

Appendix C

Verification and Validation of the EPA's Composite Model for Transformation Products (EPACMTP), and its Derivatives

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Appendix C Verification and Validation of the EPA's Composite Model for Transformation Products (EPACMTP), and its Derivatives

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C.1 Introduction

C.1.1 General Background

The U.S. Environmental Protection Agency's (EPA) Office of Solid Waste (OSW) has developed a probabilistic (Monte Carlo) groundwater flow and transport modeling approach to assess potential exposure of groundwater to toxic chemical constituents in wastes that are managed in Subtitle D industrial waste management units under RCRA regulations. The exposure to groundwater is expressed as concentrations of potential contaminants at a drinking water well (receptor well) located downgradient of the waste disposal facility. This modeling methodology has been incorporated into the EPA's Composite Model for Leachate Migration with Transformation Products Code (EPACMTP) (U.S. EPA, 1996a, b, c, d). The U.S. EPA OSW has applied EPACMTP, as a general fate and transport model, to establish regulatory levels for concentrations of chemicals in the Subtitle D industrial waste management units for several proposed rules and listing determinations.

In 1999, the flow and transport components for the vadose-zone and aquifer modules were extracted from EPACMTP for U.S. EPA's Hazardous Waste Identification Rule (HWIR99)-Multimedia, Multipathway, and Multireceptor Assessment (3MRA). The pseudo-3-D module was developed for the aquifer during this period (U.S. EPA, 1999a). At the same time, three ancillary modules were developed to include the effects of fractures, heterogeneity, and anaerobic biodegradation. The technical framework of the flow and transport modules along with the three ancillary modules were subject to peer and public reviews. A number of comments were received. The comments have recently been compiled and priorities assigned. One of the high priority issues is the verification and validation of various 3MRA simulation modules. Prior to performing additional verification and validation to the 3MRA vadose-zone and aquifer modules, it is necessary to compile all pertinent information relating to the verification and validation of the agent validation of the agent validation of the agent validation and val

C.1.2 Definition of Verification and Validation

For a simulation module to gain credibility that it can be used to simulate natural phenomena with reasonable accuracy, apart from good documentation and rigorous reviews, it has to undergo a two-step process: verification, and validation; which are the two most important steps in the quality assurance program of a module (van der Heijde, 1987). A description of verification and validation of a module is presented below.

Verification. The objective of the code verification process is two-fold (National Research Council, 1990): (1) to demonstrate that the computational algorithms can accurately solve the governing equations and (2) to assure that the computer code is fully operational. A module is said to be 'verified' when it can be demonstrated that the mathematical framework embodied in the module is correct. A module may be verified by comparing its simulation results against known analytical solutions or numerical solutions from simulators based on similar or identical mathematical frameworks.

Most modules (especially those based on legacy codes) may have been verified according to the definition above.

Validation. Model validation is conducted after the verification step. (Note that the word 'model' is used under the subject of validation instead of 'module.' A model in this case is a combination of the mathematical framework embodied in the module and data representing a site or hydrogeologic system of interest.) The objective of model validation is to determine how well the mathematical representation of the processes describes the actual system behavior in terms of the degree of correlation between model calculations and actual observed data (National Research Council, 1990). Ideally, results should be compared to the results of a well-defined field experiment or a well-conditioned laboratory experiment.

Validation of the predictive capabilities of the model is accomplished through comparison with experimental data by using independent estimates of the parameters. In principle, this is the ideal approach to validation. However, unavailability and inaccuracy of field characterization data often prevent the application of such a rigid validation approach to actual field systems.

Methods that may be used to validate a model include:

- Using field data. Typically, parts of the field data are designated as calibration data, and a calibrated site model is obtained through reasonable adjustment of parameter values. Other parts of the field data are designated as validation data; the calibrated site model is used in a predictive mode to generate similar data for comparison. For instance a groundwater model may be calibrated against water level measurements, and then validated by comparing predicted against measured contaminant concentrations in downgradient wells. Although this procedure will not allow complete validation of a modeling process, it will provide some insight into potential problems of the model. This approach is limited because splitting of a (groundwater) data set into two components does not yield completely independent data sets. Two completely independent sets of data usually do not exist, so that the verification (calibration) data and validation data are related.
- Using synthetic data. At times, the implementation of the above-described validation approach using field data is not possible nor practical due to lack of adequate, complete and high-quality field data. Thus, testing of groundwater models is limited to extended verifications, and code comparisons. In this case, a newly developed model is compared with established models designed to solve the same type of problems. If the results from the new code do not deviate significantly from the those obtained with the existing codes, a relative or comparative validity is established. If code comparison is used to evaluate a new code, the code should again be validated as soon as adequate data sets become available.

Absolute validity of a model is never determined. Establishing absolute validity requires testing over the full range of conditions for which the model is designed, an exercise that is almost never possible or practical. As stated by the National Research Council (1990), a validated model should not be applied in a predictive mode beyond its historically observed range or range of calibration.

C.2 Technical Background of the Vadose-zone and Aquifer Modules

C.2.1 EPACMTP

The U.S. EPA OIW has developed a probabilistic (Monte Carlo) groundwater flow and transport modeling approach to assess potential exposure of groundwater to toxic chemical constituents in wastes that are managed in Subtitle D industrial waste management units under RCRA regulations. The exposure to groundwater is expressed as concentrations of potential contaminants at a drinking water well (receptor well) located downgradient of the waste disposal facility. This modeling methodology has been incorporated into EPACMTP (U.S. EPA, 1996 a, b, c, d). The OIW of the U.S. EPA has applied EPACMTP, as a general fate and transport model, to establish regulatory levels for concentrations of chemicals in the Subtitle D industrial waste management units for several proposed rules and listing determinations.

The current subsurface flow and transport components in EPACMTP comprise the following modules:

- 1-D vertical variably saturated flow and transport submodules—collectively referred to as the *vadose-zone module*;
- 3-D saturated flow and transport submodules—collectively referred to as the 3-D aquifer module;
- Quasi-3-D saturated flow and transport submodules—collectively referred to as the *quasi-3-D aquifer module*; and
- Areal two-dimensional (vertically averaged) saturated flow and transport submodules—collectively referred to as the *areal two-dimensional aquifer module*.

The first module is used to simulate flow and transport of constituents in the vadose-zone. The following three modules are used to simulate flow and transport in the aquifer beneath the vadose-zone.

Details of the above modules are provided in the EPACMTP background document (U.S. EPA 1996a).

C.2.2 3MRA Subsurface Flow and Transport Modules

During the past 5 years, a number of enhancements have been made to EPACMTP and EPACMTP-derived subsurface fate and transport modules. These include the development of a computationally efficient pseudo-3-D module for the modeling system (U.S. EPA, 1999c), a new surface impoundment module for the 3MRA modeling system (U.S. EPA, 1999d), methodologies to handle fractures, heterogeneity, and anaerobic biodegradation based on a new nation-wide rate database (U.S. EPA, 1999c, e). In addition, a number of comments from public and peer reviews regarding features and theoretical aspects of EPACMTP have been received and evaluated.

Based on the existing aquifer module in EPACMTP, a pseudo-3-D aquifer module was developed as a component of the 3MRA modeling system, which is an integrated framework

consisting of medium-specific pollutant fate, transport, exposure, and risk modules. The 3MRA modeling system was first applied as part of the U.S. EPA's 1999 Hazardous Waste Identification Rule (HWIR) notice. The differences between the pseudo-3-D and the EPACMTP modules are listed below:

- Solution schemes. The pseudo-3-D scheme is based on a hybrid numerical-analytical solution technique (U.S. EPA, 1999c). This solution scheme produces receptor well concentrations that are somewhat more conservative than those generated by the fully 3-D module. Its computational speed is greater than that of the fully 3-D module by a factor of approximately 300.
- Chemical concentrations at receptor wells. The concentrations at receptors are reported in complete breakthrough curves. EPACMTP reports peak and averagedaround-the-peak concentrations.
- Proximity of surface bodies. A surface water body into which groundwater discharges pollutant flux is part of the pseudo-3-D module but not the EPACMTP-based module.
- Fractured media. Correction for the magnitude of a really averaged hydraulic conductivity is allowed for fractured media in the pseudo-3-D module but not the current module in EPACMTP.
- Heterogeneity. Correction for receptor well concentration to account for local variability in hydraulic conductivity and porosity is allowed for heterogeneous media in the pseudo-3-D module but not the EPACMTP-based module.
- Anaerobic biodegradation. A national database for anaerobic biodegradation was developed for a number of organic chemicals (U.S. EPA, 1999e). The pseudo-3-D module generates a value of chemical-specific anaerobic biodegradation rate based on the probabilistically selected pH regime, temperature range, and redox environment.
- **Time-dependent infiltration rate.** A new module has been developed to simulate a series of different infiltration rates to represent various stages of surface impoundment operation. In each stage, the corresponding flow field is assumed to be steady. At this time, this module has not yet been fully tested.

C.3 Verification History

EPACMTP has been verified extensively by comparing its simulation results against both analytical and numerical solutions. Numerous verification cases were conducted from 1991–2000. A summary of the verification cases is provided in the following subsections. The accompanying figures for selected test cases are presented in the designated appendices.

C.3.1 ORD Verification (1992-1993)

In 1992, a verification analysis of the newly developed EPACMTP was performed by the Office of Research and Development (ORD) of the U.S. EPA (U.S. EPA, 1992). A list of verification cases in the verification exercise is listed in Table C.3-1. As shown in the table, two steps of code verification were conducted: a re-verification of the original test problems and data files provided by HydroGeoLogic, Inc. and independent verification using alternative test criteria.

Based on the analysis of Tetra Tech, some technical limitations in the EPACMTP code were identified. One of the weaknesses, which occurred in the aquifer module, pertained to potential mass loss of contaminants from the system due to the upstream boundary proximity and conditions in EPACMTP. The code was modified in response to the comments (HydroGeoLogic, Inc., 1993).

Table C.3-1Summary of EPACMTP Verification by the Office of Research and Development,
U.S. EPA, and Tetra Tech, Inc. in 1992-1993 (from U.S. EPA, 1992)

Case	Description	Reverification of HGL Tests	Independent Verification
1	Steady-state, aquifer flow, single layer	Yes	Yes
2	Steady-state, vadose-zone transport, two layers		Yes
3	Transient vadose-zone transport, single layer - analytical solution	Yes	Yes
4	Transient vadose-zone transport, single layer - numerical solution		Yes
5	Transient vadose-zone transport, single layer, nonconservative solute—numerical solution		Yes
6	Transient vadose-zone transport, three layers, nonconservative solute—numerical solution		Yes
7	Transient vadose-zone transport, single layer, nonlinear adsorption—numerical solution	Yes	
8	Multiple species transport; 3-member chain decay; source decay	Yes	
9	Steady-state, aquifer flow	Yes	Yes
10	Quasi-3-D aquifer transport—numerical solution	Yes	Yes
11	Nonlinear aquifer transport	Yes	Yes
12	3-species transport, 2-D (x,y)	Yes	Yes
13	7-species transport, 2-D (x,y)	Yes	Yes
14	Full-3-D aquifer flow and transport	Yes	

C.3.2 Module-Level Verification (1993-1994)

A module-level verification task was performed between 1993-1994 and reported in EPACMTP Background Documents (U.S. EPA, 1996 b, d). Numerous components of EPACMTP's flow and transport sub-modules, in both the vadose and aquifer modules, were verified between 1993-1994 against analytical solutions, and numerical solutions from a number of simulators with similar mathematical frameworks. Details of the verification are presented below.

C.3.2.1 Vadose-Zone Module Verification

The vadose-zone and the aquifer modules were subdivided into the flow and transport submodules. The ten verification cases for the vadose-zone module are summarized in Table C.3-2 and are briefly described below. Excerpts of verification results for the vadose-zone test cases are presented in figures in Attachment C-1. Reference to the figures in Attachment C-1 is provided in Table C.3-2. Additional information regarding the test cases and respective verification results may

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Case	Description	Verification Method	Excerpts of Verification Result Presented in
-	Steady-state infiltration	Semi-analytical FECTUZ module vs. fully numerical finite element VADOFT	Figure C-1-1, Attachment C1
0	Steady-state infiltration in a layered soil	Semi-analytical FECTUZ module vs. fully numerical finite element VADOFT	Figure C-1-2, Attachment C1
3	Steady-state infiltration in a layered soil with a ponding depth	Semi-analytical FECTUZ module vs. fully numerical finite element VADOFT	Figure C-1-3, Attachment C1
4	Steady-state transport in a layered soil	Steady-state analytical solution vs. finite element numerical solution of FECTUZ	Figure C-1-4, Attachment C1
5	1-D transient transport under pulse input conditions	Semi-analytical solution vs. numerical finite element module of FECTUZ and the HYDRUS code	Figure C-1-5, Attachment C1
9	1-D transport of a conservative solute species in a saturated soil column of semi-infinite length	Numerical solution of FECTUZ vs. analytical solution of Ogata and Banks (1961)	Figure C-1-6, Attachment C1
7	1-D transport of a conservative and nonconservative solute species in a saturated soil column of finite length	FECTUZ vs. analytical solution of van Genuchten and Alves (1982)	Figure C-1-7, Attachment C1
8	Transport of a conservative species in a layered soil column	FECTUZ vs. Shamir and Harleman (1967) and Hadermann (1980)	Figures C-1-8 and C-1-9, Attachment C1
6	Transient transport under conditions of nonlinear adsorption with a pulse source	FECTUZ vs. finite difference code MOB1	Figure C-1-10, Attachment C1
10	Multispecies transport with three member, straight decay chain with a decaying source boundary condition	FECTUZ vs. analytical solution modified from Hodgkinson and Maul (1985)	Figure C-1-11, Attachment C1

be found in U.S. EPA (1996b). The vadose-zone module of EPACMTP was originally called FECTUZ. The numerical transport simulation in FECTUZ is no longer part of EPACMTP.

The first three test cases of the vadose-zone module are for the flow sub-module and focus on steady-state flow within layered and nonlayered soils. They were verified by comparing the results of the semi-analytical FECTUZ (U.S. EPA, 1989) module against the numerical finite element VADOFT model (Huyakorn, et al., 1987). Test Case 1 evaluated steady-state infiltration in a soil. Test Cases 2 and 3 are similar, both involving steady-state infiltration in a layered soil, whereas Test Case 3 introduced the surface impoundment boundary condition (ponding depth) to the system.

The last seven test cases, summarized in Table C.3-2, pertain to transport sub-module verification. Test Cases 4 and 5 tested the analytical steady-state transport module and the semianalytical transport solution, respectively. Test Case 4 involved steady-state transport in a layered soil and verification against the FECTUZ numerical solution, while Test Case 5 evaluated transient transport with verification against both the FECTUZ numerical solution and the HYDRUS code (Kool and van Genuchten, 1991).

Test Cases 6 through 10 utilize the FECTUZ numerical solution to examine transport of a contaminant in a soil column. Case 6 concerns 1-D transport of a conservative solute species and is verified against the analytical solution of Ogata and Banks (1961). Test Case 7 considers downward vertical transport of both conservative and nonconservative constituents. The results are compared against the analytical solution given by van Genuchten and Alves (1982). Test Case 8 concerns 1-D transport of a conservative solute species in a layered soil column. Two sub-cases with different dispersivity values were compared with the analytical solutions presented by Shamir and Harlemann (1967) and Hadermann (1980). Test Case 9 considers solute transport with both linear and nonlinear adsorption. This is verified against the MOB1 finite element solution (van Genuchten and Alves, 1982). Test Case 10 examines transport of a 3-member, straight decay chain and is verified against the analytical solution, modified from Hodgkinson and Maul (1985).

C.3.2.2 Aquifer Module Verification

The saturated zone module of EPACMTP was originally developed on a stand alone basis and called CANSAZ-3-D. Seven benchmark problems were analyzed to verify the flow and transport solutions in the CANSAZ-3-D modules; (Sudicky et al., 1990) and are summarized in Table C.3-3. Excerpts of verification results for the test cases are presented in Attachment C-2. Reference to the figures in Attachment C-2 is provided in Table C.3-3. Additional information regarding the test cases and respective verification results may be found in U.S. EPA (1996b). Test Case 1 was designed to verify the 3-D steady-state groundwater flow solution. For this purpose, the hydraulic head and groundwater flow velocities obtained from CANSAZ-3-D were compared against the MNDXYZ analytical solution (Ungs, 1986; Attachment C-2 of U.S. EPA, 1996b). Test Case 2 was designed to compare the analytical and numerical transport solutions for the case of single species transport in a uni-directional steady-state groundwater flow field. Test Case 3 involved 2-dimensional transport of a 3-member decay chain. The CANSAZ-3-D results for this test problem were verified against the numerical VAM2D code (Huyakorn et al., 1992). Test Case 4 involved verification of CANSAZ-3-D against an analytical solution (Sudicky et al., 1991) for a case involving a complex, seven-member branched decay chain. Test Case 5 was designed to verify the nonlinear sorption option. This problem involves 1-D flow and transport with a nonlinear Freundlich isotherm. CANSAZ-3-D was verified against the numerical MOB1 (van Genuchten, 1981) and FECTUZ. Test Case 6 involves fully 3-D flow and transport. The CANSAZ-3-D solution was compared against results obtained with the 3-D DSTRAM flow and transport code (Huyakorn and

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Table C-3-3. Verification Cases for the Saturated Zone Module (1993-1994)

Case	Description	Verification Method	Excerpts of Verification Results Presented in
1	Steady-state groundwater flow in a 3-D domain	CANSAZ-3-D vs. analytical solution of MNDXYZ	Figure C.2-1, Attachment C.2
7	Single species transport in a uni-directional flow field— analytical and numerical transport modules	CANSAZ-3-D vs. 3-D analytical solution	Figure C.2-2, Attachment C.2
3	2-D transport of a 3-member decay chain. Steady-state flow and transient solute transport in an unconfined aquifer	CANSAZ-3-D vs. VAM2D	Figure C.2-3, Attachment C.2
4	2-D transport of a complex, seven-member branched decay chain with 1-D groundwater flow	CANSAZ-3-D vs. Gaussian source analytical solution of Sudicky et al. (1991)	Figure C.2-4, Attachment C.2
5	Nonlinear sorption reactions in a 1-D, steady-state flow and transient transport. Pulse source using a Freundlich isotherm	CANSAZ-3-D vs. MOB1 and FECTUZ	Figure C.2-5, Attachment C.2
9	Steady-state flow and transport modeling of a single conservative species in 3-D aquifer domain	CANSAZ-3-D vs. DSTRAM	Figure C.26, Attachment C.2
L	Steady-state groundwater flow and transient solute transport in 3-D aquifer domain with a horizontal patch source	CANSAZ-3-D vs. VAM3-D	Figure C.2-7, Attachment C.2

Panday, 1991). Test Case 7 was designed to evaluate the automatic model domain discretization option for a 3-D flow and transport problem and was verified against the numerical VAM3-D code (Panday et al., 1993).

C.3.2.3 Metals Transport Module

The major modifications to accommodate metals transport with nonlinear sorption were made to the vadose-zone module, therefore the verification cases are applicable to this module. Five verification test cases are summarized in Table C.3-4 and excerpts of verification results are presented in Attachment C.3. Reference to the figures in Attachment C.3 is provided in Table C.3-4. Additional information regarding the test cases and respective verification results may be found in U.S. EPA (1996d). Test Case 1 involved continuous release of a nonsorbing solute to test the linear adsorption partitioning capabilities. An analytical solution from Ogata (1970) was compared against the EPACMTP results. Test Case 2 involved nonlinear Freundlich adsorption isotherms. The Freundlich isotherm was represented by its closed form. Two different source conditions were utilized: continuous and finite sources. Freundlich exponents greater than and less than one were examined. The results from EPACMTP were compared with those from HYDRUS. Test Case 3 involves transport of lead in a fully saturated soil column. The verification of this case was performed by comparing the computed cumulative mass against the total input mass. Test Case 4 involves 1-D transport of a solute, with Freundlich exponents of less than and greater than 1 and was verified against HYDRUS.

C.3.3 Verification of Individual Modules and a Composite Model in EPACMTP (1997)

In 1997, a testing plan was developed for EPACMTP code verification (U.S. EPA, 1997), in accordance with the ASTM, "*Standard Guide for Developing and Evaluating Ground-Water Modeling Codes*" (ASTM, 1996). The verification process focused on a single problem geometry, representative of waste disposal scenarios in terms of spatial dimensionality and climatic/ hydrogeological conditions. The verification process first subdivided the problem setting into individual hydrogeologic components, assessed their functionality relative to an overall fate and transport problem, and then compared each component to analytical solutions or other codes.

The vadose-zone module, the aquifer module, and the composite model were verified following the ASTM standards. The vadose-zone problem geometry was a 1-D column extending from the land surface to the water table. Boundary conditions for numerical contaminant transport involved a continuous source on the water table beneath the waste management unit. The region of the water table outside the source area received constant recharge from the ground surface. Ten test cases were conducted. These test cases may be subdivided into those for the vadose-zone module, the aquifer module, and the composite model, and are summarized in Tables C.3-5a, C.3-5b, and C.3-5c, respectively.

C.3.3.1 Vadose-Zone Module Verification

The vadose-zone module verification is summarized with four cases in Table C.3-5a and excerpts of verification results are presented in Attachment C.4. Reference to the figures in Attachment C.4 is provided in Table C.3-5a. Additional information regarding the test cases and respective verification results may be found in U.S. EPA (1997). Test Case 1 evaluated steady-state variably saturated flow and Test Case 2 considers infiltration through a clay liner and ponding depth.

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Table C.3-4. Verification Cases for Metals Transport in the Vadose-Zone (1993-1994)

Case	Description	Verification Method	Excerpts of Verification Results Presented in
←	Linear adsorption partitioning with continuous release of a nonsorbing solute	Analytical solution (Ogata, 1970) vs. EPACMTP result	Figure C.3-1, Attachment C.3
2	Nonlinear adsorption isotherm. The Freundlich isotherm was represented by its closed function form. Freundlich isotherms greater than and less than one were considered for continuous and finite source conditions	EPACMTP vs. HYDRUS	Figures C.3-2 and C.3-3, Attachment C.3
З	Transport of lead using MINTEQA2-generated isotherms	Cumulative vs. total input mass	Figure C.3-4, Attachment C.3
4	Pulse source and Freundlich exponents of 0.5, 0.8, and 1.5	Analytic solution vs. HYDRUS	Figure C.3-5, Attachment C.3

	
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Case	Description	Verification Method	Excerpts of Verification Results Presented in
-	Steady-state variably saturated flow	EPACMTP vs. STAFF3D	Figure C.4-1, Attachment C.4
2	Infiltration through a clay liner from a surface impoundment	EPACMTP vs. STAFF3D	Figure C.42, Attachment C.4
с	Contaminant transport with linear sorption and decay	Numerical EPACMTP vs. analytical EPACMTP	Figure C.4-3, Attachment C.4
4	Contaminant transport with branched chain decay and linear sorption	EPACMTP vs. VAM2D	Figure C.4-4, Attachment C.4

Table C.3-5a. Verification Cases for the Vadose-Zone Module (1997)

Table C.3-5b. Verification Cases for the Aquifer Module (1997)

Case	Description	Verification Method	Excerpts of Verification Results Presented in
5	3-D steady-state groundwater flow	EPACMTP vs. MNDXYZ analytical solution	Figure C.5-1 and C.5-2, Attachment C.5
9	3-D contaminant transport with linear sorption and decay	EPACMTP numerical module vs. EPACMTP analytical module; EPACMTP vs. VAM3DF	Figure C.5-3, Attachment C.5
7	3-D contaminant transport with four species, branched chain decay and linear sorption	EPACMTP vs. STAFF3D	Figure C.5-4, Attachment C.5

Table C.3-5c. Verification Cases for Composite Module (1997)

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Case	Description	Verification Method	Excerpts of Verification Results Presented in
8	Composite flow and contaminant transport	EPACMTP vs. VAM3DF	Figure C-6-1, Attachment C-6
6	Monte Carlo analysis based on composite flow and contaminant transport	EPACMTP vs. VAM3DF	Figure C-6-2, Attachment C-6

Test Cases 3 and 4 both considered contaminant transport with linear sorption, but Test Case 3 examined linear decay while Test Case 4 evaluated four species with branched chain decay. Test Cases 1 and 2 were verified against STAFF3D (HydroGeoLogic, Inc., 1995a), Test Case 4 was verified against VAM2D while Test Case 3 compared the steady-state results from numerical and analytical transport modules of EPACMTP.

C.3.3.2 Aquifer Module Verification

The aquifer module verification is summarized with three cases in Table C.3-5b and excerpts of verification results are shown in Attachment C.5. Reference to the figures in Attachment C.5 is provided in Table C.3-5b. Additional information regarding the test cases and respective verification results may be found in U.S. EPA (1997). The 3-D steady-state fully saturated flow module in EPACMTP was verified against the analytical solution MNDXYZ in Test Case 5. Test Cases 6 and 7 examined contaminant transport and were verified against VAM3DF (HydroGeoLogic, Inc., 1995b) and STAFF3D, respectively. Test Case 6 involved transport of a contaminant with linear sorption and decay, while Test Case 7 involved linear sorption and a four species, branched chain decay.

C.3.3.3 Composite Model Verification

The EPACMTP composite model comprises the following fate and transport modules: a vadose-zone module, and a aquifer (saturated zone) module. These modules are connected according to the detailed description in U.S. EPA (1996). The composite model verification is summarized with two test cases in Table C.3-5c and excerpts of verification results are shown in Attachment C.6. Reference to the figures in Attachment C-6 is provided in Table C.3-5c. Additional information regarding the test cases and respective verification results may be found in U.S. EPA (1997). Test Case 8 considered the composite flow and contaminant transport structure. Test Case 9 assessed the sensitivity of the geometric assumptions used to develop EPACMTP. A limited Monte-Carlo analysis was performed to assess the sensitivity of the confined water table assumption to predicting the probability of exceedance at a monitoring well. Both Test Cases 8 and 9 were verified against VAM3DF which is a 3-D, variably saturated numerical flow and transport code.

C.3.4 Verification of 3MRA Subsurface Flow and Transport Modules (1999)

In 1999, the flow and transport components for the vadose-zone module and aquifer module were extracted from EPACMTP to provide the groundwater pathway module for the 3MRA system. The basic premise for verification of the vadose-zone and aquifer modules was that EPACMTP had been rigorously verified, so it was sufficient to show that the modules reproduced EPACMTP results. Therefore, both the steady-state flow and transport sub-modules of the aquifer module (U.S. EPA, 1999c) and the flow and transport sub-modules of the vadose-zone module (U.S. EPA, 1999b,c) were compared against the numerical results from EPACMTP to ensure that the extracted modules remained intact. There are two exceptions that will be discussed below. The new saturated zone pseudo-3-D module was also developed during this period (U.S. EPA, 1999a).

The 18 test cases for the vadose-zone, aquifer, and pseudo 3-D modules are summarized in Tables C.3-6a, C.3-6b, and C.3-6c, respectively. The figures are presented in Attachment C.7 through C.9. The vadose-zone problem geometry was a 1-D column extending from the land surface to the water table. Boundary conditions for numerical contaminant transport involved a continuous source on the water table beneath the waste management unit. The region of the water table outside the source area was also considered to be a recharge boundary.

C.3.4.1 Vadose-Zone Module Verification

There are eight vadose-zone module verification cases (Table C.3-6a). Excerpts of results for the verification cases are presented in Attachment C-7. Reference to the figures in Attachment C.7 is provided in Table C.3-6a. Additional information regarding the test cases and respective verification results may be found in U.S. EPA (1999b,c). All of the cases concern contaminant transport. Test Case 1 evaluated an exponentially depleting source. Test Case 2 involved transport of a contaminant with no sorption and no hydrolysis. Test Case 3 examined sorption and hydrolysis with one species, while Test Case 4 involved two species with chain decay. Test Case 5 examined linear and nonlinear metal transport using the MINTEQA2 isotherms. Test Case 6 evaluated biodegradation resulting in chain decay reactions with four species. Test Cases 7 and 8 examined contaminant concentration at a receptor well and pressure heads at each grid node, respectively. In this instance, both Test Case 7 and 8 were verified against MODFLOW-SURFACT (HydroGeoLogic, Inc., 1996), a 3-dimensional numerical groundwater flow and transport code.

C.3.4.2 Aquifer Module Verification

There are seven aquifer module verification cases (Table C-3-6b) with excerpts of verification results presented in Attachment C-8. Reference to the figures in Attachment C-8 is provided in Table C.3-6b. Additional information regarding the test cases and respective verification results may be found in U.S. EPA (1999c). Test Case 1 evaluated an exponentially depleting source. Test Case 2 involved transport of a conservative contaminant with no sorption and no hydrolysis. Test Case 3 examined sorption and hydrolysis with one species, while Test Case 4 involved two species with chain decay. Test Case 5 examined linear and nonlinear metal transport using the MINTEQA2 isotherms. Test Case 6 evaluated biodegradation resulting in chain decay reactions with four species. Test Case 7 evaluated the generated Monte Carlo distributions.

C.3.4.3 Pseudo-3-D Module Verification

There are three verification cases for the pseudo-3-D aquifer module (Table C.3-6c). Excerpts of verification results are shown in Attachment C-9. Reference to the figures in Attachment C.9 is provided in Table C.3-6c. Additional information regarding the test cases and respective verification results may be found in U.S. EPA (1999a,c). Test Case 1 examined the average groundwater specific flow rate determined by the saturated flow sub-module and was verified using Darcy's Law. Test Case 2 examined the numerical component of the contaminant transport sub-module and is verified using the analytical solution by Ogata (1970). Test Case 3 verified the combined analytical-numerical contaminant transport sub-module using verification results of Test Case 2 subject to the analytical portion of the aquifer transport sub-module.

C.3.5 Comprehensive Verification of the 3MRA Vadose-zone Pseudo-3-D Aquifer Modules (2000)

In 2000, a comprehensive verification was conducted of all of the components in the extracted aquifer and the vadose-zone modules (U.S. EPA, 2000a,b). For testing purposes, each component was executed as a stand-alone program outside of the 3MRA Software System environment.

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Case	Description	Verification Method	Excerpts of Verification Results Presented in
-	Exponentially depleting conservative source with no sorption or hydrolysis	Vadose-Zone Module vs. EPACMTP	Figure C.7-1, Attachment C.7
2	Constant-concentration source pulse with no sorption or hydrolysis	Vadose-Zone Module vs. EPACMTP	Figure C.7-2, Attachment C.7
3	Constant-concentration source pulse with sorption and hydrolysis, one species	Vadose-Zone Module vs. EPACMTP	Figure C.7-3, Attachment C.7
4	Constant-concentration source pulse with sorption and hydrolysis, and chain decay	Vadose-Zone Module vs. EPACMTP	Figure C.7-4, Attachment C.7
5	Metals: (mercury and lead), with constant-concentration source pulse with $MINTEQ\text{-}based$ sorption and no hydrolysis	Vadose-Zone Module vs. EPACMTP	Figures C.7-5 and C.7-6, Attachment C.7
Q	Constant-concentration source pulse with biodegradation, sorption and chain decay	Vadose-Zone Module vs. EPACMTP	Figure C.7-7, Attachment C.7
7	1-D contaminant transport between a top boundary at the bottom of the source zone and the water table with mass loading to the top boundary from the leachate flux from the source module	Vadose-Zone Module vs. MODFLOW-SURFACT	Figure C.7-8, Attachment C.7
ω	1-D variable saturated flow between a top boundary at the bottom of the source zone and the water table with mass loading to the top boundary from the leachate flux from the source module	Vadose-Zone Module vs. MODFLOW-SURFACT	Figure C.7-9, Attachment C.7

	
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Case	Description	Verification Method	Excerpts of Verification Results Presented in
-	Exponentially depleting source with no sorption or hydrolysis	Aquifer Module vs. EPACMTP	Figure C.8-1, Attachment C.8
7	Constant-concentration source pulse with no sorption or hydrolysis	Aquifer Module vs. EPACMTP	Figure C.8-2, Attachment C.8
e	Constant-concentration source pulse with sorption and hydrolysis, one species	Aquifer Module vs. EPACMTP	Figure C.8-3, Attachment C.8
4	Constant-concentration source pulse with sorption and hydrolysis, and two species with chain decay	Aquifer Module vs. EPACMTP	Figures C.8-4 and C.8-5, Attachment C.8
5	Metals: (mercury and lead), with constant-concentration source pulse with sorption and no hydrolysis	Aquifer Module vs. EPACMTP	Figure C.8-6, Attachment C.8
9	Constant-concentration source pulse with biodegradation, sorption and four species chain decay	Aquifer Module vs. EPACMTP	Figures C.8-7, C.8-8, C.8-9, and C.8-10, Attachment C.8
7	Comparison of Monte Carlo saturated zone simulations	Aquifer Module vs. EPACMTP	Figure C.8-11, Attachment C.8

Table C.3-6b Verification Cases for the 3MRA Aquifer Module (1999)

Table C.3-6c Verification Cases for the 3MRA Pseudo-Three Dimensional Aquifer Module (1999)

Case	Description	Verification Method	Excerpts of Verification Results Presented in
-	Average Groundwater Specific Flow Rate	Aquifer Module vs. Darcy's Law analytical solution	Figure C.9-1, Attachment C.9
7	Numerical Component of the Contaminant Transport Sub-module	Aquifer Module vs. analytical solution by Ogata	Figure C.9-2, Attachment C.9
ю	Analytical-Numerical Component of the Contaminant Transport Sub-module	Aquifer Module vs. analytical solution	Figure C.9-3, Attachment C.9

C.3.5.1 Vadose-Zone Module Verification

There are 40 vadose-zone module verification test cases summarized in Table C.3-7a. Selected figures for Test Areas 4 and 5, the nonmetals and metals transport components, are presented in Attachment C.10. Reference to the figures in Attachment C.10 is provided in Table C.3-7a. Additional information regarding the test cases and respective verification results may be found in U.S. EPA (2000a). The reading and screening of source and site-specific input data was verified in three cases. Verification of the pre-simulation processing of input data was performed with two cases. The flow, nonmetal transport, and metals transport components were verified with 1, 5, and 4 cases, respectively. The post simulation processing of output data was verified with two cases. The robustness testing verified the stability of the simulation when executed with extreme values for selected parameters. The parameters were selected based on the results of a parameter sensitivity analysis (U.S. EPA, 1996e). The vadose-zone module's robustness was verified with 13 cases.

Test Area	General Requirements	Number of Verification Cases	Excerpts of Verification Results Presented in
1	Verification of reading and screening of source and site-specific input data	3	N/A
2	Verification of pre-simulation processing of input data	2	N/A
3	Verification of the flow component	1	N/A
4	Verification of the nonmetals transport component	5	Figure C.10-1, Attachment C.10
5	Verification of the metals transport component	4	Figure C.10-2, Attachment C.10
6	Verification of post simulation output	2	N/A
7	Verification of the vadose-zone module's robustness	13	N/A

 Table C.3-7a.
 Verification Cases for the 3MRA Vadose-Zone Module (2000)

C.3.5.2 Aquifer Module Verification

There are 69 aquifer module verification cases summarized in Table C.3-7b. Selected figures for Test Area 8, the fate and transport component, are present in Attachment C.11. Reference to the figures in Attachment C.11 is provided in Table C.3-7b. Additional information regarding the test cases and respective verification results may be found in U.S. EPA (2000b). The reading and screening of source and site-specific input data was verified in four cases. Verification of the pre-simulation processing of input data was performed with 17 cases. The fractured media, and heterogeneous saturated media components were verified with 3, and 1 cases, respectively. The reading and screening of chemical-specific, biodegradation and metal-specific data was verified in six tests. The numerical grid generation was verified in four cases. The flow component was verified with 4 cases, while the contaminant fate and transport component was verified in 19 cases. The robustness testing verified the stability of the simulation when executed with extreme values for selected parameters. The parameters were selected based on the results of a parameter sensitivity analysis (U.S. EPA, 1996e). The aquifer module's robustness was verified with 11 cases.

Test Area	General Requirements	Number of Verification Cases	Excerpts of Verification Results Presented in
1	Verification of reading and screening of source and site-specific input data	4	N/A
2	Verification of pre-simulation processing of input data	17	N/A
3	Verification of the fractured media component	3	N/A
4	Verification of the heterogeneous saturated media component	1	N/A
5	Verification of reading and screening of chemical- specific, biodegradation, and metal-specific data	6	N/A
6	Verification of numerical grid generation	4	N/A
7	Verification of the flow component	4	N/A
8	Verification of the contaminant fate and transport component	19	Figures C.11-1 and C.11-2, Attachment C.11
9	Verification of the aquifer module's robustness	11	N/A

Table C.3-7b	. Verification Cases	s for the 3MRA Aquif	er Module (2000)
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C.4 Validation History

Validation, as defined previously, may be conducted using actual measured field data. It is helpful to assess the validity of simplifying assumptions and the predictive capabilities of EPACMTP against well documented realistic site data. EPACMTP and its predecessors (from which flow and transport components in EPACMTP were derived) have been validated based on actual observations at four sites, although no validation has been performed using the 3MRA vadose-zone and aquifer zone modules. In 1990, EPACMS (CANSAZ) was validated against the data from the Borden Landfill site, along with the data from a second agricultural field site on Long Island, New York (U.S. EPA, 1990). This validation included the combination of the saturated and the vadose-zone modules in EPACMS. In 1993, the composite model was validated against data from a Dodge City, Kansas site (Kool et al., 1994). Then, in 1995 EPACMTP was validated against the data from the EBOS Site 24 in New York (U.S. EPA, 1995). The four validation cases are presented in the following subsections. Note that all the figures referenced to in the following subsections are presented.

C.4.1 Borden Site

The Borden landfill is located in Borden, southern Ontario, Canada, and occupies an area of approximately 4 ha (Figure C.12-1). The landfill was operational for 36 years and at its closure was capped with a thin layer of sand. The site overlies 8 to 20 meters of a glaciofluvial sand aquifer, which overlies a confining silty clay deposit. The chloride plume extends about 700 meters northward of the landfill and occupies nearly the entire vertical thickness of the aquifer. The waste material was deposited just above the water table, therefore transport did not occur in the vadose-zone.

Generally, the flow and transport parameters and the procedure described by Frind and Hokkanen (1987) were used for the EPACMS simulation. The exception is that Frind and Hokkanen assigned a higher recharge rate to some areas outside of the source area, but this refinement was not

utilized for the EPACMS simulation. A curvilinear grid was utilized to describe the aquifer geometry and because EPACMS assumes a constant saturated thickness, the VAM3-D-CG code (Huyakorn and Panday, 1989) was used to perform the groundwater flow simulation. Next, the CANSAZ (EPACMTP module) module was used to simulate transient transport. CANSAZ utilized the same finite grid as the groundwater flow simulation, as well as the groundwater velocity distribution from the VAM3-D simulation.

The chloride concentrations were compared for the observed values (Figure C.12-2), the CANSAZ simulation values (Figure C.12-3) and the Frind and Hokkanen simulation values (Figure C.12-4). The CANSAZ model accurately predicted the extent of the plume and the overall plume shape compared to both the Frind and Hokkanen model and the field values.

C.4.2 Long Island Site

The site is located on the south shore of the North Fork of Long Island, New York (Figure C.12-5). The agricultural site was contaminated with the pesticide aldicarb in the 1970's. The source was a 2.5 ha potato field overlying sandy loam soils with a high infiltration rate. An unconfined aquifer is located approximately 2 meters below the surface.

Both site-specific data and monitoring data are limited at this site. The site characterization was obtained from previous studies by INTERA (1980) and Carsel et al. (1985). The EPA Pesticide Root Zone Model (PRZM) (Carsel et al., 1984) was used to predict the 3-year average recharge rate and average aldicarb concentration at the base of the root zone as input for the EPACMS vadose-zone module. The steady-state groundwater flow field was generated using the analytical 2-D solution on EPACMS, followed by a 3-year transient aldicarb transport simulation.

The simulated concentrations of aldicarb in groundwater with distance from the source were compared with the observed values (Figure C.12-6). The agreement between the simulated and observed concentrations was quite reasonable, with the relative error decreasing with increasing distance from the source.

C.4.3 Dodge City Site

The Dodge City, Kansas site (Figure C.12-7), located in the Arkansas River valley, is documented by Ourisson et al. (1992). The source is a controlled release of Triasulfuron pesticide (nonconservative) and bromide (conservative) which, over a 2-year period, is transported through the vadose-zone and the aquifer. The site covers an area of approximately 2.3 acres (approximately 1 ha) overlying one meter of sandy loam soil which overlies a sand and gravel unit. The water table is located at a depth of 3 meters.

The Dodge City site was well characterized, the source was well defined and the monitoring data were available for both soil and groundwater. Site-specific values were obtained from Ourisson et al., (1992), Carsel and Parrish (1988), Gelhar et al. (1985), Carsel et al. (1984), and derived values. EPACMTP was used to simulate the flow and transport of both conservative and nonconservative constituents.

The groundwater concentration model predictions were compared against the observed values. The model tended to underestimate bromide concentrations (Figure C.12-8) slightly and overestimate Triasulfuron concentrations (Figure C.12-9). The application of the model to the

Kansas field site showed reasonably good agreement between model predictions and groundwater monitoring results.

C.4.4 EBOS Site

The Electric Power Research Institute (EPRI) research site referred to as EBOS site 24 is a disposal site for a coal tar Manufacturing Gas Plant located in New York state (Figure C.12-10). Initially, the coal tar was disposed of in a trench on the site, then over time migrated downward into the aquifer (Figure C.12-11 and C.12-12). The site is characterized by 15 to 30 feet of typical glacial outwash sand deposits overlying a clay confining layer. Napthalene was labeled the constituent of concern, since it was the polyaromatic hydrocarbon (PAH) with the highest concentrations in the coal tar.

The site-specific parameters were provided by EPRI in 1993 and consisted of both known and estimated values. EPACMTP was used to simulate the flow and transport of constituents through the vadose-zone and the aquifer. One point to note was that since the coal tar had moved down into the aquifer, the constituents could be leached out through direct contact with ambient groundwater. In the EPACMTP simulation, it was necessary to leach the constituents out of the waste by infiltration from the vadose-zone.

Napthalene concentrations near the source before (Figure C.12-13) and after (Figure C.12-14) source removal were predicted by EPACMTP. The results from EPACMTP were qualitatively similar to the observed concentration in terms of groundwater concentrations near the source.

C.5 Summary

EPACMTP, its predecessors (EPACMS, CANSAZ, and FECTUZ), and its derivatives (3MRA vadose-zone and aquifer modules) have been verified extensively during the past decade at each of the developmental stages. The model has been verified, in numerous cases, by comparing the simulation results against both analytical and numerical solutions. Additionally, EPACMTP and its predecessors have been validated using actual site data from four different sites. The relevant verification and validation history, discussed in the previous sections of this document, is summarized below.

The preliminary verification of EPACMTP was performed by ORD in 1992. Following the preliminary verification, detailed module-level verification was conducted on the flow and transport sub-modules of the vadose-zone and the aquifer modules between 1993-1994. The modules were verified against analytical solutions, and numerical solutions from a number of well-documented simulators. In 1997, the EPACMTP code was verified utilizing a testing plan developed according to ASTM standards. The vadose-zone and the aquifer modules, as well as the composite model (based on the sequentially linked vadose-zone and aquifer modules), were verified against analytical and numerical solutions. In 1999, the vadose-zone and aquifer modules were extracted from EPACMTP to be included as part of the 3MRA software system. The flow and transport sub-modules of both modules were verified against the results from EPACMTP. Additionally, for the 3MRA software system a pseudo-3-D aquifer module was developed. An exhaustive verification was conducted of all of the components in the extracted vadose-zone module and the new pseudo-3-D aquifer module in 2000. The modules were verified against analytical solutions and EPACMTP results.

EPACMTP and its predecessors have been validated using field data from four unique sites from 1990-1995. These sites include: the Borden site, the Long Island site, the Dodge City site, and the EBOS site 24.

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Pulse Source with Chemical Decay and Retardation Unsaturated Zone Transport



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Unsaturated Zone Mercury Transport



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Anaerobic Biodegradation

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Figure C.12-1 Location of the Borden Landfill showing the monitoring network. Cross-section A-A' is along longitudinal plume axis (from Frind and Hokkanen, 1987).

Borden Landfill Site

REFUSE





Observed chloride plume along cross-section A-A' in August 1979 (from Frind and Hokkanen, 1987).

Distance (m)

230 ۲

-195

(a)

BORDEN LANDFILL SITE



Figure C.12-3 The simulated chloride plume obtained by the CANSAZ simulation.

BORDEN LANDFILL SITE



Figure C.12-4 The simulated chloride plume obtained by Frind and Hokkanen, 1987.



PLAN VIEW OF LONG ISLAND FIELD SITE

Figure C.12-5 The plan view of the Long Island field site. Groundwater flow directions are shown with arrows.

COMPARISON OF EPACMTP WITH FIELD RESULTS



Figure C.12-6 Comparison between observed and predicted groundwater aldicarb concentrations for the Long Island site in December, 1970 and May, 1980.



Figure C.12-7 Plan view of Dodge City, Kansas site located in the Arkansas River valley.


COMPARISON OF PREDICTED AND OBSERVED BROMIDE BREAKTHROUGH CURVES

Figure C.12-8 Comparison of predicted and observed Bromide breakthrough curves at the Dodge City, Kansas site.





C.12-13

COMPARISON OF PREDICTED AND OBSERVED TRIASULFURON BREAKTHROUGH CURVES

EBOS SITE 24 LAYOUT AND LOCATION OF SOURCE AREA



Figure C.12-10 EBOS Site 24 layout and location of source area.

EBOS SITE 24 GROUNDWATER SAMPLING LOCATIONS



Figure C.12-11 EBOS Site 24 groundwater sampling locations.

CHANGES IN GROUNDWATER NAPTHALENE PLUME OVER TIME

a) Before Source Removal



b) After Source Removal



Figure C.12-12 Changes in the groundwater Napthalene plume over time a) June 1990: Before source removal b) October 1992: After source removal.







