	7 1	HUMAN	d.								_											•			•			63
	7.1. 7.2.		TTC .	•	• •	•	•	•		-		-		•										•				63
	1.2.	NUUN	110.		• •		•	•	•		•	•		-	-	-	-	-										

# TABLE OF CONTENTS

																										<u>Page</u>
8.	RISK ASSESSMENT	•	•			.•		•	•	.•		٠	•	•	•	•	•		٠,	•		•	•	•	•	64
9.	REFERENCES	•			•		٠	•		•	•	•	•	٠	٠.		•	•		•	•	•	•	•	•	66
APPE	NDIX: LITERATURE	SE	AR	CHE	D.	•	٠	•	٠	•	•		•	•	.•	•	•	٠	٠	•	•	•	÷	•	.•	94

# LIST OF TABLES

No.	<u>Title</u>	Page
4-1	Carbaryl Metabolites Detected in Rats, Guinea Pigs, Monkeys, Swine, Sheep, Dogs and Humans	. 19
5-1	Summary of Mutagenicity Test Results for Carbaryl	. 27
5-,2	Teratogenicity Testing of Carbaryl	. 38
5-3	Acute Lethal Toxicity of Carbaryl	. 52
6-1	Acute Lethal Effects of Carbaryl to Freshwater Fish in a 96-Hour Bioassay	. 55
6-2	Acute Lethal Effects of Carbaryl to Marine Fish in a Static Bioassay	. 57
6-3	Acute Lethal Effects of Carbaryl to Aquatic Invertebrates	. 59

## LIST OF ABBREVIATIONS

Acceptable daily intake ADI Bioconcentration factor BCF Body weight bw Deoxyribonucleic acid DNA Gastrointestinal GI **Intraperitoneal** 1.p. Concentration lethal to 50% of recipients LC<sub>50</sub> Dose lethal to 50% of recipients LD<sub>50</sub> Lowest lethal dose LDLo Maximum tolerable dose MTD Parts per million ppm Subcutaneous s.c. Sister-chromatid-exchange SCE Short-term exposure limit STEL Threshold limit value TLV Time-weighted average TWA Ultraviolet UV

#### 1. INTRODUCTION

#### 1.1 STRUCTURE AND CAS NUMBER

Carbaryl is the common name for 1-naphthyl N-methylcarbamate. It is sold in the United States under the trade name Sevin. The Chemical Abstracts Service (CAS) Registry Number for carbaryl is 63-25-2, and the structure is given below.

Molecular formula:  $C_{12}H_{11}N_{2}$ 

Molecular weight: 201.2

There are numerous additional synonyms and trade names for carbaryl (IARC, 1976; NIOSH, 1976), including 1-naphthol N-methylcarbamate; 1-naphthyl methylcarbamate; methylcarbamate 1-naphthalenol; methylcarbamic acid, 1-naphthyl ester; N-methyl-1-naphthyl carbamate; N-methyl-\alpha-naphthyl-urethan; Atoxan; Caprolin; Carpolin; Compound 7744; Dicarbam; Sevidol and Union Carbide 7744.

#### 1.2. PHYSICAL AND CHEMICAL PROPERTIES

Carbaryl is a white crystalline solid that is formulated as wettable powders, pellets, granules, dusts, suspensions and emulsifiable concentrate solutions (IARC, 1976). Important physical and chemical properties are summarized below (IARC, 1976; NIOSH, 1976):

Melting point:

142-145°C

Specific gravity:

1.232 at 20°C

Vapor pressure:

0485p

0.000041 mm Hg at 25°C 0.00015 mm Hg at 40°C

Henry's Law Constant:

13.1 (unitless) (U.S. EPA, 1981a)

Solubility in water: 40 ppm at 30°C

31 ppm at 8°C in seawater (Karinen et al., 1967)

Solubility in organic

solvents:

soluble in acetone, cyclohexanone, and

dimethylformamide

Log octanol/water partition coeffi-

2.36 (Kenaga and Goring, 1980; Karickhoff, 1981)

#### 1.3. PRODUCTION DATA

Carbaryl is prepared by the reaction of 1-naphthol and methyl isocyanate, or of 1-naphthol, phosgene and methylamine (Martin and Worthington, 1977).

Carbaryl is manufactured by Union Carbide, Institute, West Virginia (SRI, 1983). The estimated production for 1972 was 53 million pounds/year. Domestic consumption was ~25 million pounds in 1972, including 19 million pounds for agricultural purposes alone (von Rumker et al., 1974).

#### 1.4. USE DATA

Carbaryl is a widely used carbamate insecticide with a broad spectrum of effectiveness on insects. It is used primarily on corn, soy beans, cotton, fruit and nut crops, and vegetable crops (von Rumker et al., 1974), and is approved for home yard and garden use.

# 2. ENVIRONMENTAL FATE AND TRANSPORT PROCESSES

#### 2.1. AIR

Specific information regarding the fate and transport of carbaryl in the atmosphere was not located in the available literature. Data for photode-gradation in water (Section 2.2.) indicate that carbaryl will decompose in air through direct photolysis or UV light accelerated hydrolysis. Carbaryl is likely to enter the atmosphere as a result of spraying and dusting operations, but evaporation from water or soil is not likely to occur to a significant extent (Sections 2.2. and 2.3.). Relatively low soil sorption constants ( $K_{\rm oc}$ ) in the range of 200-400 and other data discussed in Sections 2.2. and 2.3. suggest that carbaryl may exist in the atmosphere to a limited extent in the particulate sorbed phase.

#### 2.2. WATER

Numerous studies have demonstrated that carbaryl does not persist in natural water when studied in the laboratory (Karinen et al., 1967; Eichelberger and Lichtenberg, 1971; Kanazawa, 1975; Rodriguez and Dorough, 1977; Freitag et al., 1979; Szeto et al., 1979; Sharom et al., 1980; Odeyemi, When tested at concentrations of <10 ppm at pH 6.5-8, complete 1982). disappearance generally occurred within a week at room temperature; at 9°C, Szeto et al. (1979) found 61% loss after 50 days and 80% loss after 42 days in creek and pond water, respectively. Carbaryl is degraded in water by and biological processes, and persistence is significantly increased in the presence of sediments or if pH is lowered. In a field study, Stanley and Trial (1980) determined the rate of disappearance of carbaryl from streams that received drift from spraying of nearby forests. An average disappearance constant of 0.028 hour was determined from measurements from nine Maine brooks and rivers (12 sites), and the constant

was not influenced by the size of the stream or the initial concentration of carbaryl in the water (2-16  $\mu g/R$ ). The corresponding half-life is 24.75 hours, a value that is comparable to that reported for natural water in the laboratory. It should be noted, however, that the analytical method used in this study measured both the parent compound and its major chemical degradation product, 1-naphthol in the Maine water (pH ~5.5-7.0).

Carbaryl is degraded in water predominantly by hydrolysis, yielding 1-naphthol, methylamine and CO<sub>2</sub> (Aly and E1-Dib, 1971; Wolfe et al., 1978). Kinetic studies have shown that carbaryl is stable to hydrolysis in acidic pH, and that degradation is rapid in basic pH (Aly and El-Dib, 1971; Wauchope and Haque, 1973; Wolfe et al., 1976); hydrolysis half-lives at pH values normally found in the aquatic environment at 27°C were calculated from measured rate constants to be 3.6 years (pH 5), 4.4 months (pH 6), 13 days (pH 7), 1.3 days (pH 8) and 3.2 hours (pH 9) (Wolfe et al., 1976). The hydrolysis of carbaryl is also significantly affected by temperature. Aly and El-Dib (1971) found that the rate of hydrolysis at pH 8.0 (i.e., second order rate constant) increased 2.9 times with a temperature increase from 13-23°C. In experiments conducted with seawater at pH 7.8, the amount of carbaryl (10 ppm) hydrolyzed after 4 days at 3.5, 17, 20 and 28°C was 0% (not detected), 44, 55 and 93%, respectively (Karinen et al., 1967). These data (Karinen et al., 1967) also suggest that salt content (ionic strength) may be an additional important factor affecting the rate of hydrolysis, since hydrolysis at comparable pH and temperature in freshwater is faster (Wolfe et al., 1976) than water with higher salt content.

Limited data are available on the photolysis of carbaryl in water. Wolfe et al. (1976) calculated half-lives of 51 and 64 hours (spring), 46 and 52 hours (summer), 68 and 102 hours (fall) and 103 and 200 hours

(winter) for latitudes of 30°N and 40°N, respectively, for the direct photolysis of carbaryl near the surface (<10 cm) of distilled water. These half-lives were calculated from disappearance quantum yields determined at 20°C and 313 nm. and data further indicated that photolysis is slowed by the presence of oxygen (i.e., under air-saturated reaction conditions) and is pH-dependent in the pH 5-7 range. Experiments conducted under midday summer sunlight (June, latitude 34°N) yielded data that were consistent with the calculated half-lives; the half-life was found to be ~45 hours in distilled water buffered at pH 5.5, and dark controls showed no decomposition. Wolfe et al. (1976) found that 1-naphthol and methyl isocyanate were not products of the direct photolysis of carbaryl at wavelengths >290 nm in either degassed or air-saturated water (pH 5.5), but did not identify the photoproducts. Aly and El-Dib (1972) reported that irradiation of aqueous carbaryl (20 ppm) with 254 nm light for 1 hour resulted in 50, 57 and 78% decomposition at pH 5.0. 7.0 and 8.0, respectively, and that 1-naphthol appeared as a product after 5 minutes of exposure in all cases. Karinen et al. (1967) noted that fluorescent light seemed to have a slight acceleration effect on hydrolysis of carbaryl in seawater at 20°C; hydrolysis averaged ~63% in 4 days in the dark and 72% under fluorescent light.

The oxidation of carbaryl in water under environmental reaction conditions is not expected to be significant when compared with its chemical and photolytic reactivity (Wolfe et al., 1976). The hydrolysis product 1-naphthol should, however, be very readily oxidized. Wauchope and Haque (1973) have reported that 1-naphthol in basic solution (i.e., 1-naphthoxide ion), but not in weakly acidic solutions, photooxidizes to 2-hydroxy-1,4-naphtho- quinone in the presence of room light.

Carbaryl can be degraded by microorganisms in aquatic environments, but the rate of degradation appears to be significant only in waters where hydrolysis is limited (i.e., neutral to acidic pH). Biodegradation of carbaryl has been demonstrated in freshwater samples (Aly and El-Dib, 1972; Szeto et al., 1979; Odeyemi, 1982), in simulated aquatic environments (Liu et al., 1981) and in pure and mixed culture experiments with bacteria isolated from freshwater (Guthrie et al., 1981) and marine water (Sikka et al., 1975). 1-Naphthol, the hydrolysis product of carbaryl, is readily utilized by microorganisms (Karinen et al., 1967; Aly and El-Dib, 1972; Sikka et al., 1975; Bollag et al., 1975), indicating that the hydrolysis step may be the limiting factor for biodegradation in natural waters.

In a study with Nile river water of pH 7.2 (25°C), Aly and El-Dib (1972) found that the concentration of carbaryl decreased progressively with time and that 89% of the added amount (4.75 ppm) disappeared in 6 days. 1-Naphthol appeared as a degradation product (2.2 and 0.8 ppm were detected after 2 and 6 days, respectively), apparently resulting from biological oxidation; these quantities of 1-naphthol did not result from chemical hydrolysis, since a sterile solution showed negligible hydrolysis after 6 days and a half-life of 16 days. The concentration of 1-naphthol reportedly decreased with time, indicating the biodegradation of this compound. Subsequent additions of increasing concentrations of carbaryl disappeared in shorter periods of time with no high buildup of 1-naphthol, thereby indicating microbial acclimatization; at the fourth addition, 17.5 ppm of carbaryl essentially disappeared after 1 day and no 1-naphthol was detected. Respirometric studies with higher levels of compound showed that, following a lag period of one day, oxygen uptake with 40 ppm carbaryl reached 95% of the theoretical amount required to oxidize it completely in 5 days; at 82 ppm carbaryl, an initial low rate of oxidation for 6 days was followed by rapid oxygen uptake, with 74% of the theoretical oxygen was used in 10 days (Aly and El-Dib, 1972).

Studies conducted in a simulated aquatic environment (i.e., in modified cyclone fermentors with a lake sediment/silt loam/activated sludge inoculum) suggest that biodegradation may also contribute significantly to the fate of carbaryl in anaerobic environments or in the presence of sediments with a rich organic nutrient content (Liu et al., 1981). Under anaerobic conditions, the half-life for biodegradation was reported to be 11.6 days when carbaryl was used as the sole carbon source and 6.1 days in the presence of glucose and peptone; the respective half-lives under aerobic conditions were 54 and 7.6 days. These results suggest that while the metabolism process played an insignificant part in carbaryl degradation in an aerobic environment, the biodegradation rate was accelerated somewhat in an anaerobic situation.

Measured soil sorption coefficients  $(K_{\rm oc})$  of 200-400 (Section 2.3.), measured octanol/water partition coefficients  $(K_{\rm ow})$  of 230 (Hansch and Leo, 1979; Kenaga and Goring, 1980; Karickhoff, 1981), and the reasonably low water solubility (see Section 1.2.) suggest that some adsorption of undegraded carbaryl onto bottom mud or sediments will occur; this is confirmed by positive sediment monitoring data (Section 3.1.). Volatilization of 1 ppm carbaryl from natural water in laboratory flasks at 9°C was not observed, a result that is consistent with its Henry's Law Constant value of 13 (U.S. EPA, 1981a).

Szeto et al. (1979) reported that addition of 1 ppm carbaryl to pond water (pH 7.5-7.9) in the presence of bottom sediment resulted in approxi-

mate averages at 40 and 33% recovery of carbaryl from the water and sediment, respectively, after 2 days of incubation at 9°C in laboratory flasks. Initial recovery from the water at day 0 was 90%, and recovery after 7, 14, 21 and 42 days from the water and sediment was approximately 26, 19, 11 and 14%, and 30, 28, 29 and 14% for water and bottom sediments, respectively.

In another study, the carbaryl concentration in seawater (10 ppm, pH 7-8, 8°C) in a laboratory aquarium without mud decreased ~50% in 38 days (Karinen et al., 1967). Most of this decrease was accounted for by the production of 1-naphthol. When 5, 10 or 25 ppm carbaryl was added to similar aquaria that contained estuarine mud, there was a sharp decrease in carbaryl concentration after ~3 days; 1-naphthol production was slight during this period, indicating that absorption by the mud was the major reason for the decline. When mud was present, the concentration of carbaryl plus 1-naphthol decreased to ~10% of the peak (3-day) value within 10 days. The concentrations of 1-naphthol in the aquaria with mud remained at low levels during the course of the 38-day study, indicating a slower rate of decomposition in the mud. In a field study, a 25x25 ft. estuarine mud flat plot was treated with carbaryl (at a rate of 10 lb A.I./acre) mixed with seawater at low tide (Karinen et al., 1967). Analysis showed that carbaryl was detectable in the mud 42 days after application; in 42 days, the concentration of carbaryl decreased from 5.4 to 0.1 ppm in the top 1 inch, from 0.32 to 0.20 ppm at the 2-3 inch level, and was detected at a level of 0.08 ppm at the 4-6 inch level at the termination of the study. 1-Naphthol levels were reportedly low after the first day, which supports the laboratory findings that hydrolysis proceeds slowly in the mud.

#### 2.3. SOIL

Carbaryl does not persist in soil. In an extensive field study, Caro et al. (1974) applied carbaryl granules with corn seed in furrows 1 meter apart

in Coshocton silt loam (average pH 5.2) at a rate of 5 kg/ha, and persistence was measured by sampling in the corn rows at 10 locations on 7 occa- sions throughout the crop season. Regression analysis of the sampling data suggested that ~135 days were required for 95% of the carbaryl to disappear, but this is an overall value for the entire field. The disappearance did not conform to a first order reaction, and data for the individual sample points indicated that the carbaryl remained stable in the soil for 25 to more than 116 days and then degraded rapidly. Although the rate of degrada- tion varied with location, these lag periods indicated that carbaryl degra- dation in the soil was primarily microbiological.

The importance of soil organisms in the degradation of carbaryl was further illustrated by Rodriguez and Dorough (1977), who studied the persistence of carbaryl in Maury soil samples (pH not stated) that differed only in previous pesticide treatment. When 10 ppm carbaryl-naphthyl-l-14C was incubated at 27°C in laboratory flasks with soil that had received no recorded pesticide treatment or with soil that had received 4 pounds/acre of carbaryl 6 months prior to collection, ~10 and 72% of the radioactivity, respectively, was lost from the soils after 4 days. Loss of radioactivity continued at a faster, although steady, rate in the previously untreated soil, but little subsequent loss occurred in the previously treated soil; after 120 days, the amount of label lost from the two soils was ~80 and 86%, respectively. Most of the lost radioactivity was attributed to liberation of CO<sub>2</sub> from microbial degradation of the naphthalene ring, and almost all of the terminal residual radioactivity was unextractable from the soil with acetone. Confirmatory studies showed that the most rapid loss of label from the carbaryl-pretreated soil occurred between the first and second day, and that loss of carbaryl was much faster from nonautoclaved soil than from autoclaved soil, although carbaryl was effectively degraded in the latter. Kazano et al. (1972) found that the persistence of <sup>14</sup>C-carbonyl-labeled carbaryl incubated with soil (pH 5-6) at 25°C in the laboratory was influenced by soil type. Production of <sup>14</sup>CO<sub>2</sub> varied from 2.2% (loamy sand) to 37.4% (clay loam) of initial radiocarbon during 32 days of incubation, and the amount of residual <sup>14</sup>C in the soil was reportedly roughly proportional to soil organic matter content.

The degradation of carbaryl by isolated soil bacteria and fungi has been reported by numerous investigators, including Boush and Matsumura (1967), Matsumura and Boush (1968), Bollag and Liu (1971a,b), Kazano et al. (1972), Sud et al. (1972), Tu and Miles (1976) and Rodriguez and Dorough (1977). These studies indicate that the bacterial and fungal isolates generally degrade carbaryl in the same manner as observed with soil incubations. Degradation appears to proceed via oxidative modification of the molecule (e.g., N-alkyl and aromatic ring hydroxylation, ring cleavage), leading eventually to hydrolytic cleavage of the ester linkage. A number of metabolites have been identified, including 1-naphthol, 1-naphthyl N-hydroxymethylcarbamate, 4-hydroxy-1-naphthyl methylcarbamate, 5-hydroxy-1-naphthylmethylcarbamate and CO<sub>2</sub>, but 1-naphthol (which is also readily degraded by soil microorganisms) is the major metabolite (Sanborn et al., 1977; Mount and Oehme, 1981).

Although microbial transformation apparently plays a major role in the degradation of carbaryl in soil and the rate of degradation may be accelerated in soils recently treated with carbaryl (Rodriguez and Dorough, 1977), it is anticipated that chemical hydrolysis will be the predominant route of degradation in basic soils with a high moisture content (Section 2.2.). Photolysis is expected to be a significant route of degradation for carbaryl at the surface (Section 2.2.).

Some leaching of undegraded carbaryl in soil is likely to occur. Measured soil sorption coefficients (K  $_{\rm oc}$  ) in the range of 200-400 (Swann et al., 1980; Kenaga and Goring, 1980) and a water solubility of ~40 ppm (see Section 1.2.) suggest that the compound will be moderately mobile. Results of soil sorption and leaching studies indicate that the overriding restraint on movement is organic matter content rather than mineral composition (i.e., clay content) (LaFleur, 1976; Sharom et al., 1980; Aly et al., When a thin layer of soil was surface-treated with 1-100  $\mu mol$ carbaryl/kg, mixed, and extracted with water at room temperature (1/1 soil/ water ratio) to near equilibrium (2 hours), desorption partition constants  $(K_h)$  ranged from 0.12 for Norfolk scl (abbreviation not defined in study) (0.15% organic matter) to 3.7 for Okenee sl (abbreviation not defined; assumed sandy loam) (5.16% organic matter) (LaFleur, 1976). Similar studies with model sand showed that 2% added peat holds carbaryl much more effectively than 20% added kaolin, and that the  $\mathbf{K}_{\mathbf{b}}$  for sand, kaolin and peat was 0.04, 1.1 and 160, respectively. When one pore volume of water ( $\sim$ 6 months rainfall in South Carolina) was added dropwise to the top of a l meter soil column that contained 10  $\mu mol$  carbaryl/kg, carbaryl movement was greatest in Norfolk scl (47% of applied carbaryl in effluent) and least in Okenee sl (0% of applied carbaryl in effluent). In another soil column study, Sharom et al. (1980) found that only 53% of the carbaryl was leached from organic soil (75% organic matter) by 10 rinses of water, whereas 52% of the carbaryl was leached from sand (0.7% organic matter) by the first rinse alone.

In a field study, Caro et al. (1974) determined runoff losses of carbaryl (5.03 kg/ha) that was incorporated into the soil at a depth of 5 cm simultaneously with seed corn. Of the 4 kg of carbaryl applied to the

field, only 5.77 g (0.14%) was lost during the growing season in runoff water and sediments. Over 90% of this loss occurred in a single rainfall 19 days after application, and most of the carbaryl (4.34 g) was lost in the runoff water. Felley (1971) noted that losses of carbaryl in runoff are apparently minor, even when it is applied on the surface and not incorporated into the soil, but these data indicate that a high volume rainfall occurring shortly after carbaryl administration can generate low-level transport.

The low vapor pressure of carbaryl suggests that unadsorbed compound will not significantly volatilize from soil. Lafleur (1976) reported the results of a desorption study indicating that mean carbaryl loss by volatilization during application to soil was <3%. In this study, thin layers of various soils were surface-treated with 1-100 µmol carbaryl/kg in 0.001 M or 0.01 M ethanol solution, the treated soil was tumbled to obtain homogenous distribution and the solvent was permitted to evaporate at least 24 hours. When 10 µmol carbaryl/kg in 0.01 M ethanol was added dropwise to the surface of a l meter soil column, mean carbaryl loss during application was ~1% (Lafleur, 1976). Although not specifically stated, the reported molar concentrations reflect the concentration of aqueous ethanol and not the concentration of carbaryl in the ethanol. Additionally, the volumes of application were not reported.

#### 3. EXPOSURE

#### 3.1. WATER

The mean levels of carbaryl that have been detected in water samples collected at 111 stations in STORET are 4.7  $\mu$ g/ $\Omega$  in unfiltered water (195 samples, range 0-335  $\mu$ g/ $\Omega$ ), 4.8  $\mu$ g/ $\Omega$  in filtered water (67 samples, range 0.04-62  $\mu$ g/ $\Omega$ ), 38.9  $\mu$ g/kg in wet sediment (10 samples, range 0.1-231  $\mu$ g/kg) and 87.3  $\mu$ g/kg in dry sediment (27 samples, range 0.1-770  $\mu$ g/kg). The preponderance of samples was taken from rural stream/river and lake water.

#### 3.2. FOOD

In a 1963-1969 survey of residues in United States foods (Duggan et al., 1971), the percentage of samples of large fruit and grains/cereals with carbaryl residues was 4.1 and 1.4, respectively. Most of the detected residues were between 0.03 and 2.0 ppm.

#### 3.3. INHALATION

Monitoring data on levels of carbaryl in ambient air were not located in the available literature. The mean air concentration of carbaryl in the air inhaled by 38 urban applicators (application of carbaryl incidental to their employment or leisure activities) who made a total of 50 individual applications was determined to be 0.02  $\mu$ g/2, with a maximum recorded level of 0.28  $\mu$ g/2 (Gold et al., 1982); the maximum total respiratory exposure was 0.70  $\mu$ g/kg bw/hour.

#### 3.4. DERMAL

The relative contribution of dermal exposure to total ambient exposure could not be determined from the available literature. The dermal exposure of carbaryl applicators was measured by Gold et al. (1982). The mean rates of exposure were 3.85 and 0.26  $\mu g$  cm<sup>-2</sup> hr<sup>-1</sup>, respectively, for the

outside of the clothing and the skin beneath the clothing. The rate of exposure to the hands of applicators was 2.36 and 24.96  $\mu g$  cm<sup>-2</sup> hr<sup>-1</sup>, respectively, for applicators with and without gloves. The maximum dermal exposure recorded in this study was 2.86 mg kg<sup>-1</sup> hr<sup>-1</sup>.

05/02/84

Ē

#### 4. PHARMACOKINETICS

An abundance of information exists in the literature concerning the absorption, distribution, metabolism and excretion of carbaryl following administration to humans and experimental animals by various exposure routes and under many conditions. Several excellent reviews are available, including IARC (1976), Mount and Oehme (1981) and NIOSH (1976); these reviews were used in conjunction with selected primary papers in the preparation of Chapter 4.

#### 4.1. ABSORPTION

Rapid absorption of carbaryl by both humans and experimental animals following oral, dermal and inhalation exposure has been well documented. Carbaryl absorption by humans occurs following oral ingestion (Farago, 1969; Lopez, 1970), during inhalation exposure (Best and Murray, 1962) or percutaneously during direct contact with the skin (Feldmann and Maibach, 1974; Maibach et al., 1971). Hwang and Schanker (1974) instilled 14C-labeled carbaryl into the lungs and intestines of rats and found that the compound was rapidly absorbed by both the lungs and intestines through the process of simple diffusion; absorption in the lungs was 2.5 times faster than in the Following oral administration of 14C-labeled carbaryl to intestines. rats, Casper et al. (1973) observed rapid and near complete (82%) gastric absorption of the compound within 1 hour after dosing. carbaryl from the GI tract of animals was also reported by other investigators (Houston et al., 1975; Pekas, 1974; Cambon et al., 1981; Mount et al., 1981).

#### 4.2. DISTRIBUTION

A 2-compartment open pharmacokinetic model has been described in experimental animals for carbaryl distribution between a well-perfused central compartment and a more slowly-perfused peripheral or tissue compartment

(Houston et al., 1974, 1975; Pipy et al., 1980). Strother and Wheeler (1980) also described a biphasic model for both pregnant and nonpregnant rats. Fernandez et al. (1982) applied open 2- and 3-compartment models to the kinetics of carbaryl and its metabolites, respectively, in rats. Carbamate metabolites of carbaryl have been reported to undergo enterohepatic circulation in rats, lengthening the time of residence in the body (Marshall and Dorough, 1979; Houston et al., 1974).

Declume and Benard (1977) administered [methyl-14C]-labeled carbaryl orally to pregnant rats, and Strother and Wheeler (1980) administered [ring-14C]- or [carbonyl-14C]-labeled carbaryl by i.p. injection to pregnant rats. Radioactivity crossed the placenta within 1 hour after dosing and was rapidly distributed in the developing fetus. Declume and Benard (1977) detected radiolabeled compound in the eye, liver and brain of the fetus, while Strother and Wheeler (1980) reported radioactivity in the fetal brain, heart and lungs.

Bukin and Filatov (1965) administered a single oral dose of 400 mg carbaryl/kg bw to rabbits, and analyzed the resulting distribution of parent compound by paper chromatography. Carbaryl was rapidly distributed in tissues and excretory fluids within 30 minutes after dosing, as follows (in decreasing order of concentration): bile, urine, kidney fat, heart, liver, spleen, testes, kidneys, lumbar muscles, femoral muscles, cerebellum, medulla oblongata, brain (assumed to mean cerebrum) and lungs. Following a single oral dose of 100, 200 or 300 mg carbaryl/kg bw to rabbits, Bukin and Filatov (1965) did not detect any parent compound in the tissues or organs at 24-80 hours after dosing. Mount et al. (1981) found significant residues of carbaryl in the liver, heart and brain of male rats 24-48 hours after administering single oral doses of 450, 800 or 1200 mg carbaryl/kg bw.

Following a single oral dose of 14C-labeled carbaryl (label location not specified) at a level of 0.9 mg/kg, 14C was detected in the testes, prostate gland and seminal vesicles of male mice (Thomas et al., 1974). Thomas (1981), in his introductory remarks for the 10th Target Organ Sympostum (The Testes), reported that carbaryl has been found in the testes of raits, mice and dogs administered the pesticide.

Following percutaneous administration of carbaryl to steers and cows, carbaryl residues were found in the liver, kidneys, muscles and omental and perirenal fat of steers 3 days after treatment but not 7 days after treatment, and in the milk of cows up to 69 hours after application (Hurwood, 1967). Distribution of carbaryl residues to the tissues was rapid but only temporary in both studies. In most cases, elimination of carbaryl residues from adult body tissues was completed within a few days after treatment.

#### 4.3. METABOLISM

Extensive reviews on the metabolism of carbaryl in mammals are available, including Ryan (1971), Kuhr and Dorough (1976) and Menzie (1969). The metabolism of carbaryl has been investigated in rats (Knaak et al., 1965; Krishna and Casida, 1966; Houston et al., 1975; Sullivan et al., 1972; Pipy et al., 1981; Strother, 1970; Ryan, 1971; Pekas, 1979; Mehendale and Dorough, 1971; Benson and Dorough, 1979; Lin and Dorough, 1974), mice and gerbils (Benson and Dorough, 1979), guinea pigs (Knaak et al., 1965, 1968; Knaak and Sullivan, 1967; Rickard and Dorough, 1979; Ryan, 1971; Benson and Dorough, 1979), rabbits (Ryan, 1971), dogs (Knaak et al., 1965, 1968; Knaak and Sullivan, 1967; Ryan, 1971), monkeys, pigs, sheep (Knaak et al., 1965, 1968; Knaak and Sullivan, 1967), dairy cattle (Whitehurst et al., 1963; Dorough, 1967) and humans (Knaak et al., 1965, 1968; Knaak and Sullivan, 1967; Ryan, 1971; Strother, 1970; Chin et al., 1974). These studies have

identified some of the hydrolytic and oxidative metabolites of carbaryl, as well as several conjugated metabolites of carbaryl. The liver appears to be the primary site of carbaryl metabolism, regardless of species or route of exposure.

Mount and Oehme (1981) described four metabolic pathways for carbaryl, all of which generally yield metabolites of lesser toxicity than the parent compound. These metabolic schemes include the methylol route, by which the methyl group on the nitrogen (carbamic acid) is converted to -CH<sub>2</sub>OH; formation of 1-naphthol by hydrolysis; hydroxylation of the naphthyl ring (at position 3, 4, 5, 6 or 7), possibly via epoxide intermediates, to yield hydroxylated carbaryl metabolites that are subsequently conjugated and then excreted; and glucuronidation of the carboxy group.

Following single oral doses of [naphthyl-14C]- or [methyl-14C]- labeled carbaryl to rats, guinea pigs, monkeys, swine, sheep, dogs and humans, or [carbonyl-14C]-labeled carbaryl to rats, several l-naphthyl and 4-(methylcarbamyloxy) l-naphthyl glucuronide and sulfate conjugates were detected in the urine and/or feces of treated animals (Knaak et al., 1965, 1968; Knaak and Sullivan, 1967). The occurrence of these metabolites in each species is shown in Table 4-1. Knaak et al. (1968) reported that the major difference in carbaryl metabolism between humans and animals is that carbaryl is hydrolyzed to l-naphthol to a greater extent by humans than by other mammals tested. l-Naphthol is conjugated and excreted as l-naphthyl sulfate or glucuronide. Sulfate or glucuronide conjugates of l-naphthol were excreted by all species tested except dogs; monkeys and swine excreted little of these metabolites. All species except dogs also hydroxylated carbaryl, excreting the hydroxylated metabolites as a glucuronide or sulfate conjugate [4-(methylcarbamoyloxy)-l-naphthyl] glucuronide or sulfate]. Dogs

TABLE 4-1

Carbaryl Metabolites Detected in Rats, Guinea Pigs, Monkeys, Swine, Sheep, Dogs and Humans<sup>a</sup>

-	Label	Labeled Forms of Carbarylb		•
set sholl the	Naphthy1-14CC	Methyl-14Cd	Carbonyl-14Ce	Unlabeled Carbaryl <sup>f</sup>
Unidentified neutrals	rats, guinea pigs, monkeys, swine,	rats, guinea pigs, monkeys, swine,	rats	rats, humans (follow- ing both oral and in- halation exposures)
. washing methylcarba-	sheep, dogs guinea pigs	guinea pigs	Q.N	rats
mate N-glucuronide	rats, guinea pigs,	rats, guinea pigs, monkėys, swine,	rats	rats, humans (follow- ing oral exposure)
carbonate-o-glucuronide		sheep, dogs rats, guinea pigs, monkeys, swine,	rats	rats, humans (follow- ing oral exposure)
1-naphthyl glucuronide 1-Naphthyl glucuronide	monkeys, swille, sheep swille, swille, sheep	sheep	QN	rats, humans (follow- ing both oral and in- halation exposures)
4_(Methylcarbamoyloxy)-	rats, guinea pigs,	rats, guinea pigs, monkeys, sheep	rats	rats
i-naphthyl sulfate I-Naphthyl sulfate	rats, guinea pigs, monkeys, sheep	ON	QN	rats, humans (follow- ing both oral and in- halation exposures)

(cont.)	
TABLE 4-	
T	

		Instance of second		
	Labe	Labeled forms of calbaly:		
Metabolites	Naphthy1-14CC	Methyl-14Cd	Carbonyl-14Ce	Unlabeled Carbaryl <sup>f</sup>
Unidentified Metabolite A	rats, guinea pigs, dogs	rats, guinea pigs, monkeys, dogs	rats	rats, humans (follow- ing oral exposure)
massactified Metabolite 8	quinea pigs, dogs	guinea pigs, dogs	ON	ON
Unidentified Metabolite C		sheep	QN	ON
Unidentified Metabolite D	sheep	sheep	ON	ND

ASource: NIOSH (1976); Knaak et al. (1965, 1968); Knaak and Sullivan (1967)

DAnalysis by radiometric techniques; rats and guinea pigs treated intraperitoneally; monkeys, swine, sheep and dogs treated orally.

CAnimals tested include rats, guinea pigs, monkeys, swine, sheep and dogs.

dAnimals tested include rats, guinea pigs, monkeys, swine, sheep and dogs.

eRats were the only species tested.

fAnalysis by fluorometric techniques; rats treated orally; humans treated orally or by inhalation.

ND = Not detected

appeared to conjugate carbaryl (enol form) directly, forming 1-naphthyl methylimidocarbonate 0-glucuronide. All other species tested also excreted this metabolite (Knaak et al., 1965, 1968; Knaak and Sullivan, 1967).

Other identified metabolites of carbaryl include 5,6-dihydro-5,6-dihydroxycarbaryl in rats and guinea pigs (Sullivan et al., 1972) and dairy cows (Dorough, 1967), thioether conjugates in rats (Ryan, 1971), glucuronide or sulfate conjugates of 5,6-dihydroxycarbaryl or N-hydroxymethylcarbaryl in rats (Chen and Dorough, 1979), and mercapturic acids, S-(4-hydroxy-1-naphthyl)cysteine and S-(5-hydroxy-1-naphthyl)cysteine (Bend et al., 1971).

Of the identified metabolites of carbaryl, 5-hydroxycarbaryl and 1-naph-thol may be more toxic than the parent compound under certain circumstances (Carpenter and Weil, 1970; Bollag et al., 1975).

#### 4.4. EXCRETION

The major route of excretion for carbaryl and its metabolites following oral administration for most mammalian species is the urine; however, the dog excretes ~50% of the administered dose in the feces (Knaak et al., 1965, 1968; Knaak and Sullivan, 1967). For rats and dogs administered single oral doses of [naphthyl-14C]-labeled carbaryl, 10 and ~50%, respectively, of the administered radioactivity was detected in the feces. Following single oral administration of [methyl-14C]-labeled carbaryl, 68 and 23% of the radioactivity was detected in the urine of rats and dogs, respectively. When rats were given 14C-carbaryl labeled at various positions, excretion was nearly complete by 3 days. 95% of the naphthyl-label was eliminated in the urine and feces; 99% of the carbonyl-label was eliminated in the urine, feces and respiratory CO2; and ~91% of the methyl-label was eliminated in the urine, feces, CO<sub>2</sub> and carcasses. Guinea pigs receiving i.p. injec-[naphthyl-14C]- or [methyl-14C]-labeled carbaryl either of tions

excreted 85% of the administered doses in the urine within 24 hours. Rats given i.p. injections of [naphthyl=14C]-, [methyl=14C]- and [carbonyl=14C]-labeled carbaryl excreted 73, 47 and 48%, respectively, of the administered doses in the urine within 24 hours.

Less Shah and Guthrie (1977) reported that rabbits rapidly absorbed a dermal dose of 14C-carbaryl (label location not reported), and eliminated the highest concentrations of carbaryl metabolites within 24 hours in the urine and feces. Hurwood (1967) reported that dairy cows sprayed with carbaryl (percutaneous absorption) excreted the intact compound in the milk in decreasing concentrations from 5-77 hours postdosing.

Two human volunteers who ingested a single dose of 2 mg carbaryl/kg bw were reported to have 25% of the administered dose as carbaryl metabolites (measured by fluorometric techniques) in the urine within 4 days after ingestion (Knaak et al., 1968).

#### 5. EFFECTS

### 5.1. CARCINOGENICITY

At present, only equivocal evidence of carbaryl's carcinogenicity in laboratory animals has been reported. The IARC (1976) found the data to be insufficient for an evaluation of the carcinogenicity of carbaryl.

! A contaminant of carbaryl, 2-naphthyl methylcarbamate, can be produced if contaminated 1-naphthol (a carbaryl precursor) is used in manufacture. The 2-naphthyl isomer has been reported to be carcinogenic (Argauer and Warthen, 1975). Four domestic carbaryl samples were analyzed and contained no detectable 2-naphthol methylcarbamate, but four samples of foreign carbaryl contained 0.52-5.60% of the contaminant.

Innes et al. (1969) conducted a carcinogenicity bioassay on many compounds, including carbaryl, for the National Cancer Institute in two strains of first filial generation mice, designated B6C3 and B6AK. Groups of 18 male and 18 female neonates of both strains received 0 or 4.64 mg carbaryl/kg by gavage on days 7-28 of age and thereafter in the diet at a level of 0 or 14 ppm for 18 months. No significant increase in tumor incidence was found among treated groups in any tissues examined histopathologically. Similar groups of mice given an s.c. injection of 100 mg carbaryl/kg dissolved in dimethylsulfoxide on day 28 of age and observed until age 78 weeks had tumor incidences that were not significantly different from the dimethylsulfoxide-injected mice (Innes et al., 1969).

Carpenter et al. (1961) reported that the lung tumor incidence was not increased as compared to nontreated controls in groups of 30 male A/Jax or C3H mice given weekly s.c. injections of 10 mg carbaryl for 5 months.

Carpenter et al. (1961) also reported the absence of increased tumor incidence in CF-N rats fed carbaryl for 2 years. Groups of 20 male and 20

female rats were given diets containing 0, 50, 100, 200 or 400 ppm carbaryl; survivors were killed after 732-736 days of treatment. In addition, gross and histopathological evaluations were performed on auxiliary (concurrently maintained) groups of control and exposed rats at 6, 9, 12 and 24 months of exposure. Survival was unaffected by treatment and no specific tumor type of tumor sites were associated with treatment. The numbers of tumor-bearing rats among the 40 rats/group were 9, 11, 7, 6 and 11 for dietary exposures of 0, 50, 100, 200 and 400 ppm, respectively.

Triolo et al. (1982) observed no significant increase in lung tumor induction in two separate experiments wherein female A/J mice were fed 1000 ppm carbaryl for 20 weeks. Of 16 treated mice in the first experiment, 5 developed lung tumors, compared with 1 lung tumor-bearing animal of 11 untreated control mice. Of a second group of 31 treated mice, 3 had lung tumors compared with 7 lung tumor-bearing animals of 31 untreated control mice. Differences between treated and control groups were not statistically significant.

The carcinogenicity of carbaryl after dietary and s.c. administration has been reported in a Russian study. Andrianova and Alekseev (1970) administered by gavage 30 mg carbaryl/kg to 60 mongrel male rats twice weekly for up to 22 months. A group of 48 untreated male mongrel rats served as controls. Of the 12 surviving treated rats, 3 had fibrosarcomas and 1 had an osteosarcoma. Of the 46 surviving untreated rats, only 1 had a fibrosarcoma. Elevated tumor incidence was reported to be statistically significant (p<0.01). Subcutaneous implantation of a paraffin pellet containing 20 mg carbaryl/kg killed 38 of 48 mongrel rats after 22 months. The 10 survivors had implantation site sarcomas. Of the 48 untreated controls 46 survived 22 months, and 1 animal had a fibrosarcoma. The results of this

study are difficult to interpret due to the high mortality in treated groups, lack of information on control groups and the possible contamination of the test compound (NIOSH, 1976).

Shimkin et al. (1969) gave 16 male A/He mice 12 i.p. injections of 0.5 mg carbaryl in tricaprylin over a 4-week period. Lung tumors were observed in 6 of 15 survivors 20 weeks after treatment. Controls consisted of 28 tricaprylin injected rats and 31 untreated rats. Lung tumors developed in 7 and 2 animals, respectively, in these groups. The percentage of lung tumor-bearing animals in the treated group was not significantly increased over controls (p>0.05). This bioassay was developed as a short-term test to indicate the possible carcinogenic potential of a compound. The usual criteria for a positive response is both an increase in the number of animals with lung tumors and an increase in the number of lung tumors per tumor-bearing animals.

Carbaryl can be nitrosated in the presence of nitrite in mildly acid milieu, such as the human stomach (IARC, 1976), in vitro in dilute aqueous solution at pH 1-3.5 (Eisenbrand et al., 1975; Elespuru and Lijinsky, 1973) or in rat gastric juice in vitro (Beraud et al., 1979). N-Nitrosocarbaryl induced lung tumors in all surviving (14/16) rats given a single s.c. injection of 1000 mg/kg and observed for 450 days. No tumors were observed in controls (Eisenbrand et al., 1975).

No significant incidence of malignant tumors was observed in female Sprague-Dawley rats given 300 mg carbaryl by gavage over a 10-day period or 90 mg carbaryl and 120 mg sodium nitrite by gavage over a 3-day period and observed until death (Lijinsky and Taylor, 1977).

#### 5.2. MUTAGENICITY

Carbaryl has been evaluated for its potential mutagenicity in several short-term tests using bacteria, fungi, <u>Drosophila</u>, mammalian cells in culture, and whole-mammal assays. The biological endpoints for determination of the potential mutagenicity of carbaryl were: gene mutation, structural/numerical chromosome aberrations, <u>DNA</u> repair synthesis, mitotic recombination, and sister chromatid exchange. The available studies on carbaryl are outlined in Table 5-1. A detailed evaluation of the mutagenicity data on carbaryl can be found in U.S. EPA (1981b).

The ability of carbaryl to cause gene mutations in bacteria has been extensively studied (see Table 5-1). The majority of the test results were reported as negative. It should be emphasized that the negative results are not wholly unequivocal. Appropriate concurrent controls and adequate concentration ranges were sometimes not used, an exogenous metabolic activation system was not always included, and in some cases the available data are not sufficient to determine whether an adequate test was conducted. There were some reported occurrences of weak responses (Egert and Greim 1976, Jaszczuk et al. 1979). Because these responses were weak at high doses, the possibility of an impurity or impurities with mutagenic activity should be considered.

Carbaryl has also been reported as weakly mutagenic in eukaryote assays, namely a <u>Drosophila</u> sex-linked recessive lethal test (Brzeskij and Vaskov 1971) and a forward mutation assay using Chinese hamster V79 cells (Ahmed et al., 1977a). These studies have deficiencies and are considered merely suggestive evidence of weak mutagenicity. Wojciechowski et al. (1982) reported carbaryl as negative in a Chinese hamster V79 gene mutation test.

Assay	Indicator Organism	Application	Concentration or dose	Exogenous Activation System	Responsea	Comments	Reference
A. Gene mutation tests:	ssts: Bacteria		ļ	Ç			De Lorenzo et
2 2	S. typhimurium TA1535 TA1537 TA1537	plate incorporation	10-1500 µg/plate	50 +1			
	TA100 TA100 S. typhimurium	plate	60-1000 µg/plate	6S+I	я и «	·	Marshall et al., 1976
mutation	TA1536 TA1537 TA1538	i i	ה מחחר ה	6S+	, и в		McCann et al., 1975
Reverse mutation	S. typhimurium TA1535 TA1537 TA100 TA98	plate incorporation		1			Shirasu et al.,
Reverse mutation	<u>s. typhimurium</u> TA1535 TA1536 TA1537	plate incorporation	0.02 µg/plate		1111		1976
Reverse mutation	1A1538 <u>1 typhimurium</u> 1A98 TA100	spot test	50 nmo 1		111		1977
Reverse mutation	1A1537 1A1538 S. <u>typhimurium</u> 1A198 1A1535	plate incorporation	0.00115-11.5 µg/plate		1 (1)(	<del>-</del> -	Blevins et al., 1977
	TA1537 TA1538				•		

Assay	Indicator	Application	Concentration or dose	Exogenous Activation System	Responsed	Comments	Reference of the second of the
Reverse mutation	S. <u>typhimurium</u> S. <u>typhimurium</u> TA1535 TA1000 C3076 TA1537	gradient plates		6S+1			1981 1981
Reverse mutation	03052 TA1538 TA98 S. <u>typhimurlum</u>	liquid suspension	100 PM	+ mouse microsomes plus cofactors	, A	Weak response after activation. A dose-response was not deter- mined and positive findings were not confirmed on a repeat test.	Egert and Greim, 1976
Reverse mutation	S. typhimurium TA1535 TA1536	plate incorporation	500-1000 µg/ plate	6S+	,		Marshall et al., 1976
Reverse mutation	171537 171538 5. <u>typhimurium</u> 17198	æ	0.25-1000 ng/ plate		KKE P P	Although a weak increase in the number of revertants was reported, there was no indication whether this increase was 2-fold or greater.	Jaszczuk et al., 1979
;	TA1535 TA1537	spot test	10% in buffer		•		Ashwood-Smith et al., 1972 probst et al.,
Meverse mutation Reverse mutation	E. CO11 MP2	plate incorporation	100 nmol/mt	6S+1	и и		1981 . Uchłyama
Reverse	E. CO11 8/r MP2	spot test	0.10 mg/plate		g g 3	-	Shirasu et al., 1976
Meverse mutation	MP2 MP2 UVEA						

						<b>6.</b> F.	
	Indicator	Application	Concentration	Exogenous Activation System	Responsea	Comments	Reference
Assay	Organism						Nagy et al.,
67 Lexe0	E. col1	spot test	NR.		1 1		6.6
mutation	WP2 UVEA			657	, ,		probst et al.,
Reverse	E. <u>coll</u> WP2	gradient plates		<b>2</b> 1			Egert and
Reverse	E. CO11	11qu1d suspension	MR.	+ mouse microsomes plus cofactors	N.		Greim, 1976
mutation	nad", arg",				1		Fiscor and Nil
9740	E. col1	spot test	æ				1972
mutation	K12, lac- leu-, cys-			97,0	,		DeGlovann1- Donnelly
Reverse	Bacillus		<b>~</b>	t 5			et al., 1968
mutation	-1891 -1891				,		Fahrig, 1974
Reverse	Serratia	spot test	œ Z		,		Fahrlg, 1974
Forward	E. col1	liquid incubation test, spot test	æ				Elespuru
Forward	Haemoph 1 lus	11qu1d suspension	0.1 mM for 1-9 min.		ŧ		
mutation	(novoblocin resistance)			,		•	

-29-

Assay	Indicator	Application	Concentration or dose	Exogenous Activation System	Responsea	Comments	
	Organism	ernit in a second	a titte				
A. Gene mutation sex-linked sublethal mutations	D. melanogaster Fed		1% suspension in a diluted sugar syrup for 24 hours		+wk b	Although authors conclude carbary! is a weak mutagen to <u>D</u> .  melanogaster, inducing a low rate of recessive sex-linked lethal and sublethal mutations, their evidence is considered only suggestive because of an inadequate sample size.	Vaskov, 1972
Fortward	Chinese hamster lung V79 cells	cell culture	0.01 mM for 2 days		+wkb	No dose-response relationship demonstrated	Ahmed et al., 1977a
mutation (ouabain Forward mutation (ouabain resistance)	Chinese hamster lung V79 cells	cell culture	0.01-0.1 mM for 3 days	+ irradiated Syrian hamster cells	1		Wojciechowski et al., 1982
B. Cytogenetic tests: Nondisjuction Aspe	tests: Lower eukaryotes Aspergillus P	votes plate test	0.1 mg/m£				Morpurgo et al., 1979
B. Cytogenetic tests: Root +in mitotsis:	tests: Plants	root treatment	25 ppm and 50 ppm for 24 hours		. <b>.</b>	Dose-related C-mitotic effects (multipolar anaphases, mitotic arrest, chromosome lagging, tetraploidy).	
Allium cepa Root tip mitosis:		root treatment	25, 50, and 100 ppm for 4 hours		•	Dose-related C-mitotic effects. (chromosome lagging, tetra-ploidy, chromosome stickiness, bridges, mitotic arrest).	. Amer et al., 1971
Victa faba and Gossyptum barbadense		seed treatment	100 ppm for 6, 12, 24, and 48 hours		+		

0485p

-30-

05/09/84

						-	
Assay	Indicator Organism	Application	Concentration or dose	Exogenous Activation System	Responsea	Comments	Reference
Pollen mother cell melosis: Vicia faba		2-week-old plants sprayed weekly for one month	saturated		<b>+</b>	Abnormal melosis (anaphase bridges, chromosome lagging, stickiness)	Amer and Farah, 1968
Melotic effects in corn		seed treatment and injection into anthers of plants	0.12% and 0.25% aqueous solution for 48 hours (seeds) and 6 hours (anthers)		*	Abnormal metosis (bridges, stickiness)	Brankovan, 1972
M1 tos1s: Hordeum		seed treatment	500, 1000, 1500 ppm for 6, 12, and 24 hours		•	C-mitotic effects (chromosome lagging, fragments, bridges)	Wuu and Grant, 1966
Metosts: Hordeum vulgare		seed and plant treatment	1000 ppm for 12 hours (seeds) and 500 ppm (plants)		•	Abnormal metosis (e.g., chromosome stickiness, bridges, univalents, polyploidy, fragments, micronuclei)	Muu and Grant, 1967
B. Cytogenetic tests:	tests: Mammallan cells in culture	s in culture					Sabbarwal and
Polyploidy	Chinese hamster V79 cells		5x10-* to 10-* M		<b>+</b>		Lockard, 1979
Polyploidy	Chinese hamster fibroblast cells		0.0075-0.03 mg/mt		+	Increase at 0.0075 mg/mm	Ishidate and Odashima, 1977
Aneuploidy/ polyploidy	Chinese hamster V79 cells		0, 50, and 100 for 1-3 days	6S+1	*	Increase in frequency of cells with elevated chromosome numbers (>22) at 100 pM. This effect was decreased by addition of glutathione or S9. Multiple chromatid exchanges observed.	Onfelt and Klasterska, 1983
Human embryonic fibroblasts			20, 40, and 80 yg/mt for 24 hours		•	29.2% of cells aneuploid (primarily hypodiploidy) at 80 µg/mL; chromosome fragments also found.	Kazarnovskaya and Vasilos, 1977

Assay	Indicator	Application	Concentration or dose	Exogenous Activation System	Responsed	Comments	Reference
Human embryonic fibroblasts			20, 40, and 80 µg/mt for 6, 24, and 48 hours, technical pro-		+	Dose-related C-mitotic effects	Vasilos et al., 1972
Human embryon1c f1broblasts		j.	duct reported as containing 84% active ingredient 0.001, 0.01, 100, and 1000 mg/mk		<b>.</b>	Dose-related C-mitotic effects	Shp1rt, 1975
B. Cytogenetic tests: Mitotic studies in epithelium from small	sts: Whole mammals		85% commercial preparation, acute: 400 mg/kg (one-half LDsn).		•	C-mitotic effects and chromosome fragmentation	Vasilos et al., 1975a
intestine crypts and cornea of rats Mitotic studies follicles,			80 mg/kg, 40 mg/kg, and 20 mg/kg 85% commercial preparation, sub- acute: 5 and 20 mg/kg (28 adminis-		•	C-mitotic effects and chromosome fragmentation	Vasilos et al., 1975b
lium, and epithelium of glandulae intes- tinales of rats Micronucleus test in mice			thronto: 0.05 to 8 mg/kg/day for 6 months orally 10** M 1.p. or dally for one week via intuba- tion		1	•	Degraeve et al., 1976
Micronucleus test in mice	Swiss male	fed in distilled water	two doses of 146 mg/kg separated by a 24-hour interval		,	-	Rant et al., 1980

Assay	Indicator	Application	Concentration or dose	Exogenous Activation System	Responsea	Comments	Reference
	Organism		- 1				Well et al., 1973
Dominant lethal	male rats	diet	200 mg/kg/day on 7 successive days				Well et al.,
Dominant	male rats	oral	100 mg/kg on 7 successive days		1		1973 Epstein et al.,
letha! Dominant Tethal	m1ce.	oral	50 and 100 mg/kg/day on 5 succes-sive days		t .		Rani et al.,
Host- mediated reverse mutation	S. typhimurium G46	of S. typhimur- of S. typhimur- lum into male Swiss mice 24 hours after last oral treatment	438 mg/kg/day orally on 3 successive days		1		200
C. Other studie	Other studies indicative of DNA-damaging activity	imaging activity	10"s to 10"4 M		*	Dose-related increase at	Sabharwal and Lockard, 1979
Sister chromatid exchange	lung V79 cells		, ;	87	*	Dose-related increase at 21	Ahmed et al., 1977b
Unscheduled DNA synthesis	SV-40 transformed human fibroblasts	cell culture	1-1000 pM for 8 hours	Ŝ.		um; Sy dia not increase uos activity	probst et al.,
Unscheduled DNA synthesis	primary rat hepatocytes (Fischer 344)	cell culture	100 nmol/mt for 5 hours		, •	22% inhibition on DNA synthe-	Rocchi et al.,
DNA synthesis	rat thymocytes (Wistar)	cell culture	1-100 ug/mt		. +	sis at 10 ug/mt 62% inhibition on DNA synthe-	Rocchi et al.,
DNA synthesis	human lympho- cytes	cell culture	50 µg/mg		. 1	sis at 50 µg/ml	Rocchi et al.,
Unscheduled	human lympho- cytes	cell culture	50 µg/m£				

05/09/84

Assay	Indicator	Application	Concentration or dose	Exogenous Activation System	Responsed	Comments	Reference
	Organism				1		Regan et al.,
DNA strand breaks as	human skin fibroblasts	cell culture	100 pM for 1 hour		ı		9/61
by sedimenta- tion profiles							Stebert and
Mitotic gene	S. cerevistae	cell culture	1000 ppm (4.97mM) dissolved in DMS0		ŧ		Elsenbrand, 1974
conversion			for 16 hours				Fahr19, 1974
Mitotic gene	S. cerevisiae	liquid incuba- tion test	Œ.		i		Uchiyama
Recombination	Bacillus	æ	0-10 mg/plate		1 1		et al., 1975
	Marburg 17a						

\*Responses: +, positive; -, negative; -, negative with and without S9 activation system; <u>+</u>, positive with and negative without S9 activation system negative with and positive without S9 activation system -34-

buk = weak or borderline mutagenic response

NR = Not reported; 1.p. = intraperitoneal

Other tests indicative of DNA-damaging activity have primarily been negative. These included tests for DNA strand breakage and mitotic recombination (Uchiyama et al., 1975; Fahrig, 1974; Regan et al., 1976; Siebert and Eisenbrand, 1974). Inconsistent results have been reported for DNA repair tests (Ahmed et al., 1977b; Probst et al., 1981; Rocchi et al., 1980). Sister chromatid exchange formation in V79 cells was reported as positive (Sabharwal and Lockard, 1979).

Negative results have been reported for carbaryl in whole-mammal tests These included which detect clastogens (chromosome-breaking agents). dominant lethal assays in rats (Weil et al., 1973) and mice (Epstein et al., 1972), and micronucleus assays in mice (Degraeve et al., 1976; Rani et al., 1980). The micronucleus assay is also thought to detect agents that affect However, several other cytogenetic studies in the spindle apparatus. meiotic and mitotic cells of plants, mitotic mammalian cells in culture, and mitotic cells in whole mammals strongly suggest that carbaryl affects the Chromosome nonspindle apparatus and causes chromosome nondisjunction. disjunction leads to aneuploidy (loss and gain of whole chromosomes) or polyploidy (increase of chromosome number in multiples of the basic number), which are considered to be significant mutagenic effects. Although much of the evidence for aneuploidy induction is indirect [i.e., observations on cell division where treatment caused mitotic arrest, chromosome lagging, etc.; generally referred to as -mitotic effects (i.e., colchicine-like action)], there were some studies on aneuploidy induction. For example, the recent study of Onfelt and Klasterska (1983) showed an increase in aneuploid and polyploid cells after treatment of Chinese hamster V79 cells with carbaryl. Although the fungi test for chromosome nondisjunction was reported as negative (Morpurgo et al., 1979), colchicine, a well-known spindle inhibitor, appears to be ineffective at inducing aneuploidy/polyploidy in lower eukaryotes [see Bond and Chandley (1983) for review].

It should be pointed out that there were deficiencies in many of these cytogenetic studies; e.g., the purity of the carbaryl was sometimes not given, the toxicity of the concentrations tested was not described, and the frequency of each specific aberration was not reported. Nevertheless, the consistency of the positive results obtained by different investigators using different test systems strongly suggests that carbaryl has the potential to cause aneuploidy/polyploidy. In addition, carbaryl is a carbamate pesticide, and other carbamates, such as methyl benzimidazole carbamate, have been reported to cause chromosome nondisjunction [see Bond and Chandley (1983) for review].

If a mutagen reached the germinal tissue, it would have the potential to cause mutations that may contribute to the burden of genetic disease. Germinal numerical chromosome aberrations would contribute to human morbidity and mortality by spontaneous abortions and various genetic disorders (e.g., Down's syndrome, Turner's syndrome, etc.) [see Hook (1983) for Although there are no studies for meiotic nondisjunction in mammals, other studies provide suggestive evidence that carbaryl (or an active form or forms of the chemical) may reach mammalian germ tissue. For example, Wyrobek et al. (1980) reported elevation of sperm abnormalities in carbaryl-exposed workers. Krylova and Denisova (1973) also reported pathological changes in spermatozoa of a small rodent called the Mongolian tree creeper, which inhabited an area sprayed with carbaryl. In laboratory studies, adverse effects on spermatozoa were seen in carbaryl-treated rats and mice (Kitagawa et al. 1977, Degraeve et al. 1976, Shtenberg and Rybakova In contrast, some investigators have reported no significant gonadal effects attributable to carbaryl (Dikshith et al. 1976, Weil et al. 1972).

In summary, while the majority of studies concerning the mutagenicity of carbaryl contain differences, the body of evidence nevertheless indicates that carbaryl is not very active at causing gene mutations or structural chromosome aberrations. The available studies, however, do strongly suggest that carbaryl may induce numerical chromosome aberrations (aneuploidy/polyploidy) as its mutagenic endpoint. All of the mammalian studies were on mitotic cells. It should be noted that with respect to somatic-numerical cell risk, aneuploidy induction has been proposed as a possible chromosomal mechanism in carcinogenesis and tumor promotion (Tsutsui et al. 1983, Onfelt and Klasterska 1983). Although no testing of meiotic mammalian nondisjunction has been performed, other studies have suggested that carbaryl may reach the germ tissue in mammals, and thus may carry a first-generation risk (e.g., embryonic, fetal, and infant death, and genetic disorders).

#### 5.3. TERATOGENICITY

Oral administration of carbaryl was teratogenic to rabbits, guinea pigs, beagle dogs and miniature swine, but not to mice, rats, hamsters or monkeys (Murray et al., 1979; Robens, 1969; Weil et al., 1973; Smalley et al., 1968; Earl et al., 1973; Benson et al., 1967; Coulston, 1971; Dougherty and Coulston, 1975). These teratogenicity tests are summarized in Table 5-2.

Equivocal determinations were obtained for fetotoxicity with mice and for teratogenicity with rabbits and guinea pigs. It is possible to explain these varying results by examining inadequacies in experimental design (small number of animals tested, only a single dose level tested), different methods of administration (gavage and dietary), the presence or absence of maternal toxicity, or some combination of these. Oral administration of carbaryl by gavage at a dose of 150 mg/kg/day to mice produced significant maternal toxicity, but no embryotoxicity or teratogenic effects, while

TABLE 5-2

Teratogenicity Testing of Carbaryl

F

e U	.	· ·	# :	et al.
Reference		Murray et al., 1979	Muray et al., 1979	Benson et al., 1967
Second Leten	retel nespons	No significant incl- dence of fetal mal- formations	Significantly decreased mean fetal body weight and length; delayed ossification of skull-bones and of sternebrae	No difference in fetal mortality or body weight between treated and control fetuses. At the 30 mg/kg level, an incleated abnormalities in 2 litters, as compared with 2 such abnormalities in controls, was not controls. Was not controls authors
	Maternal Response	100 mg/kg/day: 1 of 23 treated died 150 mg/kg/day: significant maternal toxicity (salivation,	statistically signif- icant (P<0.05) number of maternal deaths (10 of 37) No significant ad- verse effects	No difference between treated and control dams in mortality. behavior, physical condition, resorptions or fetal deaths
Obcorva-	tion Day*	81	. <b> </b>	normal partu- rition
	Treatment Days	6-15 of gestation	6-15 of gestation	6 through end of gestation
	Daily Dose or Exposure	0, 100 or 150 mg/kg/ day	0 or 5660 ppm (equiv- alent doses of 0 or 1166 mg/kg/day)	0, 10 or 30 mg/kg
	Vehicle	cottonseed o11	diet	diet
	No. Dams	41 control; 23 low dose; 37 high dose	35 control;	20/dose level
	Species/	Strain mice/CF-1	m1ce/CF-1	a) ce
		Route oral, gavage	oral, diet	oral, diet

	Species/	No. Dams at Start	Vehicle	Daily Dose or Exposure	Treatment Days	Observa- tion Day*	Maternal Response	Fetal Response	Reference
oral, gavage	rabbits/ New Zea- land white	28 control; 20 low dose; 14 high dose	cottonseed	0, 150 or 200 mg/kg/ day	6-18 of gestation	83	Significant maternal toxicity at 200 mg/kg/day level; mild maternal toxicity at 150 mg/kg/day level	single incidence of shall be incidence of tically significant decrease in fetal body weight (P<0.05) 200 mg/kg/day: significantly increased incidence of omphalocele (P<0.05): statistically significant decrease in delayed ossification of the fifth sternebra (P<0.05)	Murray et al., 1979
oral	rabbits/ New Zea-	21 control; 9 low dose; 4 mid dose;	gelatin capsules	0, 50, 100 or 200 mg/kg	5-15 of gestation	<b>58</b>	No treatment-related maternal toxicity	No treatment-related embryotoxicity or in- duction of terata at any dose tested	Robens, 1969
gavage	white hamsters/ Golden Syrian		0.5 g sodium carboxymethyl cellulose, 0.4 g lween 60, 0.9 g sodium chloride, 2.0 g benzyl alco-	0 or 125 mg/day	6-8 of gestation	14 or 15	Maternal toxicity, manifested in diar- rhea, salivation and incoordination	No treatment-related effect on fetal mortality or body weight; no increase in incidence of bone anomalies	Robens, 1969
oral, gavage	hamsters/ Golden Syrian	/ 10 control; 6 treated/ dose level	hol and water 0.5 g sodium carboxymethyl cellulose, 0.4 g Iween 60, 0.9 g sodium chloride, 2.0 g benzyl alcohol and water	0 or 250 mg/day	7 or 8 of gestation	14 or 15	Maternal toxicity, manifested in diar- rhea, salivation and incoordination; lethal to 2 of 6 dams	Increased fetal mor- tality; no effect on fetal body weight; no increase in incidence of bone anomalies	Robens, 1969

	Species/ Strain	No. Dams at Start	Vehicle	Daily Dose or Exposure	Treatment Days	ubserva- tion Day*	Maternal Response	Fetal Response	Reference
	guinea pigs/ Coulson	31 control; 26 multiple doses; 40 single doses	gelatin	0 or 300 mg/kg	multiple doses on doses on or single doses on day 11, 12, 13, 14, 15, 16, 17, 18, 19 or 20 of gesta- tion	<b>95</b>	Multiple doses on days il-20 were lethal to 10 of 26 dams treated. Single doses were lethal to 5 of 40 dams treated.	Multiple doses: In- creased fetal mortal- ity; decreased litter size; incidence of il terata as compared to 2 for controls. Ter- ata in treated ani- mals were bone de- fects occurring most frequently in the cer- vical vertebrae. Single doses: average litter size and per- cent fetal mortality were similar to con- trol values; inci- trol values; inci- dams were exposed on day 12, 13, 14, 15 or 16; treatment on day 13 resulted in 2 fetuses with no kid- neys or genital or- gans, one of which also had fused tho- racic vertebrae and ribs; other terata in treated animals were bone defects occurring most fre- quently in the cer- vical vertebrae.	
oral, peroral intuba- tion	guinea	controls and treated: 5/ dose level except 10/ dose level on days 10-24 of gestation	corn of l	0, 50, 100 or 200 mg/ kg	10 and 11; 12; 13; 14; 12-14; 15; 16; 15 and 16; 17-19; 20- 24; 10-24 of gesta- tion	35	Decreased mean body weights	No treatment-related embryotoxic or teratogenic effects at any dose level tested	Weil et al., 1973

Species/ Strain	No. Dams at Start	Vehicle	Dally Dose or Exposure	Treatment Days	Observa- tion Day*	Maternal Response	Fetal Response	Reference	
guinea	controls and treated: 5/ dose level except 10/ dose level on days 10- 24 of gesta- tion	diet	0, 100, 200 or 300 mg/kg	10 and 11; 12; 13; 14; 12-14; 15; 16; 15 and 16; 17-19; 20- 24; 10-24 of gesta- tion	34 or 35	No treatment-related maternal effects at any dose level tested	No treatment-related embryotoxic or teratogenic effects at any dose level tested	Ne11 et al., 1973	
	¥	corn of)	0, 3, 7, 25 or 100 mg/kg	various intervals throughout gestation or until weaning of the pups	18 or 19	100 mg/kg: signifi- cantly increased per- cent mortality (P<0.001)	cantly decreased mecantly decreased median number of total and vlable fetuses (P<0.05); significantly increased percentage of litters with resorption sites (P<0.05); no treatment-related teratodose level tested	Weil et al., 1973	
	¥	d tet	0, 7, 25, 100 or 200 mg/kg	various intervals throughout gestation or until weaning of the pups	18 or 19	No treatment-related maternal effects at any dose level tested	No treatment-related embryotoxic or tera-togenic effects at any dose level tested	Neil et al., 1973	

Reference	Smalley et al., 1968; Earl et al., 1973	
Fetal Response	At increasing dose levels, decreased body weight gain and increased mortality in nursing pups when compared to controls, probably due to an effect of carbaryl on lactation; decreased lactation; decreased lactation; decreased lavels tested. Teratogenic effects: 0 pups (DK) at 0 and 3.25 mg/kg; 3 pups (DK) at 6.25 mg/kg; 14 pups (18%) at 12.5 mg/kg; 1 pup (14%) at 25 mg/kg; 1 pup (14%) at 60 mg/kg; teratogenic effects were defined to include abdominalthoracic fissures with varying degrees of intestinal agenesis and displacement, varying degrees of brachygnathia, acaudate pups, failure of skeletal formation and superfluous phalanges	
Maternal Response	Increased number of difficult births (dystocia) among treated animals — placental separation from uterus and atonture was observed in these animals; significantly decreased reproductive capacity at 50 mg/kg level	
Observa- tion Day*	normal partu- rition, except 6 on day 55	
Treatment Days	throughout gestation (average gestation period = 62 days)	
Dally Dose or Exposure	0, 3.125, 6.25, 12.5, 25 or 50 mg carbaryl/kg body weight	
Vehicle	diet	
No. Dams	16, 10, 10, 18, at the given dose levels	
Spec1es/	beagle dogs	
	Route oral, diet	

	Reference	Earl et al.,	F 1
	Fetal Response	9 mg/kg/dav: 22% re-	sorptions (control percent not determined); in litter, 2 fetuses with crooked tails, I runt, and ananous fetus with anomalotrophic effects. 16 mg/kg/day: 8x resorptions (control percent not determined); increased percent still-born (13.0% as compared to 6.1% for controls); i abnormal fetus with ecaudate nanosomus and phocomelia of all limbs.
	Maternal Response		8 mg/kg/day: 3 sows had irregular estral periods: 1 sow had pyometra; decreased percent of pregnan- cles 16 mg/kg/day: decreased percent of pregnancles
	Observa-		normal partu- rition
	Treatment	udys	20 days before mating and throughout gestation or from 7 days after mating to the end of gestation
	Daily Dose	or Exposure	0, 4, 8 or 16 mg/kg/ day
-		Vehicle	diet
		Mo. Dams at Start	6, 5, 7 or 7, respectively, at the given dose levels
		Species/ Strain	miniature swine
		Route	oral,

\*Day O of gestation was designated as the day of mating (Murray et al., 1979) or the day after mating (Robens, 1969). Day 1 of gestation was designated vaginal plug for band of gestation was designated vaginal plug for beagle dogs (Smalley et al., 1968) and miniature swine (Earl et al., 1973), or as the observation of a vaginal plug for guinea as the day of mating for beagle dogs (Smalley et al., 1968) and miniature swine (Earl et al., 1976), and the designation of day O or la plug and rats (Well et al., 1973). The remaining studies were only available as abstracts or summaries in NIOSH (1976), and the designation of gestation was not reported.

NR = Not reported

dietary administration of carbaryl to mice at a much larger dose produced mild fetotoxicity in the absence of maternal toxicity, (Murray et al., 1979) and at a smaller dose produced equivocal evidence of fetotoxicity (Benson et Significant maternal toxicity, embryotoxicity and terata al., 1967). (emphalocele) were observed when rabbits were given gavage doses of 200 mg carbaryl/kg/day (Murray et al., 1979), while neither maternal toxicity, embryotoxicity, nor teratogenic effects were observed at the same dose given orally as gelatin capsules (Robens, 1969). The small number of animals (4 or 9/dose level) used by Robens (1969), however, may not have been sufficient. A significantly increased incidence of teratogenic effects was observed in the fetuses of guinea pigs when the dams were given a single dose level of carbaryl that induced significant maternal lethality (Robens, 1969). Weil et al. (1973) did not find similar results in guinea pigs after testing carbaryl under various dose levels, two methods of administration and many different exposure periods during organogenesis.

Earl et al. (1973) administered carbaryl to miniature swine at a level of 0, 4, 8 or 16 mg/kg/day in the diet for 20 days before mating and throughout gestation, or from 7 days after mating to the end of gestation. One malformed fetus was observed at the 8 mg/kg/day level, and another at the 16 mg/kg/day level. In a follow-up experiment, Earl et al. (1973) administered carbaryl to miniature swine in dietary concentrations of 0, 16 or 32 mg/kg/day for 20 days before mating and throughout gestation, or from 7 days after mating to the end of gestation. Neither maternal toxicity nor teratogenic effects were observed at the two dose levels. The carbaryl used in this follow-up experiment, however, had been stored for 12-15 months (between the two experiments). Therefore, Earl et al. (1973) speculated that the carbaryl may have partially decomposed, resulting in a loss of teratogenic potential.

# 5.4. OTHER REPRODUCTIVE EFFECTS

Several studies of the effects of carbaryl administration on the reproduction of experimental animals were encountered.

Collins et al. (1971) administered carbaryl to groups of 20 pairs of weanling Osborne-Mendel rats (mated at 100 days of age) in dietary levels of 0; 2000, 5000 or 10,000 ppm through three generations. For each dietary level, the parental rats  $(F_0)$  were mated to produce two  $F_1$  generations, the second of which  $(F_{1b})$  produced two  $F_2$  generations, and the  $F_{2b}$  rats produced the two  $F_3$  generations. Only the 10,000 ppm level caused decreased female fertility, with  $F_2$  rats failing to produce an  $F_{3b}$  generation. Mortality of pups was increased at the 5000 and 10,000 ppm levels. Dose-related decreases in litter size and mean weanling body weights were significantly different from control values.

The results in Mongolian gerbils treated similarly (Collins et al., 1971) were as follows.  $F_3$  generation gerbils receiving carbaryl in the diet at a level of 10,000 ppm produced one litter, but not a second litter. At lower dose levels, reduced fertility, reduced litter size and reduced pup viability were observed in  $F_2$  and  $F_3$  rats, but not in a dose-related manner. The number of pups surviving to weaning was significantly reduced from control values at  $\geq$ 4000 ppm in all generations and at 2000 ppm in  $F_2$  and  $F_3$  gerbils.

In another 3-generation reproduction study (Weil et al., 1973), 4-month-old  $f_0$  rats were mated to produce the  $f_{1a}$  generation, and at 3-month intervals thereafter they were remated to produce the  $f_{1b}$  and  $f_{1c}$  generations.  $f_{1a}$  rats (4 months old) were mated to produce  $f_{2a}$  rats, which were mated at 4 months of age to produce  $f_{3a}$  rats, at 6 months of age to produce  $f_{3b}$  rats and at 7.3 months of age to produce the  $f_{3c}$  generation.

The  $F_0$  parents were given the first dose of carbaryl at 5 weeks of age. The pesticide was administered daily either in the diet at 0, 7, 25, 100 or 200 mg/kg bw (a 100 mg/kg plus corn oil group was also included, probably to observe effects of the oil on absorption; however, the reason for this group was not discussed by the authors) or by gavage in corn oil at 0; 3, 7, 25 or 100 mg/kg. Rats receiving 100 mg/kg/day by gavage had daily signs of cholinesterase inhibition throughout the 3 generations, significantly increased mortality of male and female parents of each generation, decreased number of pups born alive ( $F_2$  and  $F_3$  pups), prolonged periods between first mating and birth of a litter (F<sub>lb</sub>), decreased number of mated females to produce litters ( $F_{1b}$ ), reduced median number of total viable fetuses and an increased percentage of litters with resorption sites. Lower doses by gavage were without effect. Dietary administration resulted in no increased mortality at any dose in any generation. The  $F_{2a}$ generation receiving dietary administration of 100 or 200 mg, carbaryl/kg/day had prolonged periods between first mating and birth of a litter (Weil et al., 1973).

As reviewed by NIOSH (1976), a 3-generation reproduction study from the Russian literature (Shtenberg and Ozhovan, 1971) reported that very low doses of carbaryl, 2 and 5 mg/kg, administered by gastric intubation to "second ( $F_2$ ) through fifth ( $F_5$ ) generation" rats for 6 months (matings were performed after 4 months of treatment) resulted in reproductive effects. Treated male rats of the  $F_2$ ,  $F_3$  and  $F_4$  generations had significantly reduced sperm motility, spermatogenesis and sperm survival when compared with controls. Treated females of the  $F_3$  and  $F_4$  generations had estrus cycle changes (shortened estrus, prolonged interestrus) which were apparent after 3 months of treatment with 5 mg/kg and at 6 months with

2 mg/kg doses. Dose-related reductions of litter size occurred in all generations as did decreased pup survival. Histological examination of testes and ovaries from treated rats revealed dystrophic spermatogenic epithelia and sclerotic ovarian follicles.

Administration of 50 mg carbaryl/kg by gavage to pregnant Sprague-Dawley rats on day 18 of gestation caused inhibition of acetylcholinesterase activity in blood, brain and liver of both the dams and the fetuses within 30 minutes of treatment (Cambon et al., 1978, 1980). The effects persisted in the dams for up to 12 hours in blood, 5 hours in brain and 24 hours in liver. In fetuses, the inhibition of the enzyme persisted for 12 hours in blood, 12 hours in brain and 5 hours in liver. When carbaryl at doses of 6.25, 12.5, 25 or 50 mg/kg was administered, inhibition of acetylcholinesterase in the blood of dams and fetuses was seen at the lowest dose. Maternal brain tissue enzyme was also inhibited at 6.25 mg/kg, but effects on fetal brain acetylcholinesterase activity did not occur at doses <50 mg/kg. Doses of 1-5 mg/kg administered from day 11 or 19 of gestation until term were without effect (Declume et al., 1979).

Mature female rhesus monkeys treated with carbaryl at 2 or 20 mg/kg/day by gavage throughout gestation had increased, although not directly dose-related, incidences of abortions as compared with vehicle-treated controls (Coulston, 1971). The number of monkeys/group was too small to allow any conclusions to be drawn.

A follow-up study, in which 15-16 monkeys/group were treated orally with carbaryl at 0, 0.2, 2.0 or 20 mg/kg/day (in capsules) from day 20-38 of gestation, did not show an association between carbaryl treatment and an increased incidence of abortion or stillbirth among rhesus monkeys (Dougherty and Coulston, 1975).

Oral doses of carbaryl at 100 or 300 mg/kg for 3 months caused decreased frequency or absence of estrus, proestrus and diestrus in rats (Vashakidze, 1965). Treated females had prolonged pregnancies, reduced litter sizes and deformed wombs. Treated males were infertile.

Treatment of male rats with 0, 7, 14 or 70 mg carbary1/kg orally for 12 months (Shtenberg and Rybakova, 1968) resulted in a dose-dependent decrease in sperm motility, edema of testicular tissue, spermatogenic epithelia destruction and reduced numbers of spermatozoa. Treated female rats had estrus cycle changes at the 14 and 70 mg/kg doses.

Carbaryl in oral doses of 8.5, 17 or 34 mg/kg daily for 5 days in male Swiss-Webster mice caused no changes in testes or prostate weights and failed to inhibit the uptake of \*H-testosterone by the mouse prostate (Thomas, 1974; Thomas et al., 1974). When \*C-carbaryl was administered, only small amounts of radioactivity were detected in prostate, seminal vesicles, testes, seminal plasma and epididymal fat.

## 5.5. CHRONIC AND SUBCHRONIC TOXICITY

Carpenter et al. (1961) reported the absence of permanent degenerative changes in rats and dogs upon chronic and subchronic oral carbaryl administration. Groups of 5 male and 5 female CF-N rats were given 0, 1500 or 2250 ppm carbaryl diets for 96 days. Upon gross and histopathological examination, the only effects noted among low-dose animals were increased kidney weights in the females. Among high-dose animals, minor diffuse cloudy swelling of the kidney tubules occurred, and females had increased kidney and body weights, while males had an increased liver/body weight ratio. Food consumption was not affected (Carpenter et al., 1961).

Groups of 20 male and 20 female CF-N rats were given 0, 50, 100, 200 or 400 ppm carbaryl in the diet for 2 years. Additional animals were killed at intervals during the experiment as described in Section 5.1. Histological and hematological evaluations revealed no increased tumor incidence, no treatment-related effect on hematocrit values and no lesions in any tissues other than slight changes in the kidney and liver of high dose rats. A significantly increased incidence (p<0.002) of cloudy swelling of the hepatic cords was observed in randomly selected animals from the high dose group only at the end of the experiment. Specific incidences were not enumer- ated. Cloudy swelling of the kidney was observed in high dose animals; the incidence was statistically significantly elevated at the end of 1 year of treatment but not at the end of 2 years of treatment (Carpenter et al., 1961).

Groups of 2 male and 2 female Basenji cocker dogs were given gelatin capsules containing 0, 0.45, 1.8 and 7.2 mg carbaryl/kg 5 days/week for 1 year. No significant treatment-related effects on blood cholinesterase levels or other hematological parameters were observed. Histopathological evaluation revealed a diffuse cloudy swelling in the kidneys of dogs only at the high dosage level. This effect was judged to be transient, as it was observed in controls, but to a lesser extent (Carpenter et al., 1961).

Shtenberg and Rybakova (1968) orally administered 0, 7, 14 or 70 mg carbaryl/kg to groups of 24 male and 24 female rats (strain not specified) for up to 12 months. Growth inhibition and decreased blood cholinesterase activity were observed in the two high dose groups, but not in the low dose group. Increased gonadotrophic hormone production in the hypophysis, increased adrenal gland activity and decreased thyroid activity in treated groups were also observed. Reproductive effects were also observed as discussed in Section 5.4.

Dikshith et al. (1976) conducted a subchronic study in which groups of 7 male albino rats received 0 or 200 mg carbaryl/kg by gavage 3 days/week for 90 days. Biochemical evaluation showed significant increases in testicular succinic dehydrogenase and adenosine triphosphatase and hepatic glucose-6phosphatase activities compared with controls. In addition, significant décreases in hepatic acid phosphatase and blood cholinesterase activities gross or were attributable to carbaryl treatment. No abnormalities were observed in the blood, liver, kidneys, testes, sperm or brains of treated rats. No adverse effects on survival or fertility were observed. (The biochemical effects reported here are based upon data and statistical significance levels provided in a table in the referenced publication. The authors, however, conclude that carbaryl produces no biochemical changes in rats, although the quantitative data reported are in contrast with their conclusions).

Makakura et al. (1978) observed ultrastructural and metabolic anomalies in the livers of male Wistar rats given 50 doses of carbaryl (3 mg/dose) by gavage over a period of 1 year. Of the rats treated, four received restricted daily diets, 3 hours of feeding followed by 21 hours of fasting. Carbaryl treatment was given to three of the rats and they were placed on a random feeding schedule. A control group of 13 rats received physiological saline by gavage. All groups were observed for 60 days after treatment, during which a random feeding schedule was followed. Determinations of blood glucose and immunoreactive insulin levels were made before and after treatment. Carbaryl treated rats had elevated blood glucose levels and slight reductions in serum immunoreactive insulin. Upon histopathological evaluation, morphological changes in the livers of treated rats were observed (i.e., virtual absence of hepatic glycogen granules and swollen granular endoplasmic reticulum).

# 5.6. OTHER RELEVANT INFORMATION

Many data are available regarding the acute toxicity of carbaryl in a variety of species. The lethal doses (or concentrations) are listed in Table 5-3 (NIOSH, 1983).

Several investigators have examined the effects of short-term occupational exposure to carbaryl. The hydrolysis product of carbaryl, 1-naphthol, has been identified in the urine of production plant workers (Best and Murray, 1962) and formulating plant workers (Durham and Wolfe, 1962). Whorton et al. (1979) evaluated the exposure of carbaryl production workers and reported the absence of correlation between carbaryl exposure and reduction of sperm count.

Leavitt et al. (1982) and Gold et al. (1982) evaluated groups of pesticide applicators involved in spraying carbaryl. Results indicated that the predominant exposure route is dermal, with the forearms and hands having the highest exposure. Acetylcholinesterase levels of exposed applicators were not significantly different from normal. No alterations in blood acetylcholinesterase activity was observed in humans given 0.06 or 0.12 mg carbaryl/kg daily for 6 weeks (Wills et al., 1967). A decrease in the urinary amino acid:creatinine nitrogen ratio, however, was observed at the higher dose level (Wills et al., 1968).

The effects of carbaryl administration (usually by i.p. injection) on certain parameters of behavior in the rat have been reported by several investigators. Operant behavior in obtaining food or water and in avoiding shock can be affected by doses as low as 5-20 mg/kg (Anger and Wilson, 1980), and activity on the running wheel can be impaired with a dose of <1 mg/kg (Singh, 1973).

TABLE 5-3
Acute Lethal Toxicity of Carbaryl\*

	Route	LD <sub>50</sub> (or LC <sub>50</sub> )
Species	Route	
Rat	oral	250 mg/kg
Rat	dermal	<b>400 mg/kg</b>
Rat	inhalation	721 mg/m³
Rat	intraperitoneal	<b>4</b> 8 mg/kg
Rat	intravenous	<b>4</b> 1.9 mg/kg
Mouse	oral	<b>438</b> mg/kg
Mouse	intraperitoneal	25 mg/kg
Rabbit	oral	710 mg/kg
Rabbit	dermal	2000 mg/kg
	oral	280 mg/kg
Guinea pig Cat	oral	150 mg/kg

<sup>\*</sup>Source: NIOSH, 1983

Subacute or acute oral carbaryl administration has been reported to inhibit brain cholinesterase in rats (Desi et al., 1974), and to reduce body temperature and plasma cholinesterase activity in mice (Ahdaya et al., 1976). Carbaryl administration has been reported to alter the immune response of rabbits (Street and Sharma, 1975) and mice (Wiltrout et al., 1978).

### 6. AQUATIC TOXICITY

The data on carbaryl toxicity in aquatic organisms are quite extensive (U.S. EPA, 1983; Mount and Oehme, 1981). Because of the limited nature of this profile, those studies dealing with fish and aquatic organisms that are not native to the United States (Bailey et al., 1970) will not be reviewed.

6.1. ACUTE

The acute toxicity of carbaryl has been tested in many species of freshwater fish (Table 6-1). The most sensitive species tested were the lake trout, <u>Salvelinus namaycush</u>, yellow perch, <u>Perca flavescens</u>, and the Coho salmon, <u>Oncorhynchus kisutch</u>, having 96-hour LC<sub>50</sub> values of 0.690, 0.745 and 0.764 mg carbaryl/1, respectively (Macek and McAllister, 1970; Johnson and Finley, 1980). The most resistant species tested, having respective 96-hour LC<sub>50</sub> values of 15.80, 20.0 and 31.80 mg carbaryl/1 were the channel catfish, <u>Ictalurus punctatus</u>, the bullhead, <u>Ictalurus melas</u>, and the mosquito fish, <u>Gambusia affinis</u>. Woodward and Mauck (1980) demonstrated the dependence of carbaryl toxicity to cutthroat trout, <u>Salmo clarkii</u>, on the pH of water. The 96-hour LC<sub>50</sub> values determined in pH 6.5, 7.5 and 8.5 water were 6.00, 3.95 and 0.97 mg carbaryl/2, respectively, in this trout species. Post and Schroeder (1971) demonstrated that small and immature fish were more sensitive than larger, more mature fish of the same species.

The acute toxicity of carbaryl to marine species of fish has been investigated less extensively (Table 6-2). The 96-hour LC<sub>50</sub> value for striped bass, Morone saxatilis, was 1.0 mg carbaryl/2 (Korn and Earnest, 1974). Stewart et al. (1967) reported 24-hour static median lethal concentrations of 3.9, 4.1 and 6.7 for the shiner perch, Cymatogaster oggregata, English sole, Parophrys vetulus, and the three spined stickleback, Gasterosteus aculeatus, respectively.

TABLE 6-1

Acute Lethal Effects of Carbaryl to Freshwater Fish in a 96-Hour Bioassay<sup>d</sup> (LC<sub>50</sub>)

Species	Mean Concentration (mg/l)	Method	Reference
Rainbow trout, Salmo gairdneri	4.34	static	Macek and McAllister, 1970
Rainbow trout, S. gairdneri	1.470	static	Post and Schroeder, 1971
Cutthroat trout, Salmo clarkii	0.970 <sup>b</sup>	static	Woodward and Mauck, 1980
Cutthroat trout, Salmo clarkii	1.500 <sup>c</sup>	static	Post and Schroeder, 1971
Brown trout, Salmo trutta	1.95	static	Macek and McAllister, 1970
Brook trout, Salvelinus fontinalis	1.070 <sup>c</sup>	static	Post and Schroeder, 1971
Lake trout, Salvelinus namaycush	0.690	static	Johnson and Finley, 1980
Coho salmon, Oncorhynchus kisutch	0.764	static	Macek and McAllister, 1970
Coho salmon, 0. kisutch	1.300	static	Post and Schroeder, 1971
Chinook salmon, Oncorhynchus tshawytscha	2.40	flow-through	Johnson and Finley, 1980
Atlantic salmon, Salmo solar	4.50	static	Johnson and Finley, 1980
Blueqill sunfish, Lepomis macrochirus	6.76	static	Macek and McAllister, 1970
Green sunfish, Lepomis cyanellus	11.2	static	Johnson and Finley, 1980
Redear sunfish, <u>Lepomis microlophus</u>	11.20	static	Macek and McAllister, 1970

.

Species	Mean Concentration (mg/l)	Method	Reference
Black crappie, Pomoxis nigromaculatus	2.60	static	Johnson and Finley, 1980
Largemouth bass, Micropterus salmoides	6.4	static	Macek and McAllister, 1970
Goldfish, Carassius auratus	13.20	static	Macek and McAllister, 1970
Carp, Cyprinus carpio	5.28	static	Macek and McAllister, 1970
Carp, C. carplo	1.7	static	Chin and Sudderruddin, 1979
Fathead minnow, Pimephales promelas	14.60	static	Macek and McAllister, 1970
Mosquito fish, Gambusia affinis	31.80	static	Chalyarach et al., 1975
Channel catfish, <u>Ictalurus punctatus</u>	15.80	static	Macek and McAllister, 1970
Channel catfish, <u>I</u> . <u>punctatus</u>	1.30	static	Brown et al., 1979
Black bullhead, <u>Ictalurus</u> melas	20.00	static	Macek and McAllister, 1970
Yellow perch, Perca flavescens	5.10	static	Johnson and Finley, 1980
Yellow perch, P. flavescens	0.745	static	Macek and McAllister, 1970

avalues for low body weight group of trout.

CToxicity data from tests of <96-hours duration have not been reported here, although they may exist in the original reference for a specific species. <sup>b</sup>Carbaryl was most toxic to trout in pH 8.5, soft water, as reported here.

TABLE 6-2

Acute Lethal Effects of Carbaryl to Marine Fish in a Static Bioassay (LC50)

Mean Concentration Reference (mg/L)	3.9 Stewart et al., 1967	ta c		24	Korn and Earnest, 1974
Species		ğ	English sole, Parophrys retulus	Stickleback, Gasterosteus aculeatus	

The acute lethal effects of carbaryl have been studied in many aquatic invertebrates including crustaceans, insects and mollusks (Table 6-3). In general, the mollusks, both freshwater and marine species, tended to be the most resistant to carbaryl, with  $LC_{50}$  values between 2.2 and 125.0 mg/%. The lethal concentrations (24-, 48- or 96-hour values) among crustaceans range from 0.00026 mg/% for the 24-hour  $LC_{50}$  in Daphnia magna (Rawash et al., 1975) to a 96-hour  $LC_{50}$  of 2.43 mg/% in the crayfish, Procarnbarus simulans (Chaiyarach et al., 1975).

#### 6.2. CHRONIC

The effects of 0.1 mg/2 carbaryl exposure for a period of 5 months was studied in the estuarine spot, <u>Leiostomus xanthurus</u> (Lowe, 1967). Although mortality was high in both control and experimental groups (65%), no carbaryl-related mortality could be determined. Histopathological examination showed no exposure-induced tissue changes. Parasitic lesions of the brain detected in treated fish may have been due to a treatment-related, lower resistance to infestation by the parasite (Lowe, 1967). Carbaryl was shown to inhibit brain acetylcholinesterase activity (68% of normal) in spot treated for 13 days with 1.0 mg/2, but only slight inhibition was noted after 2.5 and 5 months exposure at 0.1 mg/2.

The effects of carbaryl exposure to embryos of the killifish, Fundulus heteroclitus, and medaka, Oryzias latipes, have been examined (Weis and Weis, 1974; Solomon and Weis, 1979). Circulatory anomalies and cardiac malformations were associated with carbaryl exposure in embryos of both fish. In the killifish, treatment at 10 mg/2 caused a reduction in successful axis formation (at 2 days), reduced pigment formation and feeble or nondiscernible heartbeat (at 3 days) and arrested development at stage 22 or 24 (Weis and Weis, 1974). Exposure to 10 mg carbaryl/2 for 3 days caused the

TABLE 6-3

Acute Lethal Effects of Carbaryl to Aquatic Invertebrates

Ł

Species	Durationa (hours)	Mean Concentration (mg/L)	Method	Effect	Reference
	48	0.04	static	1,650	Stewart et al., 1967
Mud shrimp, Upogebla pugettensis	95	0.12	static	LC <sub>50</sub>	Chalyarach et al., 1975
Grass shrimp, Palaemonetes kadiakensis	. es	0.03	static	1050	Stewart et al., 1967
Ghost shrimp, Lalitanassa californisis	96	2.43	static	1C <sub>50</sub>	Chalyarach et al., 1975
Crayfish, Procambarus Simulans	87	0.550	flowthrough	EC50ª	Butler, 1963
Blue crab, Callinectes Sapidus	2 <del>7</del> 2	0.27	static	1°50	Stewart et al., 1967
Shore crab, Hemigrapsus oregonensis	<b>7</b>	0.60	static	1c <sub>50</sub>	Stewart et al., 1967
Dungeness crab, Cancer magister	· **	0.029	static	LC <sub>50</sub>	Bluzat and Seuge, 1979
Amphipod, Gammarus pulex		0.0025	flowthrough	ECSOB	Butler, 1963
Prawn, Penaeus aztecus	÷ ÷	0.007 <sup>b</sup>	static	ຂ້າ	Woodward and Mauck, 1980
Amphipod, Gammarus pseudolimnaeus	2	0.00026	static	05 <sub>27</sub>	Rawash et al., 1975
Water flea, Daphnia magna	; \$	0.0064	static	רנים	Johnson and Finley, 1980
Water flea, Daphnia pulex	÷ \$	0.011	static	1C <sub>50</sub>	Woodward and Mauck, 1980
Stone fly, Pteronarcella Dadia	70	0.075	static	rc <sub>50</sub>	Rawash et al., 1975
Mosquito larvae, Culex <u>pipiens</u>	. 87	2.3	static	10.50	Stewart et al., 1967
Bay mussel, Mytilus edulis	<b>.</b>	2.5	static	0537	Stewart et al., 1967
Pacific oyster, Crassostrea gigas	96	125.0	static	05 27	Chalyarach et al., 1975
Mactrid clam, Rangia cuntata	\$ <del>7</del>	7.3	static	0537	Stewart et al., 1967
Cockle clam, Cilnocardium nucleilli	<b>8</b>	21.0	static	05 ) 1	Bluzat and Seuge, 1979
Fresheater Clam, Limited Staumarts	96	8.2	static	LC50	Balley and Llu, 1980

deffect was loss of equilibrium or mortality.

boata for pH 7.5 water

CData for pH 6.5 water

development of a defective "tube heart" in up to 15% of the exposed killifish embryos. In the medaka, carbaryl at 5 mg/2 or greater caused heart abnormalities in 96% or more of the developing embryos. Once again, many anomalies were classified as undifferentiated tubular-type hearts. Circular abnormalities, edema and clotting were also prevalent in treated medaka eggs (Solomon and Weis, 1979). Developmental arrest was reported in embryos treated with 30.0 mg carbaryl/2.

When embryos of the South African clawed toad, <u>Xenopus laevis</u>, were treated with 1 mg carbaryl/2, slight edema and ventral curvature were reported (Elliott-Feeley and Armstrong, 1981). The LC<sub>50</sub> for embryos for 24 hours postgastrulation was 4.7 mg/2. In the embryos surviving at 10 mg/2, severe abnormalities were reported. Tadpoles (larvae) were treated at 0.1, 1.0 or 10.0 mg/2 for 24 hours and showed dose-related changes in swimming behavior. Ninety-six, 92 and 94% of the tadpoles treated at 0.1, 1.0 and 10.0 mg carbaryl/2, respectively, for 24 hours survived and had developed normally at 2 weeks after exposure (Elliott-Feeley and Armstrong, 1981).

# 6.3. PLANT EFFECTS

The growth and assimilation of radiolabeled sodium carbonate in the freshwater green alga, <u>Scenedesmus quadricaudata</u>, was determined after treatment with 0.1 or 1.0 mg carbaryl/2 (Stadnyk et al., 1971). Oddly, carbaryl treatment was associated with a dose-related increase in cell growth and uptake of <sup>14</sup>C, with 44 and 57% increases in the 0.1 and 1.0 treated groups, respectively, when compared with controls. Bringmann and Kuhn (1977) reported that the toxicity threshold, the concentration where cell multiplication inhibition became evident, for carbaryl in <u>S. quadri-</u>

cauda was 1.4 mg/2. O'Kelley and Deason (1976) reported similar findings in toxicity tests with 36 species of algae belonging to the genera Chlorella, Scenedesmus, Nitzschia, Golinkiniopsis, Monoraphidium, Actinastrum, Koliella and Carteria. At 0.01 and 0.1 mg carbaryl/2, algal growth was similar to controls. At 1.0 mg/2, 18 of 36 culture species showed slightly inhibited growth. Treatment at 10.0 and 25.0 mg carbaryl/2 caused some algal species to be severely inhibited (<50% of control growth), while other species showed stimulated growth in comparison with controls (O'Kelley and Deason, 1976).

The effects of carbaryl on the growth of four marine algal species was tested by Walsh and Alexander (1980). The 4- and 12-day EC<sub>50</sub> values for growth inhibition were determined photometrically for chlorophyll content in Chlorococcum sp., Chlorella sp., Skeletonema costatum and Nitzschia angularum. The values were 2.1 and 2.7 mg/l, 1.0 and 1.2 mg/l, 1.7 and 1.8 mg/l, and 1.5 and 1.6 mg/l, respectively, for the four species at 4 and 12 days (Walsh and Alexander, 1980).

### 6.4. RESIDUE

Pertinent data regarding carbaryl residues in the aquatic environment could not be located in the available literature.

# 6.5. OTHER RELEVANT INFORMATION

The uptake, metabolism and biliary excretion of carbaryl was studied in rainbow trout, Salmo gairdneri (Statham et al., 1975). Carbaryl at 0.25 mg/2 was absorbed from the water and metabolized, with ~30% of the absorbed dose appearing in the bile within 24 hours. The biliary metabolites identified were, in decreasing prevalance, unchanged carbaryl, 1-naphthol glucuronide, 5,6-dihydrodihydroxycarbaryl and an unidentified polar metabolite (Statham et al., 1975). A bile:water concentration ratio

of radiolabeled carbaryl was 947 after 24 hours of exposure at 0.25 mg/ $\Omega$  (Statham et al., 1976). In channel catfish, <u>Ictalurus punctatus</u>, carbaryl was absorbed from the water and from a treated diet. The accumulation of carbaryl or metabolite residues was 11 ng/g tissue after 56 days of exposure to 0.25 mg carbaryl/ $\Omega$  water (Korn, 1973). Accumulation from dietary carbaryl, 2.8 mg/kg/week, was 11 and 9 ng/g tissue at 3 and 56 days, respectively (Korn, 1973). Fish accumulated 1% or less of the available pesticide. Tissue residues were eliminated rapidly in dietary treated fish, while elimination was slower in those exposed to carbaryl in the water. The author (Korn, 1973) suggested that carbaryl in the water is hydrolyzed to  $\Omega$ -naphthol, which has a greater retention in fish tissues than the parent compound.

### 7. EXISTING GUIDELINES AND STANDARDS

#### 7.1. HUMAN

The ACGIH (1982) recommends a TLV of 5 mg/m $^3$  for an 8-hour TWA and a 15-minute STEL of 10 mg/m $^3$ .

NIOSH (1983) recommends a concentration limit of 5 mg/m² for a 10-hour TWA, and OSHA (1981) has established a standard for occupational carbaryl exposure of 5 mg/m³ for an 8-hour TWA.

Tolerances for carbaryl residues have been established for a variety of raw agricultural commodities. A tolerance of 0.1 ppm has been established for the meat, fat and meat by-products of cattle, goats, horses, sheep and swine. A tolerance of 1 ppm has been established for the liver and kidney of these animals.

According to the Federal Register (FR), tolerances for carbaryl residues on vegetable raw agricultural commodities intended for human consumption range from 0 ppm for barley, oat and rye grain to 12 ppm for berries, spinach and collard and mustard greens. Carbaryl tolerances for residues on raw agricultural commodities intended for livestock consumption are set at 100 ppm for hay, straw, forage and fodder of grains and other vegetables (40 CFR 180.169).

An ADI value for humans of 0.01 mg/kg has been reported by WHO (1974).

#### 7.2. AQUATIC

Guidelines and standards specifically for the protection of aquatic organisms from the toxic effects of carbaryl could not be located in the available literature. Assuming the maximum application rate of 4.6 kg/hectare, a body of water 1 meter in depth would have a concentration of 0.34 mg carbaryl/1 (Korn, 1973). Providing a body of water were sprayed directly at this application rate, toxic effects could be significant.

#### 8. RISK ASSESSMENT

The well designed and well reported chronic studies of Innes et al. (1969), Carpenter et al. (1961) and Triolo et al. (1982) have failed to demonstrate the carcinogenicity of carbaryl. An elevated number of animals with lung adenomas has been reported to occur upon i.p. administration to strain A mice, but an increased number of tumors/tumor-bearing animal must also be observed to be regarded as a positive response in this bioassay. Andrianova and Alekseev (1970) reported an increased incidence of tumors in rats given carbaryl by gavage for up to 22 months; however, the use of a compound of questionable purity (foreign carbaryl is thought to contain 2-naphthyl methylcarbamate), excessive mortality among treated animals and the lack of experimental details (no Russian translation) detract from the usefulness of this study.

Carpenter et al. (1961) reported the absence of increased tumor incidence in rats fed 50-400 ppm of carbaryl in the diet for 2 years, and in rats given weekly s.c. injections of 10 mg carbaryl for 5 months. Cloudy swelling of the hepatic cords and a transient increased incidence of cloudy swelling of the renal tubules was observed in rats fed 400 ppm, but not 200 ppm.

Russian investigators have reported adverse reproductive effects at a dosage as low as 2 mg/kg (Shtenberg and Ozhavan, 1971; Shtenberg and Rybakova, 1968), but the test compound was of unknown purity. No adverse reproductive effects were reported in the 3-generation study of Weil et al. (1973) wherein 25 mg carbaryl/kg/day of known composition was administered by gavage to rats. Carbaryl was teratogenic in some species when administered at relatively high dosages, and there is some evidence of teratogenicity in beagle dogs and miniature swine at lower dosages (Section 5.3.).

Earl et al. (1973) reported a single incidence of terata in fetuses of swine at each of the 8 and 16 mg carbaryl/kg/day levels. In a follow-up experiment, however, Earl et al. (1973) did not observe any teratogenic effects at dose levels of 16 or 32 mg carbaryl/kg/day. It is not possible to draw definitive conclusions from this study for use in risk assessment because of the small numbers of animals tested and because the lack of teratogenic effects observed in the follow-up experiment may or may not have been due to decomposition of the carbaryl, which had been stored for 12-15 months before use. The teratogenicity study with beagle dogs (Smalley et al., 1968; Earl et al., 1973) is also not appropriate for human risk assessment, as dogs may not be a suitable experimental model for humans, because of marked differences between dogs and other mammals (including humans) in the metabolism of carbaryl (see Section 4.3.) (NIOSH, 1976). Also, the U.S. EPA, in an evaluation of the teratogenicity of carbaryl published in the Federal Register (1980), concluded that the teratogenicity study with beagle dogs was not adequate in terms of numbers of animals and observation of the dams during treatment. Studies with swine were mentioned but not discussed. The Agency concluded that carbaryl "would not constitute a potential human teratogenic or reproductive hazard under proper environmental usage."

The data of Carpenter et al. (1961) will be used to derive an ADI because a compound of known purity was administered to a sufficient number of animals (20 of each sex) via an environmentally relevant route (diet) for a significant portion of the rat's lifespan (2 years). The highest reported dosage that produced no adverse effects was 9.6 mg/kg/day (based on observed food intake) in female rats weighing ~0.3 kg. An ADI of 0.096 mg/kg/day is obtained by dividing 9.6 mg/kg/day by an uncertainty factor of 100 (10 to extrapolate from animals to humans and 10 to protect the most sensitive individuals). This ADI is equivalent to 6.72 mg/day for a 70 kg man.

## 9. REFERENCES

ACGIH (American Conference of Governmental Industrial Hygienists). 1982. TLVs, Threshold Limit Values for Chemical Substances in Work Air, adopted by ACGIH for 1982. Cincinnati, OH.

Ahdaya, S.M., P.V. Shah and F.E. Guthrie. 1976. Thermoregulation in mice treated with parathion, carbaryl or DDT. Toxicol. Appl. Pharmacol. 35(3): 575-580.

Ahmed, F.E., N.J. Lewis and R.W. Hart. 1977a. Pesticide induced ouabain resistant mutants in Chinese hamster V79 cells. Chem.-Biol. Interact. 19(3): 369-374.

Ahmed, F.E., R.W. Hart and N.J. Lewis. 1977b. Pesticide induced DNA damage and its repair in cultured human cells. Mutat. Res. 42(2): 161-173.

Aly, O.M. and M.A. El-Dib. 1971. Persistence of some carbamate insecticides in the aquatic environment. I. Hydrolysis of sevin, baygon, pyrolan and dimetilan in waters. Water Res. 5(12): 1191-1205.

Aly, O.M. and M.A. El-Dib. 1972. Persistence of some carbamate insecticides in the aquatic environment. Advan. Chem. Ser. 111: 210-243.

Aly, M.I., N. Bakry, F. Kishk and A.H. El-Sebae. 1980. Carbaryl adsorption on calcium-bentonite and soils. Soil Sci. Soc. Am. J. 44(6): 1213-1215.

•

Amer, S. 1965. Cytological effects of pesticides. I. Mitotic effects of N-methyl-l-naphthyl carbamate (Sevin). Cytologia. 30: 175-181.

Amer, S.M., and O.R. Farah. 1968. Cytological effects of pesticides. III. Meiotic effects of N-methyl-1-naphthyl carbamate (Sevin). Cytologia. 33: 337-344.

Amer, S.M., M.A. Hammouda, and O.R. Farah. 1971. Cytological and morphological effects of the insecticide N-methyl-1-naphthyl-carbamate (Sevin). Flora (Jena). 160: 433-439.

Andrianova, M.M. and I.V. Alekseev. 1970. On the carcinogenic properties of the pesticides sevin, maneb, ciram and cineb. Vap. Pitan. 29: 71. (Cited in IARC, 1976)

Anger, W.K. and S.M. Wilson. 1980. Effects of carbaryl on variable interval response rates in rats. Neurobehav. Toxicol. 2(1): 21-24.

Argauer, R.J. and J.D. Warthen, Jr. 1975. Separation of 1- and 2-naphthols and determination of trace amounts of 2-naphthyl methylcarbamate in carbaryl formulations by high pressure liquid chromatography with confirmation by spectrofluorometry. Anal. Chem. 47: 2472. (Cited in NIOSH, 1976)

Ashwood-Smith, M.J., J. Trevino, and R. Ring. 1972. Mutagenicity of Dichlorvos. Nature. 240: 418-420.

Bailey, H.C. and D.H.W. Liu. 1980. <u>Lumbriculus variegatus</u>, a benthic oligochaete, as a bioassay organism. ASTM STP 707. p. 205-215.

Bailey, R.M., J.E. Fitch, E.S. Herald, et al. 1970. A List of Common and Scientific Names of Fishes from the United States and Canada. Am. Fish. Soc. Spec. Pub. No. 6, Washington, DC.

Bend, J.R., G.M. Holder, E. Protos and A.J. Ryan. 1971. Water-soluble metabolites of carbaryl (1-naphthyl N-methylcarbamate) in mouse-liver preparations and in the rat. Austr. J. Biol. Sci. 24: 535-546. (Cited in IARC, 1976)

Benson, B.W., W.J. Scott and R.P. Beliles. 1967. Sevin -- Safety evaluation by teratological study in the mouse. Unpublished report submitted to Union Carbide Corp. 19 p. (Cited in NIOSH, 1976)

Benson, W.H. and H.W. Dorough. 1979. Comparative carbamate ester hydrolysis in four mammalian species. Toxicol. Appl. Pharmacol. 48(1): A139.

Beraud, M., B. Pipy, R. Derache and D. Gaillard. 1979. Formation of the carcinogen, N-nitrosocarbaryl, by interactions between an insecticide of the carbamate, carbaryl series and sodium nitrate in rat gastric juice. Food Cosmet. Toxicol. 17(6): 579-583. (French with English Abstract)

Best, E.M., Jr. and B.L. Murray. 1962. Observations on workers exposed to Sevin insecticide - A preliminary report. J. Occup. Med. 4: 507-517. (Cited in NIOSH, 1976; Leavitt et al., 1982)

Blevins, R.D., M. Lee and J.D. Regan. 1977. Mutagenicity screening of five methyl carbamate insecticides and their nitroso derivatives using mutants of <u>Salmonella</u> typhimurium LT2. Mutat. Res. 56(1): 1-6.

Bluzat, R. and J. Seuge. 1979. Effects of 3 insecticides (lindane, fenthion and carbaryl): Acute toxicity on 4 species of limnic invertebrates; chronic toxicity in the pulmonate mollusk <u>Lymnaea</u>. Environ. Pollut. 18(1): 51-70.

Bollag, J-M. and S-Y. Liu. 1971a. Degradation of sevin by soil micro-organisms. Soil Biol. Biochem. 3(4): 337-345.

Bollag, J-M. and S-Y. Liu. 1971b. Metabolism of carbaryl by a soil fungus. J. Agric. Food Chem. 19(3): 487-490.

Bollag, J-M., E.J. Czaplicki and R.D. Minard. 1975. Bacterial metabolism of 1-naphthol. J. Agric. Food Chem. 23: 85. (Cited in Mount and Oehme, 1981)

Bond, D.J. and A.C. Chandley. 1983. Aneuploidy. Oxford Univ. Press, Oxford, England.

Boush, G.M. and F. Matsumura. 1967. Insecticidal degradation by <u>Pseudo-monas meolphthora</u>, the bacterial symbiote of the apple maggot. J. Econ. Entomol. 60: 918-920.

Brankovan, O. 1972. The meiotic effect of Sevin 50 after treatment of corn in the embryonic and generative phases of development. Arhiv za Poljopriv-redne Nauke. 25:(90) 128-138.

Bringmann, V.G. and R. Kuhn. 1977. Grenzwerte der Schadwirkung wassergefahrdender Stoffe gegen Bakterien (<u>Pseudomonas putida</u>) und Grunalgen
(<u>Scenedesmus quadricauda</u>) im Zellvermehrungshemmtest. Z. f. Wasser und
Abwasser-Forschung. 10: Jahrgang. Nr. 3/4/77, 87-98.

Brown, K.M., D.C. Anderson, S.G. Jones, L.E. Deuel and J.D. Price. 1979. The relative toxicity of four pesticides in tap water and water from flooded rice paddies. Int. J. Environ. Stud. 14(1): 49-54.

Brzeskij, V.V. and V.I. Vaskov. 1972. Mutation studies and fertility calculation with <u>Drosophila melanogaster</u> under the effects of carbaryl. Agnew. Parasitol. 13: 23-28.

Bukin, A.L. and G.V. Filatov. 1965. Sevin toxicity for mammals and birds. Veterinariia. 42: 93-95. (Rus.) (Cited in NIOSH, 1976)

Butler, D.A. 1963. A Review of Fish and Wildlife Service Investigations During 1961 and 1962. Circ. 167, Fish Wildl. Serv., Washington, DC. 25 p. (Cited in U.S. EPA, 1983)

Cambon, C., C. Declume and R. Derache. 1978. Inhibition of acetylcholinesterase from fetal and maternal tissues after oral intake of carbaryl (1-naphthyl-N-methyl-carbamate) by pregnant rats. Biochem. Pharmacol. 27(22): 2647-2648.

Cambon, C., C. Declume and R. Derache. 1980. Fetal and maternal rat brain acetylcholinesterase: Isoenzymes changes following insecticidal carbamate derivative poisoning. Arch. Toxicol. 45(4): 257-262.

Cambon, C., Y. Fernandez. M. Falzon and S. Mitjavila. 1981. Variations of the digestive absorption kinetics of carbaryl with the nature of the vehicle. Toxicology. 22(1): 45-51.

Caro, J.H., H.P. Freeman and B.C. Turner. 1974. Persistence in soil and losses in runoff of soil-incorporated carbaryl in a small watershed. J. Agric. Food Chem. 22(5): 860-863.

Carpenter, C.P. and C.S. Weil. 1970. Cited as a personal communication.

J. Agric. Food Chem. 18: 1015. (Cited in Mount and Oehme, 1981)

Carpenter, C.P., C.W. Weil, P.E. Palm, et al. 1961. Mammalian toxicity of 1-naphthyl-N-methylcarbamate (Sevin insecticide). J. Agric. Food Chem. 9: 30-39.

Casper, H.H., J.C. Pekas and W.E. Dinusson. 1973. Gastric absorption of a pesticide (1-naphthyl-N-methylcarbamate) in the fasted rat. Pestic. Biochem. Physiol. 2: 391-396. (Cited in NIOSH, 1976)

Chaiyarach, S., V. Ratananun and R.C. Harrel. 1975. Acute toxicity of the insecticides toxaphene and carbaryl and the herbicides propanil and molinate to four species of aquatic organisms. Bull. Environ. Contam. Toxicol. 14(3): 281-284.

Chen, K.C. and H.W. Dorough. 1979. Glutathione and mercapturic acid conjugates in the metabolism of naphthalene and 1-naphthyl-N-methylcarbamate (carbaryl). Fed. Proc. 38: 585. (Cited in Mount and Oehme, 1981)

Chin, Y.N. and K.I. Sudderuddin. 1979. Effect of methamidophos on the growth rate and esterase activity of the common carp <u>Cyprinus</u> carpio L. Environ. Pollut. 18(3): 213-220.

Chin, B.H., J.M. Eldridge and L.J. Sullivan. 1974. Metabolism of carbaryl by selected human tissues using an organ-maintenance technique. Clin. Toxicol. 7(1): 37-56.

Collins, T.F.X., W.H. Hansen and H.V. Keeler. 1971. The effect of carbaryl (Sevin) on reproduction of the rat and the gerbil. Toxicol. Appl. Pharmacol. 19: 202-216.

Coulston, F. 1971. The effect of carbaryl on reproduction in the rhesus monkey. Albany Med. College, Inst. Exp. Pathol. Toxicol., Albany, NY. 18 p. (Cited in NIOSH, 1976)

Declume, C. and P. Benard. 1977. Autoradiographic study of the distribution of an anticholinesterase agent, 1-naphthyl-N-methyl-14C-carbamate in the pregnant rat. Toxicol. Appl. Pharmacol. 39(3): 451-460. (Fre.)

Declume, C., C. Cambon and R. Derache. 1979. The effects on newborn rats of repeated carbaryl administration during gestation. Toxicol. Lett. 3(4): 191-196. (Abstract)

DeGiovanni-Donnelly, R., S.M. Kolbye, and P.D. Greeves. 1968. The effects of IPC, CIPC, Sevin, and Zectran on <u>Bacillus subtilis</u>. Experientia. 24: 80-81.

Degraeve, N., M. Moutschen-Dahmen, N. Houbrechts, and A. Colizzi. 1976. The hazards of an insecticide carbaryl used alone and in association with nitrites. Bulletin dela Societe Royale des Sciences de Liege. 45: 46-57.

DeLorenzo, F., N. Staiano, L. Silengo and R. Cortese. 1978. Mutagenicity of diallate, sulfallate and triallate and relationship between structure and mutagenic effects of carbamates used widely in agriculture. Cancer Res. 38(1): 13-15.

Desi, I., L. Gonczi, G. Simon, I. Farkas and Z. Kneffel. 1974. Neurotoxi-cologic studies of two carbamate pesticides in subacute animal experiments. Toxicol. Appl. Pharmacol. 27(3): 465-476.

Dikshith, T.S.S., P.K. Gupta, J.S. Gaur, K.K. Datta and A.K. Mathur. 1976.
Ninety day toxicity of carbaryl in male rats. Environ. Res. 12: 161-170.

Dorough, H.W. 1967. Carbaryl-C14 metabolism in a lactating cow. J. Agric. Food Chem. 15: 261. (Cited in Mount and Oehme, 1981)

Dougherty, W.J. and F. Coulston. 1975. Teratogenic evaluation of carbaryl in the rhesus monkey (Macaca mulatta). Unpublished report submitted to Union Carbide Corp., June 6. (Cited in NIOSH, 1976)

Duggan, R.E., G.Q. Lipscomb, E.L. Cox, R.E. Heatwole and R.C. Kling. 1971. Pesticide residue levels in food in the U.S. from July 1, 1963 to June 30, 1969. Pestic. Monit. J. 5: 73. (Cited in Mount and Oehme, 1981.)

Darham, W.F. and H.R. Wolfe. 1962. Measurement of the exposure of workers to pesticides. Bull. WHO. 26: 75. (Cited in Leavitt et al., 1982)

Earl, F.L., E. Miller and E.J. Van Loon. 1973. Reproductive, teratogenic and neonatal effects of some pesticides and related compounds in beagle dogs and miniature swine. Pestic. Environ.: Continuing Controversy, Paper Inter-Am., Conf. Toxicol. Occup. Med. p. 253-266.

Egert, G., and H. Greim. 1976. Formation of mutagenic N-nitroso compounds from the pesticides prometryne, dodine, and carbaryl in the presence of nitrite at pH 1. Mutat. Res. 37: 179-186.

Eichelberger, J.W. and J.J. Lichtenberg. 1971. Persistence of pesticides in river water. Environ. Sci. Technol. 5(6): 541-544.

Eisenbrand G., O. Ungerer and R. Preussmann. 1975. Reaction of nitrite with pesticides. II. Formation, chemical properties and carcinogenic activity of the N-nitroso derivative of N-methyl-l-naphthyl carbamate (carbaryl). Food Cosmet. Toxicol. 13(3): 365-367.

Elespuru, R.K. and W. Lijinsky. 1973. Formation of carcinogenic nitroso compounds from nitrite and some types of agricultural chemicals. Food Cosmet. Toxicol. 41(5): 807-817.

Elespuru, R., W. Lijinsky and J.K. Setlow. 1974. Nitrosocarbaryl, as a mutagen of environmental significance. Nature. 247: 385-387.

Elliott-Feeley, E. and J.B. Armstrong. 1981. Effects of fenitrothion and carbaryl on <u>Xenopus laevis</u> development. Toxicology. 22(4): 319-335.

Epstein, S.S., E. Arnold, J. Andrea, W. Bass and Y. Bishop. 1972. Detection of chemical mutagens by the dominant lethal assay in the mouse. Toxicol. Appl. Pharmacol. 23: 288-325.

Fahrig, R. 1974. Comparative mutagenicity studies with pesticides. IARC Sci. Publ. 10: 161-181.

Farago, A. 1969. Fatal suicidal case of Sevin (1-naphthyl-N-methylcarba-mate) poisoning. Arch. Toxicol. (Berl.) 24: 309-315. (Ger.) (Cited in NIOSH, 1976)

Federal Register. 1980. Determination not to initiate a rebuttable presumption against registration (RPAR) of pesticide products containing carbaryl; availability of decision document. Vol. 45, No. 241, p. 81869-81876.

Feldmann, R.J. and H.I. Maibach. 1974. Percutaneous penetration of some pesticides and herbicides in man. Toxicol. Appl. Pharmacol. 28(1): 126-132.

Felley, D.R. 1971. Diss. Abstr. B 31: 6040. (Cited in Caro et al., 1974)

Fernandez, Y., M. Falzon, C. Cambon-Gros and S. Mitjavila. 1982. Carbaryl tricompartmental toxicokinetics and anticholinesterase activity. Toxicol. Lett. 13(3-4): 253-258.

Fiscor, G., and G.M. Nil Lo Piccolo. 1972. Survey of pesticides for mutagenicity by the bacterial plate assay method. EMS News. 6: 6-8.

Freitag, D., H. Geyer, W. Klein, A.G. Kraus, E. Lahaniatis and F. Korte. 1979. An approach for comparative screening of the environmental behavior of chemical. Ecotox. Environ. Safety. 3: 144-151.

Gold, R.E., J.R. Leavitt, T. Holcslaw and D. Tupy. 1982. Exposure of urban applicators to carbaryl. Arch. Environ. Contam. Toxicol. 11(1): 63-67.

Guthrie, R.K., V.A. Anugwelem and E.M. Davis. 1981. Responses of bacteria to the presence of carbaryl in water. II. Pure culture vs. mixed culture response. Water Res. Bull. 17: 1005-1007.

Hansch, C. and A.J. Leo. 1979. Substituent Constants for Correlation Analysis in Chemistry and Biology. John Wiley and Sons, New York.

Hook, E.B. 1983. Perspective in mutation epidemiology: Contribution of chromosome abnormalities to human morbidity and mortality and some comments upon surveillance of chromosome mutation rates. Mutat. Res. 114: 389-423.

Houston, J.B., D.G. Upshall and J.W. Bridges. 1974. Pharmacokinetics and metabolism of two carbamate insecticides, carbaryl and landrin, in the rat. Xenobiotica. 5(10): 637-648.

Houston, J.B., D.G. Upshall and J.W. Bridges. 1975. Studies using carbamate esters as model compounds to investigate the role of lipophilicity in the gastrointestinal absorption of foreign compounds. J. Pharmacol. Exp. Ther. 195(1): 67-72.

Húrwood, I.S. 1967. Studies on pesticide residues. 2. Carbaryl residues in the body tissues and milk of cattle following dermal application. Queens J. Agric. Anim. Sci. 24: 69-74. (Cited in NIOSH, 1976)

Hwang, S.W. and L.S. Schanker. 1974. Absorption of carbaryl from the lung and small intestine of the rat. Environ. Res. 7(2): 206-211.

IARC (International Agency for Research on Cancer). 1976. Carbaryl. <u>In:</u>
Some Carbamates, Thiocarbamates and Carbazides. IARC Monographs on the Evaluation of the Carcinogenic Risk of Chemicals to Man, Vol. 12. IARC, WHO, Lyon, France. p. 37-54.

Innes, J.R.M., B.M. Ulland, M.G. Valerio, et al. 1969. Bioassay of pesticides and industrial chemicals for tumorigenicity in mice: A preliminary note. J. Natl. Cancer Inst. 42: 1101-1114.

Ishidate, M., Jr., and S. Odashima. 1977. Chromosome tests with 134 compounds on Chinese hamster cells in vitro - A screening for chemical carcinogens. Mutat. Res. 48: 337-354.

Jaszczuk, E., T. Syrowatka and J. Cybulski. 1979. Mutagenic activity of propoxur, carbaryl and their nitroso derivatives: Induction of reversion in Salmonella typhimurium. Rocz. Pansw. Zakl. Hig. 30(1): 81-88. (Pol.)

4

Johnson, W.W. and M.T. Finley. 1980. Handbook of Acute Toxicity of Chemicals to Fish and Aquatic Invertebrates. U.S. Dept. Interior, Fish Wildl. Serv., Resource Pub. 137, Washington, DC. 98 p.

Kånazawa, J. 1975. Uptake and excretion of organophosphorus and carbamate insecticides by freshwater fish, motsugo, <u>Pseudorasbora parva</u>. Bull. Environ. Contam. Toxicol. 14: 346-352.

Karickhoff, S.W. 1981. Semi-empirical estimation of sorption of hydrophobic pollutants on natural sediments and soils. Chemosphere. 10: 833-846.

Karinen, J.F., J.G. Lamberton, N.E. Stewart and L.C. Terriere. 1967. Persistence of carbaryl in the marine estuarine environment. Chemical and biological stability in aquarium systems. J. Agric. Food Chem. 15(1): 148-156.

Kazano, H., P.C. Kearney and D.D. Kaufman. 1972. Metabolism of methylcar-bamate insecticides in soils. J. Agric. Food Chem. 20(5): 975-979.

Kazarnovskaya, M.L., and A.F. Vasilos. 1977. The effect of Sevin on the chromosomal appartus of cells in vitro. Zdravaookhranenie. 20(4): 14-16.

Kenaga, E.E. and C.A.I. Goring. 1980. Relationship between water solubility, soil sorption, octanol-water partitioning and concentration of chemicals in biota. <u>In</u>: Aquatic Toxicology, Proc. 3rd Ann. Symp. Aquatic Toxicol., ASTM, Philadelphia, PA. p. 79-115.

Kitagawa, K., M. Wakakura, and S. Ishikawa. 1977. Light microscopic study of endocrine organs of rats treated by carbamate pesticide. J. Tox. Sci. 2: 53-60.

Kmaak, J.B. and L.J. Sullivan. 1967. Metabolism of carbaryl in the dog. J. Agric. Food Chem. 15: 1125. (Cited in NIOSH, 1976; Mount and Oehme, 1981)

Knaak, J.B., M.J. Tallant, W.J. Bartley and L.F. Sullivan. 1965. The metabolism of carbaryl in the rat, guinea pig and man. J. Agric. Food Chem. 13: 537. (Cited in NIOSH, 1976; Mount and Oehme, 1981)

Knaak, J.B., M.J. Tallant, S.J. Kozbelt and L.J. Sullivan. 1968. The metabolism of carbaryl in man, monkey, pig and sheep. J. Agric. Food Chem. 16(3): 465-470.

Korn, S. 1973. Uptake and persistence of carbaryl in channel catfish. Trans. Am. Fish. Soc. 102(1): 137-139.

Korn, S. and R. Earnest. 1974. Acute toxicity of twenty insecticides to striped bass, Morone saxatilis. Calif. Fish Game. 60(3): 128-131.

Krishna, J.G. and J.E. Casida. 1966. Fate in rats of the radiocarbon from ten variously labeled methyl- and dimethylcarbamate-14C insecticide chemicals and their hydrolysis products. J. Agric. Food Chem. 14: 98. (Cited in Mount and Oehme, 1981)

Krylova, T.V., and A.V. Denisova. 1973. Action of Sevin on the Mongolian tree creeper. Biol. Nauki. 16(10): 25-30.

Kuhr, R.J. and H.W. Dorough. 1976. Carbamated Insecticides: Chemistry, Biochemistry and Toxicology. CRC Press, Cleveland, OH. p. 103-124, 146, 258-286. (Cited in Mount and Oehme, 1981)

LaFleur, K.S. 1976. Carbaryl desorption and movement in soil columns. Soil Sci. 121: 212-216.

Leavitt, J.R., R.E. Gold, T. Holcslaw and D. Tupy. 1982. Exposure of professional pesticide applicators to carbaryl. Arch. Environ. Contam. Toxicol. 11(1): 57-62.

Lijinsky, W. and H.W. Taylor. 1977. Transplacental chronic toxicity test of carbaryl with nitrite in rats. Food Cosmet. Toxicol. 15(3): 229-232.

Lin, T. and H.W. Dorough. 1974. Influence of insecticide exposure on the <a href="mailto:in\_vivo">in\_vivo</a> and <a href="mailto:in\_vivo">in\_vitro</a> metabolic activity of rats. Arch. Environ. Contam. Toxicol. 2(4): 364-377.

Liu, D., K. Thomson and W.M.J. Strachan. 1981. Biodegradation of carbaryl in simulated aquatic environment. Bull. Environ. Contam. Toxicol. 27(3): 412-417.

Lopez, C. 1970. Therapy with corticosteroid compounds in acute poisoning with carbaryl and organophosphate pesticides. Int. Arch. Arbeitsmed. 26: 50-62. (Cited in NIOSH, 1976)

Lowe, J.I. 1967. Effects of prolonged exposure to Sevin on an estuarine fish, <u>Leiostomus</u> <u>xanthurus</u> <u>La cepede</u>. Bull. Environ. Contam. Toxicol. 2(3): 147-155.

Macek, K.J. and W.A. McAllister. 1970. Insecticide susceptibility of some common fish family representatives. Trans. Am. Fish. Soc. 99(1): 20-27.

Maibach, H.I., R.J. Feldmann, T.H. Milby and W.F. Serat. 1971. Regional variations in percutaneous penetration in man. Pestic. Arch. Environ. Health. 23: 208-211. (Cited in NIOSH, 1976)

Marshall, T.C. and H.W. Dorough. 1979. Biliary excretion of carbamate insecticides in the rat. Pestic. Biochem. Physiol. 11: 56. (Cited in Mount and Oehme, 1981)

Marshall, T.C., H.W. Dorough and H.E. Swim. 1976. Screening of pesticides for mutagenic potential using <u>Salmonella typhimurium</u> mutants. J. Agric. Food Chem. 24(3): 560-563.

Martin, H. and C.R. Worthington, Ed. 1977. Pesticide Manual, British Crop Protection Council, 5th ed. 593 p.

Matsumura, F. and G.M. Boush. 1968. Degradation of insecticides by a soil fungus, <u>Trichoderma viride</u>. J. Econ. Entomol. 61(3): 610-612.

McCann, J., E. Choi, E. Yamaski, and B.N. Ames. 1975. Detection of carcinogens as mutagens in the <u>Salmonella/microsome</u> test: Assay of 300 chemicals. Proc. Natl. Acad. Sci. 72(12): 5135-5139.

Mehendale, H.M. and H.W. Dorough. 1971. Glucuronidation mechanisms in the rat and their significance in the metabolism of insecticides. Pestic. Biochem. Physiol. 1(3-4): 307-318.

Ménzie, C.M. 1969. Metabolism of Pesticides. U.S. Dept. Interior, Bur. Sport Fish., Wildl. Spec. Sci. Rep., Wildl. No. 127. (Cited in Mount and Oehme, 1981)

Morpurgo, G., D. Bellincampi, G. Gualandi, L. Baldinelli and C.O. Serlupi. 1979. Analysis of mitotic nondisjunction with <u>Aspergillus</u> <u>nidulans</u>. Environ. Health Perspect. 31: 81-95.

Mount, M.E. and F.W. Oehme. 1981. Carbaryl: A literature review. Residue Rev. 80: 1-64.

Mount, M.E., A.D. Dayton and F.W. Oehme. 1981. Carbaryl residues in tissues and cholinesterase activities in brain and blood of rats receiving carbaryl. Toxicol. Appl. Pharmacol. 58(2): 282-296.

Murray, F.J., R.E. Staples and B.A. Schwetz. 1979. Teratogenic potential of carbaryl given to rabbits and mice by gavage or by dietary inclusion. Toxicol. Appl. Pharmacol. 51(1): 81-89.

Nagy, Z., I. Mile and F. Antoni. 1975. The mutagenic effect of pesticides on <u>Escherichia coli</u> WP2 try. Acta Microbiol. Acad. Sci. Hung. 22: 309-314.

NIOSH (National Institute for Occupational Safety and Health). 1976. Criteria for a Recommended Standard...Occupational Exposure to Carbaryl. U.S. DHEW Pub. 77-107-192.

NIOSH (National Institute for Occupational Safety and Health). 1983. RTECS (Registry of Toxic Effects of Chemical Substances). February, 1983.

Odeyemi, O. 1982. Relative stability of carbaryl in tropical ecosystems. Environ. Pollut., Ser. A. 29(3): 207-213.

O'Kelley, J.C. and T.R. Deason. 1976. Degradation of pesticides by algae. ORD, U.S. EPA. EPA 600/3-76-022. 41 p.

Onfelt, A., and I. Klasterska. 1983. Spindle disturbances in mammalian cells. II. Induction of viable aneuploid/polyploid cells and multiple chromatid exchanges after treatment of V79 Chinese hamster cells with carbaryl. Mutat. Res. 119: 319-330.

OSHA (Occupational Safety and Health Administration). 1981. General Industry Standards. OSHA Safety and Health Standards reprinted from 29 CFR 1910. U.S. Dept. Labor. p. 631-637.

Pekas, J.C. 1974. Absorption of pesticidal carbamates from perfused intestinal loops in conscious swine. Food Cosmet. Toxicol. 12(3): 377-379.

Pekas, J.C. 1979. Further metabolism of naphthyl N-methylcarbamate (carbaryl) by the intestine. Pestic. Biochem. Physiol. 11(1): 166-175.

Pipy, B., D. Gaillard, M. Beraud and R. Derache. 1980. Phagocytic activity of the rat reticuloendothial system and the pharmacokinetics of an anti-cholinesterasic insecticide: Carbaryl. Xenobiotica. 10(10): 785-793.

Ptpy, B., D. Gaillard and R. Derache. 1981. Phagocytic activity of the reticulo-endothelial system in the rat and rates of <u>in vivo</u> excretion of metabolites of carbaryl and <u>in vitro</u> microsomal metabolism. Biochem. Pharmacol. 30(6): 669-672.

Post, G. and T.R. Schroeder. 1971. Toxicity of four insecticides to four salmonid species. Bull. Environ. Contam. Toxicol. 6(2): 144-156.

Probst, G.S., R.E. McMahon, L.E. Hill, C.Z. Thompson, J.K. Epp and S.B. Neal. 1981. Chemically-induced unscheduled DNA synthesis in primary rat hepatocyte cultures: A comparison with bacterial mutagenicity using 218 compounds. Environ. Mutagen. 3(1): 11-32.

Rani, M.V.U., O.S. Reddi and P.P. Reddy. 1980. Mutagenicity studies involving aldrin, endosulfan, dimethoate, phosphamidon, carbaryl and ceresan. Bull. Environ. Contam. Toxicol. 25(2): 277-282.

Rawash, I.A., I.A. Gaaboub, F.M. El-Gayar and A.Y. Al-Shazli. 1975. Standard curves for nuvacron, malathion, Sevin, DDT and Kelthane tested against the mosquito <u>Culex pipiens</u> L. and the microcrustacean <u>Daphnia magna</u> straus. Toxicology. 4(2): 133-144.

Regan, J.D., R.B. Setlow, A.A. Francis and W. Lijinsky. 1976. Nitroso-carbaryl: Its effect on human DNA. Mutat. Res. 38: 293-302.

Rickard, R.W. and H.W. Dorough. 1979. Formation and fate of nitrosamides in animals. Toxicol. Appl. Pharmacol. 48(1): A138.

Robens, J.F. 1969. Teratologic studies of carbaryl, diazinon, norea, disulfiram and thiram in small laboratory animals. Toxicol. Appl. Pharmacol. 15(1): 152-163.

Rocchi, P., P. Perocco, W. Alberghini, A. Fini and G. Prodi. 1980. Effect of pesticides on scheduled and unscheduled DNA synthesis of rat thymocytes and human lymphocytes. Arch. Toxicol. 45(2): 101-108.

Rodriguez, L.D. and H.W. Dorough. 1977. Degradation of carbaryl by soil microorganisms. Arch. Environ. Contam. Toxicol. 6(1): 47-56.

Ryan, A.J. 1971. The metabolism of pesticidal carbamates. CRC Crit. Rev. Toxicol. 1(1): 33-54.

Sabharwal, P.S. and J.M. Lockard. 1979. Induction of sister chromatid exchange and polyploidy by carbaryl in V79 cells. <u>In Vitro</u>. 15(3): 172-173.

Sanborn, J.R., R.M. Francis and R.L. Metcalf. 1977. The degradation of selected pesticides in soil: A review of published literature. U.S. EPA, Cincinnati, OH. EPA 600/9-77022. 616 p.

Shah, P.V. and F.E. Guthrie. 1977. Dermal absorption, distribution and the fate of six pesticides in the rabbit. <u>In</u>: Pesticides Management and Insecticide Resistance, D.L. Watson and A.W.A. Brown, Ed. Academic Press, NY. p. 547-554. (Cited in Mount and Oehme, 1981)

Sharom, M.S., J.R. Miles, C.R. Harris and F.L. McEwen. 1980. Persistence of 12 insecticides in water. Water Res. 14(8): 1089-1093.

Shimkin, M.B., R. Wieder, M. McDonough, et al. 1969. Lung tumor response in strain A mice as a quantitative bioassay of carcinogenic activity of some carbamates and aziridines. Cancer Res. 29: 2184. (Cited in IARC, 1976)

Shirasu, Y., M. Moriya, K. Kato, A. Furuhashi and T. Kada. 1976. Mutagenicity screening of pesticides in the microbial system. Mutat. Res. 40: 19-30.

Shpirt, M.B. 1975. Toxicological evaluation of dichlorodipheny-trichloroethane (DDT), hexachlorocyclohexane (HCCH), tetramethyl-thiuramdisulfide (TMTD), Sevin, and Zineb when acting on a human cell structure. Gig. Tr. Prof. Zabol. 17:(3) 32-34.

Shtenberg, A.I. and M.V. Ozhovan. 1971. The effect of low Sevin doses on the reproductive function of animals in a number of generations. Vopr. Pitan. 30: 42. (Cited in NIOSH, 1976)

Shtenberg, A.I. and M.N. Rybakova. 1968. Effect of carbaryl on the neuro-endocrine system of rats. Food Cosmet. Toxicol. 6: 461-467. (Cited in NIOSH, 1976)

Siebert, D. and G. Eisenbrand. 1974. Induction of mitotic gene conversion in <u>Saccharomyces</u> <u>cerevisiae</u> by N-nitrosated pesticides. Mutat. Res. 22: 121-126. (Cited in NIOSH, 1976)

Sikka, H.C., S. Miyazaki and R.S. Lynch. 1975. Degradation of carbaryl and linaphthol by marine microorganisms. Bull. Environ. Contam. Toxicol. 13(6): 666-672.

Singh, J.M. 1973. Decreased performance behavior with carbaryl, an indication of clinical toxicity. Clin. Toxicol. 6(1): 97-108.

Smalley, H.E., J.M. Curtis and F.L. Earl. 1968. Teratogenic action of carbaryl in beagle dogs. Toxicol. Appl. Pharmacol. 13(3): 392-403.

Solomon, H.M. and J.S. Weis. 1979. Abnormal circulatory development in medaka caused by the insecticides carbaryl, malathion and parathion. Teratology. 19(1): 51-62.

SRI (Stanford Research Institute). 1983. Directory of Chemical Producers, U.S.A. SRI, Menlo Park, CA.

Stadnyk, L., R.S. Campbell and B.T. Johnson. 1971. Pesticide effect on growth and 14C assimilation in a freshwater alga. Bull. Environ. Contam. Toxicol. 6(1): 1-8.

Stanley, J.G. and J.G. Trial. 1980. Disappearance constants of carbaryl from streams contaminated by forest spraying. Bull. Environ. Contam. Toxicol. 25(5): 771-776.

Statham, C.N., S.K. Pepple and J.J. Lech. 1975. Biliary excretion products of l-naphthyl-N-methylcarbamate-l-14C (carbaryl) in rainbow trout (Salmo gairdneri). Drug Metab. Dispos. 3(5): 400-406.

Statham, C.N., M.J. Melancon, Jr. and J.J. Lech. 1976. Bioconcentration of xénobiotics in trout bile: A proposed monitoring aid for some waterborne chemicals. Science. 193(4254): 680-681.

Stewart, N.E., R.E. Millemann and W.P. Breese. 1967. Acute toxicity of the insecticide, Sevin, and its hydrolytic product, 1-naphthol, to some marine organisms. Trans. Am. Fish. Soc. 96(1): 25-30.

Street, J.C. and R.P. Sharma. 1975. Alteration of induced cellular and humoral immune responses by pesticides and chemicals of environmental concern. Quantitative studies of immunosuppression by DDT, Aroclor 1254, carbaryl, carbofuran and methylparathion. Toxicol. Appl. Pharmacol. 32(3): 587-602.

Strother, A. 1970. Comparative metabolism of selected N-methylcarbamates by human and rat liver fractions. Biochem. Pharmacol. 19(8): 2525-2529.

Strother, A. and L. Wheeler. 1980. Excretion and disposition of (14C) carbaryl in pregnant, nonpregnant and fetal tissues of the rat after acute administration. Xenobiotica. 10(2): 113-124.

Sud, R.K., A.K. Sud and K.G. Gupta. 1972. Degradation of Sevin (1-naph-thyl-N-methyl carbonate) by <u>Achromobacter</u> sp. Ch. Mikrobiol. 87: 353-358.

Sullivan, L.J., J.M. Eldridge, J.B. Knaak and M.J. Tallant. 1972. 5,6-Dihydro-5,6-dihydroxycarbaryl glucuronide as a significant metabolite of carbaryl in the rat. J. Agric. Food Chem. 20(5): 980-985.

Swann, R.L., D.A. Laskowski, P.J. McCall, K. Vander Kuy and H.J. Dishburger. 1980. A rapid method for first approximation of environmental parameters utilizing reverse phase high performance liquid chromatography (RP-HPLC). Unpublished paper.

Szeto, S.Y., H.R. MacCarthy, P.C. Oloffs and R.F. Shepherd. 1979. The fate of acephate and carbaryl in water. J. Environ. Sci. Health. B14: 635-654.

Thomas, J.A. 1974. Actions of pesticides and other drugs on the male reproductive system. NTIS PB 237381/9GA. 41 p.

Thomas, J.A. 1981. Introductory remarks on the testes. Environ. Health Perspect. 38(0): 3-4.

Thomas, J.A., C.S. Dieringer and L. Schein. 1974. Effects of carbaryl on mouse organs of reproduction. Toxicol. Appl. Pharmacol. 28: 142-144.

Triolo, A.J., W.R. Lang, J.M. Coon, D. Lindstrom and D.L. Herr. 1982. Effect of the insecticides toxaphene and carbaryl on induction of lung tumors by benzo(a)pyrene in the mouse. Toxicol. Environ. Health. 9(4): 637-649.

Tsutsui, T., H. Maizumi, J.A. McLachlan and J.C. Barrett. 1983. Aneu-ploidy induction and cell transformation by diethylstilbestrol: A possible chromosomal mechanism in carcinogenesis. Cancer Res. 43: 3814-3821.

Tu, C.M. and J.R.W. Miles. 1976. Interactions between insecticides and soil microbes. Residue Rev. 64: 17-65.

Uchiyama, M., M. Takeda, T. Suzuki and K. Yoshikawa. 1975. Mutagenicity of nitroso derivatives of N-methylcarbamate insecticides in microbiological method. Bull. Environ. Contam. Toxicol. 14: 389-394.

U.S. EPA. 1981a. Treatability Manual. I. Treatability data. U.S. EPA, Washington, DC. EPA 600/2-82001A.

U.S. EPA. 1981b. Preliminary report on the mutagenicity of carbaryl. Reproduction Effects Assessment Group, Washington, DC. EPA 600/6-81-001.

U.S. EPA. 1983. AQUIRE: Aquatic Information Retrieval Database. Online: March 18.

Vashakidze, V.I. 1965. Some questions on the harmful influence of Sevin on the sexual function of experimental animals. Soobshch. Akad. Nauk. Gruz. SSR. 38: 471. (Cited in NIOSH, 1976)

Vasilos, A., V.D. Dmetrienko, and I.G. Shroyt. 1972. Colchicine-like action of Sevin on human embryonic fibroblasts in vitro. Bull Eskp. Biol. Med. 73: 91-93.

Vasilos, A.F., V.D. Dmitrienko, and I.G. Shroit. 1975a. Disruption of the mitotic system following acute Sevin poisoning. Izv. Akad. Nauk Mold. SSR Ser. Biol. i.Khim. Nauk. 3: 64-67.

Vásilos, A.F., V.D. Dmitrienko, and I.G. Shroit. 1975b. Changes in mitosis in chronic poisoning of rats with Sevin. Izv. Akad. Nauk Mold. SSR Ser. Biol. i.Khim. Nauk. 4: 45-47.

von Rumker, R., E.W. Lawless and A.F. Meiners. 1974. Production, distribution, use and environmental impact potential of selected pesticides. Prepared by Midwest Research Institute, RVR Consultants, Kansas City, MO for Council on Environmental Quality, Washington, DC. Contract No. RQC-311. NTIS PB 74 238-795.

Wakakura, M., S. Ishikawa and S. Uga. 1978. Ultrastructural hepatic changes by carbamate pesticide (Sevin) in rats. Environ. Res. 16(1-3): 191-204.

Walsh, G.E. and S. Alexander. 1980. A marine algal bioassay method: Results with pesticides and industrial wastes. Water Air Soil Pollut. 13(1): 45-55.

Wauchope, R.D. and R. Haque. 1973. Effects of pH, light and temperature on carbaryl in aqueous media. Bull. Environ. Contam. Toxicol. 9(5): 257-260.

Weil, C.S., M.D. Woodside, C.P. Carpenter, and H.F. Smyth, Jr. 1972. Current status of tests of carbaryl for reproductive and teratogenic effect. Toxicol. Appl. Pharmacol. 21: 390-404.

Weil, C.S., M.D. Woodside, J.B. Bernard, N.I. Condra, J.M. King and C.P. Carpenter. 1973. Comparative effect of carbaryl on rat reproduction and guinea pig teratology when fed either in the diet or by stomach intubation. Toxicol. Appl. Pharmacol. 26(4): 621-638.

Weis, P. and J.S. Weis. 1974. Cardiac malformations and other effects due to insecticides in embryos of the killifish, <u>Fundulus heteroclitus</u>. Teratology. 10(3): 263-267.

Whitehurst, W.E., E.T. Bishop, F.E. Critchfield, et al. 1963. The metabolism of Sevin in dairy cows. J. Agric. Food Chem. 11: 167. (Cited in Mount and Oehme, 1981)

WHO (World Health Organization). 1974. 1973 Evaluation of some pesticide residues in food. WHO Pestic. Res. Serv., No. 3. p. 141-176. (Cited in IARC, 1976)

Whorton, M.D., T.H. Milby and H.A. Stubbs. 1979. Testicular function among carbasyl-exposed employees. J. Tox. Environ. Health. 5: 929-941.

Wills, J.H., E. Jameson, A. Stein, et al. 1967. Effects of oral doses of carbaryl on man. Toxicol. Appl. Pharmacol. 10: 390. (Cited in Desi et al., 1974)

Wills, J.H., E. Jameson and F. Coulston. 1968. Effects of oral doses of carbaryl on man. Clin. Toxicol. 1: 265. (Cited in IARC, 1976)

Wiltrout, R.W., C.D. Ercegovich and W.S. Ceglowski. 1978. Humoral immunity in mice following oral administration of selected pesticides. Bull. Environ. Contam. Toxicol. 20(3): 423-431.

Wéjciechowski, J.P., P. Kaur and P.S. Sabharwal. 1982. Induction of ouabain resistance in V-79 cells by four carbamate pesticides. Environ. Res. 29(1): 48-53.

Wolfe, N.L., R.G. Zepp, G.L. Baughman, R.C. Fincher and J.A. Gordon. 1976. Chemical and photochemical transformation of selected pesticides in aquatic systems. U.S. EPA, Athens, GA. EPA 600/3-76-067. 153 p.

Wolfe, N.L., R.G. Zepp and D.F. Paris. 1978. Use of structure-reactivity relationships to estimate hydrolytic persistence of carbamate pesticides. Water Res. 12(8): 561-563.

Woodward, D.F. and W.L. Mauck. 1980. Toxicity of five forest insecticides to cutthroat trout and two species of aquatic invertebrates. Bull. Environ. Contam. Toxicol. 25(6): 846-854.

Wuu, K.D. and W.F. Grant. 1966. Morphological and somat chromosomal abberations induced by pesticides in barley (<a href="Hordeum vulgare">Hordeum vulgare</a>). Can. J. Genet. Cytol. 8: 481-501.

Wuu, K.D. and W.F. Grant. 1967. Chromosomal abberations induced by pesticides in meiotic cells of barley. Cytologia. 32: 31-41.

## APPENDIX: LITERATURE SEARCHED

This profile is based on data identified by computerized literature searches of:

CA SEARCH (Files 308, 309, 310, 311, 320)

TOXLINE

MEDLINE

RTECS

SCI SEARCH

OHM TADS

STORET

SRC Environmental Fate Data Bases

SANSS

**AOUIRE** 

**EPCASR** 

Chemical Industry Notes

Most of these searches were conducted in February, 1983; a few were conducted March-May, 1983. In addition, hand searches were made of Chemical Abstracts (Collective Indices 7 and 8th), and the following secondary sources were reviewed:

ACGIH (American Conference of Governmental Industrial Hygienists). 1980. TLVs: Documentation of the Threshold Limit Values, 4th ed. (Includes Supplemental Documentation, 1981). Cincinnati, OH. 486 p.

ACGIH (American Conference of Governmental Industrial Hygienists). 1982. Documentation of the Threshold Limit Values for Chemical Substances in Work Air. Cincinnati, OH. 94 p.

Bruin, P., G.J. Bergen and J.J. Desta, Ed. 1980. Handling Chemicals Safely, 2nd ed. Dutch Assoc. Safety Experts, Dutch Chemical Industry Assoc., and Dutch Safety Institute, The Netherlands. 1013 p.

Clayton, G.D. and F.E. Clayton, Ed. 1981. Patty's Industrial Hygiene and Toxicology, Vol. 2A, 3rd rev. ed. John Wiley and Sons, NY. 2878 p.

Clayton, G.D. and F.E. Clayton, Ed. 1981. Patty's Industrial Hygiene and Toxicology, Vol. 2B, 3rd rev. ed. John Wiley and Sons, NY. p. 2879-3816.

Clayton, G.D. and F.E. Clayton, Ed. 1982. Patty's Industrial Hygiene and Toxicology, Vol. 2C, 3rd rev. ed. John Wiley and Sons, NY. p. 3817-5112.

Hamilton, A. and H.L. Hardy. 1974. Industrial Toxicology, 3rd ed. Publishing Sciences Group, Inc., MA. 575 p.

ITII (International Technical Information Institute). 1982. Toxic and Hazardous Industrial Chemicals Safety Manual for Handling and Disposal with Toxicity and Hazard Data. ITII, Tokyo, Japan. 700 p.

Muir, G.D., Ed. 1977. Hazards in the Chemical Laboratory, 2nd ed. The London Chemical Society, London. 473 p.

NTP (National Toxicology Program). 1982. Carcinogenesis Testing Program. Chemicals on Standard Protocol. Management Status, December 31.

Proctor, N.H. and J.P. Hughes. 1978. Chemical Hazards of the Workplace. J.B. Lippincott Co., Philadelphia, PA. 533 p.

Sax, I.N. 1979. Dangerous Properties of Industrial Materials, 5th ed. Van Nostrand Reinhold Co., NY.