US ERA ARCHIVE DOCUMENT

ENVIRONMENTAL MONITORING AND ASSESSMENT PROGRAM-SURFACE WATERS:

WESTERN PILOT STUDY FIELD OPERATIONS MANUAL FOR WADEABLE STREAMS

Edited by

David V. Peck¹, James M. Lazorchak², and Donald J. Klemm²

¹ U.S. Environmental Protection Agency Regional Ecology Branch Western Ecology Division National Health and Environmental Effects Research Laboratory Corvallis, OR 97333

> ² U.S. Environmental Protection Agency Ecosystems Research Branch Ecological Exposure Research Division National Exposure Research Laboratory Cincinnati, OH 45268

NATIONAL HEALTH AND ENVIRONMENTAL EFFECTS RESEARCH LABORATORY
OFFICE OF RESEARCH AND DEVELOPMENT
U.S. ENVIRONMENTAL PROTECTION AGENCY
RESEARCH TRIANGLE PARK, NC 27711

NATIONAL EXPOSURE RESEARCH LABORATORY
OFFICE OF RESEARCH AND DEVELOPMENT
U.S. ENVIRONMENTAL PROTECTION AGENCY
RESEARCH TRIANGLE PARK, NC 27711

SECTION 1 INTRODUCTION

by

James M. Lazorchak¹, Alan T. Herlihy², Donald J. Klemm¹, and Steven G. Paulsen³

This manual contains procedures for collecting samples and measurement data from various biotic and abiotic components of streams in the western United States. These procedures were initially developed and used between 1993 and 1998 in research studies of the U.S. Environmental Protection Agency's (EPA) Environmental Monitoring and Assessment Program (EMAP), and published in Lazorchak et al. (1998). The purposes of this manual are to: (1) Document the procedures used in the collection of field data and various types of samples for the EMAP Western Pilot Study (EMAP-WP) and (2) provide these procedures for use by other groups participating in EMAP-WP or implementing stream monitoring programs similar to EMAP.

These procedures are designed for use during a one-day visit by a crew of four persons to sampling sites located on smaller, wadeable streams (stream order 1 through 3, or higher for semi-arid and arid regions of the western U.S.). They were initially developed based on information gained from a workshop of academic, State, and Federal experts (Hughes, 1993), and subsequent discussions between aquatic biologists and ecologists within EMAP, with scientists of the U.S. Geological Survey National Water Quality Assessment Program (NAWQA), with biologists from the U.S. Fish & Wildlife Service, and with State and Regional biologists within EPA Region 3. EMAP staff has also sought information from various Federal and State scientists in the western U.S.

U.S. EPA, National Exposure Research Laboratory, Ecological Exposure Research Division, 26 W. Martin L. King Dr., Cincinnati, OH 45268.

Department of Fisheries and Wildlife, Oregon State University, c/o U.S. EPA. 200 SW 35th St., Corvallis, OR 97333.

³ U.S. EPA, National Health and Environmental Effects Research Laboratory, Western Ecology Division, 200 SW 35th St., Corvallis, OR 97333.

EMAP initiated additional research activities in 1997 to develop field procedures for use in nonwadeable riverine systems. These procedures are currently still under development and will be published separately.

1.1 OVERVIEW OF EMAP-SURFACE WATERS

The U.S. EPA has designated EMAP to develop the necessary monitoring tools to determine the current status, extent, changes and trends in the condition of our nation's ecological resources on regional and national scales (U.S. EPA, 1998). The nation's ecological resources are a national heritage, as essential to the country now and in the future as they have been in the past. Data indicate that regional and international environmental problems may be endangering these essential resources. The potential threats include acid rain, ozone depletion, point and nonpoint sources of pollution, and climate change.

The tools being developed by EMAP include appropriate indicators of ecological condition, and statistical sampling designs to determine the status and extent of condition, and to detect regional-scale trends in condition. When fully implemented in a national monitoring framework, such as that being developed by the White House Committee on Environment and Natural Resources (CENR; Committee on Environment and Natural Resources, 1997), these tools will provide environmental decision makers with statistically valid interpretive reports describing the health of our nation's ecosystems (Whittier and Paulsen, 1992). Knowledge of the health of our ecosystems will give decision makers and resource managers the ability to make informed decisions, set rational priorities, and make known to the public costs, benefits, and risks of proceeding or refraining from implementing specific environmental regulatory actions. Ecological status and trend data will allow decision makers to objectively assess whether or not the nation's ecological resources are responding positively, negatively, or not at all, to existing or future regulatory programs.

The following three objectives guide EMAP research activities (U.S. EPA, 1998):

- Estimate the current status, extent, changes and trends in indicators of the condition of the nation's ecological resources on a regional basis with known confidence.
- Monitor indicators of pollutant exposure and habitat condition and seek associations between human-induced stresses and ecological condition.
- Provide periodic statistical summaries and interpretive reports on ecological status and trends to resource managers and the public.

The EMAP Surface Waters Resource Group (EMAP-SW) is charged with developing the appropriate tools to assess the health of lakes, streams, and wetlands in the United States. The first phase of the program started with a study of northeastern lakes between 1991 and 1996 (Larsen and Christie, 1993; Baker et al., 1997). In 1992 and 1993, a pilot study of wetland ecosystems was conducted in the Prairie Pothole region of the northern plains region of the U.S. (Peterson et al., 1997). The specific research studies dealing with streams are described in more detail in the following section.

1.2 STREAM SAMPLING COMPONENTS OF EMAP-SURFACE WATERS

The procedures presented in this manual were developed and refined during several different research projects conducted between 1993 and 1997. These projects represent two types of field activities to be performed prior to full-scale implementation of a monitoring program that addresses EMAP objectives. *Pilot projects* are intended to answer questions about proposed ecological indicators, such as plot design (how to obtain representative samples and data from each stream site), responsiveness to various stressors, evaluation of alternative methods, and logistical constraints. Pilot studies are not primarily intended to provide regional estimates of condition, but may provide these estimates for a few indicators.

Demonstration projects are conducted at larger geographic scales, and may be designed to answer many of the same questions as pilot studies. Additional objectives of these larger studies are related to characterizing spatial and temporal variability of ecological indicators, and to demonstrating the ability of a suite of ecological indicators to estimate the condition of regional populations of aquatic resources.

1.2.1 Mid-Atlantic Highlands Assessment Project

The stream sampling component of EMAP-SW was initiated in 1993 in the mid-Appalachian region of the eastern United States, in conjunction with a Regional-EMAP (R-EMAP) project being conducted by EPA Region 3. This R-EMAP study was known as the Mid-Atlantic Highlands Assessment study (MAHA), and was carried out over a 4-year period. The MAHA project was designed to test the EMAP approach in a few of the most heavily impacted ecoregions of Region 3, the mid-Appalachians, the Ridge and Valley, the Central Appalachians, the Piedmont and some of the Coastal Plain.

The Region 3 R-EMAP project was designed to answer the following questions:

- What are biological reference conditions for the Central Appalachian Ridge and Valley Ecoregion?
- Do biological communities differ between subregions?
- What is the status of mid-Atlantic Highlands stream biota?
- Can linkages be established between impairment and possible causes of impairment?
- How can an EMAP-like approach be used to design programs to restore and manage stream resources on a regional scale?

During the MAHA study, 577 wadeable stream sites throughout EPA Region 3 (DE, MD, VA, WV, PA) and the Catskill Mts. of New York were visited and sampled using the field protocols being developed by EMAP. Streams were sampled each year during a 10-week index period from April to July by field crews from EPA, the U.S. Fish and Wildlife Service, State, and contract personnel.

1.2.2 Mid-Atlantic Integrated Assessment Program

In 1997 and 1998 the EMAP Surface Waters Program became a collaborator in the Mid-Atlantic Integrated Assessment (MAIA) project, which is attempting to produce an assessment of the condition of surface water and estuarine resources. The MAIA project represented a follow-up to the MAHA study, with an expanded geographic scope (southern New York to northern North Carolina, with more sites located in the Piedmont and Coastal Plain ecoregions) and a different index period (July-September). The first year of the MAIA study, approximately 200 sites (150 wadeable sites, 13 repeated wadeable sites, and approximately 30 riverine sites) were visited for sampling.

1.2.3 Temporal Integrated Monitoring of Ecosystems Project

A special interest component of EMAP-SW is the Temporal Integrated Monitoring of Ecosystems Project (TIME). The purpose of the TIME project is to assess the changes and trends in chemical condition in acid-sensitive surface waters (lakes and streams) of the northeastern and eastern U.S. resulting from changes in acidic deposition caused by the 1990 Clean Air Act Amendments. The TIME project has three goals (Stoddard, 1990):

- Monitor current status and trends in chemical indicators of acidification in acid-sensitive regions of the U.S.
- Relate changes in deposition to changes in surface water conditions.

 Assess the effectiveness of the Clean Air Act emissions reductions in improving the acid/base status of surface waters.

1.2.4 Other Projects

The basic procedures and methods presented in this manual have also been used in other areas of the U.S. as part of R-EMAP projects being conducted by other EPA Regions. These include Regions 7 (central U.S.), 8 (Colorado), 9 (California), and 10 (Oregon and Washington). Each of these projects have modified the basic procedures to be compatible with the geographic region or other project-specific requirements.

1.2.5 Western Pilot Study

The second major geographic study within EMAP is targeted for the states and tribal nations in the western conterminous U.S. Details regarding this research initiative can be found in the peer-reviewed research plan (U.S. EPA, 2000). The purpose for this western study is to further advance the science of monitoring and to demonstrate the application of core tools from EMAP in monitoring and assessment across the West. The Western Geographic Study will serve to advance both the science of monitoring and the application of monitoring to policy, provide an opportunity to push the science and its application to new levels, both in terms of the type of systems addressed (mountainous and arid systems) and the size of the region covered (essentially one third of the conterminous U.S), and demonstrate the application of EMAP designs in answering the urgent and practical assessment questions facing the western EPA Regional Offices, while framing these unique studies in a methodology that can be extended to the entire nation.

The primary objectives of the Western Pilot Study (EMAP-WP), the surface waters component of the Western Geographic Study are to:

- Develop the monitoring tools (biological indicators, stream survey design, estimates of reference condition) necessary to produce unbiased estimates of the ecological condition of surface waters across a large geographic area (or areas) of the West; and
- Demonstrate those tools in a large-scale assessment.

The goal of EMAP-WP is to provide answers to three general assessment questions:

1. What proportion of stream and river miles in the western U.S. are in acceptable (or poor) biological condition?

- 2. What is the relative importance of potential stressors (habitat modification, sedimentation, nutrients, temperature, grazing, timber harvest, etc.) in streams and rivers across the West?; and
- 3. With what stressors are streams and rivers in poor biological condition associated?

The resource population of interest for EMAP-WP are all perennial streams and rivers as represented in EPA's River Reach File (RF3), with the exception of the "Great Rivers" (the Columbia, Snake, Colorado and Missouri Rivers). The pilot study will utilize an EMAP probability design to select sites which are statistically representative of the resource population of interest. This will allow one to extrapolate ecological results from the sites sampled to the entire population. A comprehensive set of ecological indicators (see below) will be implemented in a coarse survey of streams and rivers across all of the West (the conterminous portions of EPA Regions 8, 9 and 10), as well as in several more spatially-intensive "focus areas" in each Region (see Figure 1-1). Sample sizes (i.e., numbers of stream sites) have been chosen to allow eventual estimates of condition to be made for each state, each Regional focus area, numerous aggregated ecological regions (e.g., mountainous areas of the Pacific states, the Southern Basin and Range, etc.), major river basins, and many other potential geographic classifications.

1.3 SUMMARY OF ECOLOGICAL INDICATORS

The following sections describe the rationale for each of the ecological indicators currently included in the stream sampling procedures presented in this manual. Evaluation activities to determine the suitability of individual indicators to robustly determine ecological condition are ongoing at this time. This information is presented to help users understand the various field procedures and the significance of certain aspects of the methodologies.

Currently, EMAP considers two principal types of indicators, condition and stressor (U.S. EPA, 1998). Condition indicators are biotic or abiotic characteristics of an ecosystem that can provide an estimate of the condition of an ecological resource with respect to some environmental value, such as biotic integrity. Stressor indicators are characteristics that are expected to change the condition of a resource if the intensity or magnitude is altered.

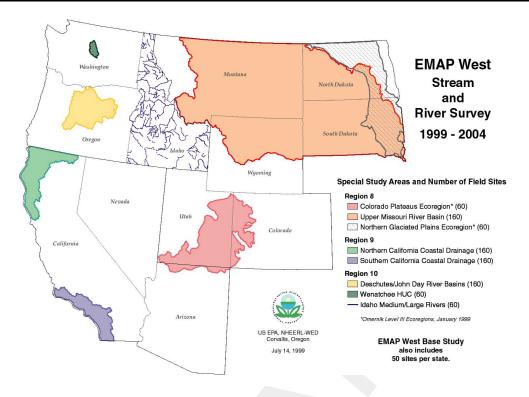


Figure 1-1. The geographic scope of the EMAP-Surface Waters Western Pilot Study, including the "special interest" study areas within each EPA Region.

1.3.1 Water Chemistry

Data are collected from each stream for a variety of physical and chemical constituents. Information from these analyses is used to evaluate stream condition with respect to stressors such as acidic deposition (of importance to the TIME project), nutrient enrichment, and other inorganic contaminants. In addition, streams can be classified with respect to water chemistry type, water clarity, mass balance budgets of constituents, temperature regime, and presence of anoxic conditions.

1.3.2 Physical Habitat

Naturally occurring differences among surface waters in physical habitat structure and associated hydraulic characteristics contributes to much of the observed variation in species composition and abundance within a zoogeographic province. The structural complexity of aquatic habitats provides the variety of physical and chemical conditions to

support diverse biotic assemblages and maintain long-term stability. Anthropogenic alterations of riparian areas and stream channels, wetland drainage, grazing and agricultural practices, and stream bank modifications such as revetments or development, generally act to reduce the complexity of aquatic habitat and result in a loss of species and ecosystem degradation.

Stressor indicators derived from data collected about physical habitat quality will be used to help explain or diagnose stream condition relative to various condition indicators. Important attributes of physical habitat in streams are channel dimensions, gradient, substrate characteristics; habitat complexity and cover; riparian vegetation cover and structure; disturbance due to human activity, and channel-riparian interaction (Kaufmann, 1993). Overall objectives for this indicator are to develop quantitative and reproducible indices, using both multivariate and multimetric approaches, to classify streams and to monitor biologically relevant changes in habitat quality and intensity of disturbance. Kaufmann et al. (1998) discuss procedures for reducing EMAP field habitat measurements and observations to metrics that describe channel and riparian habitat at the reach scale.

1.3.3 Periphyton Assemblage

Periphyton are the algae, fungi, bacteria, and protozoa associated with substrates in aquatic habitats. These organisms exhibit high diversity and are a major component in energy flow and nutrient cycling in aquatic ecosystems. Many characteristics of periphyton community structure and function can be used to develop indicators of ecological conditions in streams (Hill et al., 1999). Periphyton are sensitive to many environmental conditions, which can be detected by changes in species composition, cell density, ash free dry mass (AFDM), chlorophyll, and enzyme activity (e.g., alkaline and acid phosphatase). Each of these characteristics may be used, singly or in concert, to assess condition with respect to societal values such as biological integrity and trophic condition.

A hierarchical framework is being used in the development of the periphyton indices of stream condition. The framework involves the calculation of composite indices for biotic integrity, ecological sustainability, and trophic condition. The composite indices will be calculated from measured or derived first-order and second-order indices. The first-order indices include species composition (richness, diversity), cell density, AFDM, chlorophyll, and enzyme activity (e.g., Saylor et al., 1979), which individually are indicators of ecological condition in streams. Second-order indices will be calculated from periphyton characteristics, such as the autotrophic index (Weber, 1973), community similarity compared to refer-

ence sites, and autecological indices (e.g., Lowe, 1974; Lange-Bertalot, 1979; Charles, 1985; Dixit et al, 1992).

1.3.4 Benthic Macroinvertebrate Assemblage

Benthic macroinvertebrates inhabit the sediment or live on the bottom substrates of streams. The macroinvertebrate assemblages in streams reflect overall biological integrity of the benthic community, and monitoring these assemblages is useful in assessing the status of the water body and discerning trends. Benthic communities respond differently to a wide array of stressors. As a result of this, it is often possible to determine the type of stress that has affected a benthic macroinvertebrate community (Plafkin et al., 1989; Klemm et al., 1990; Barbour et al. 1999). Because many macroinvertebrates have relatively long life cycles of a year or more and are relatively immobile, macroinvertebrate community structure is a function of past conditions.

Two different approaches are currently being evaluated to developing ecological indicators based on benthic invertebrate assemblages. The first is a multimetric approach, where different structural and functional attributes of the assemblage are characterized as "metrics". Individual metrics that respond to different types of stressors are scored against expectations under conditions of minimal human disturbance. The individual metric scores are then summed into an overall index value that is used to judge the overall level of impairment of an individual stream reach. Examples of multimetric indices based on benthic invertebrate assemblages include Kerans and Karr (1993), Fore et al. (1996) and Barbour et al. (1995; 1996).

The second approach being investigated is to develop indicators of condition based on multivariate analysis of benthic assemblages and associated abiotic variables. Examples of this type of approach as applied to benthic invertebrate assemblages include RIVPACS (Wright, 1995), and BEAST (Reynoldson et al., 1995). Rosenberg and Resh (1993) present various approaches to biological monitoring using benthic invertebrates, and Norris (1995) briefly summarizes and discusses approaches to analyzing benthic macroinvertebrate community data.

1.3.5 Aquatic Vertebrate Assemblages

Aquatic vertebrate assemblages of interest to EMAP include fish and amphibians. The fish assemblage represents a critical component of biological integrity from both an ecosystem function and a public interest perspective. Historically, fish assemblages have

been used for biological monitoring in streams more often than in lakes (e.g., Plafkin et al., 1989; Karr, 1991). Fish assemblages can serve as good indicators of ecological conditions because fish are long-lived and mobile, forage at different trophic levels, integrate effects of lower trophic levels, and are reasonably easy to identify in the field (Plafkin et al., 1989). Amphibians comprise a substantial portion of vertebrate biomass in streams of many areas of the U.S. (Hairston, 1987; Bury et al., 1991). Reports of dramatic declines in amphibian biodiversity (e.g., Blaustein and Wake, 1990; Phillips, 1990) has increased the level of interest in monitoring these assemblages. Amphibians may also provide more information about ecosystem condition in headwater or intermittent streams in certain areas of the country than other biological response indicators (Hughes, 1993). The objective of field sampling is to collect a representative sample of the aquatic vertebrate assemblage by methods designed to 1) collect all except very rare species in the assemblage and 2) provide a measure of the abundance of species in the assemblages (McCormick, 1993). Information collected for EMAP that is related to vertebrate assemblages in streams includes assemblage attributes (e.g., species composition and relative abundance) and the incidence of external pathological conditions.

Indicators based on vertebrate assemblages are being developed primarily using the multimetric approach described in Section 1.3.5 for benthic macroinvertebrates, and originally conceived by Karr and others (Karr et al., 1986). Simon and Lyons (1995) provide a recent review of multimetric indicators as applied to stream fish assemblages. (McCormick et al. (In press) provide an example of a multimetric indicator developed for the Mid-Atlantic region using EMAP data, based on an evaluation process described by Hughes et al. (1998).

1.3.6 Fish Tissue Contaminants

Indicators of fish tissue contaminants attempt to provide measures of bioaccumulation of toxic chemicals in fish. The primary purpose of determining contaminant levels in fish tissue is to provide a measure of the potential exposure of stream systems to toxic compounds. It is also meant to be used in conjunction with the other stressor indicators (physical habitat, water chemistry, land use, population density, other records of relevant anthropogenic stresses) and condition indicators (fish, macroinvertebrates, periphyton) to help diagnose whether the probable cause of stream degradation, when it is shown by the condition indicators to occur, is water quality, physical habitat, or both.

The various studies that have been done on fish tissue contaminants have focused on different parts of the fish: whole fish, fillets, livers. For EMAP-SW, the focus is on

whole fish because of the emphasis on the ecological health of the whole stream (as opposed to a focus on human health concerns). Whole fish are a better indicator of risk to piscivorous wildlife than fillets. It is hoped to also be able to say something about risks to human health by analyzing whole fish. Whole fish also present fewer logistical problems for field crews (no gutting required in the field) and the analytical lab (no filleting necessary).

Samples are prepared for two major categories of fish species. One sample is prepared using a species whose adults are small (e.g., small minnows, sculpins, or darters). The second sample is prepared using a species whose adults are of larger size (e.g., suckers, bass, trout, sunfish, carp). In addition to being more ubiquitous than the larger fish (and therefore more likely to be present in sufficient numbers to composite), small fish have other advantages over large fish. Most importantly, it may be possible to get a more representative sample of the contaminant load in that stream segment (although it could be at a lower level of bioaccumulation) by creating a composite sample from a larger number of small individuals than by compositing a few individuals of larger species. The major advantage that larger fish could potentially offer, whether predators (piscivores) or bottom feeders, is a higher level of bioaccumulation and thus greater sensitivity to detect contaminants. The relative bioaccumulation of contaminants by large and small stream fish is not known, thus the reason for preparing two samples in this study.

In addition, specimens are collected for determination of the presence of various internal pathogens..

1.4 OBJECTIVES AND SCOPE OF THE FIELD OPERATIONS MANUAL

Only field-related sampling and data collection activities are presented in this manual. Laboratory procedures and methods (including sample processing and analytical methods) associated with each ecological indicator are summarized in Chaloud and Peck (1994); detailed procedures will be published as a separate document.

This manual is organized to follow the sequence of field activities during the 1-day site visit. Section 2 presents a general overview of all field activities. Section 3 presents those procedures that are conducted at a "base" location before and after a stream site visit. Section 4 presents the procedures for verifying the site location and defining a reach of the stream where subsequent sampling and data collection activities are conducted. Sections 5 through 14 describes the procedures for collecting samples and field measurement data for various condition and stressor indicators. Specific procedures associated with each indicator are presented in standalone tables that can be copied, laminated, and taken

into the field for quick reference. Section 15 describes the final activities that are conducted before leaving a stream site. Appendix A contains a list of all equipment and supplies required by a crew to complete all field activities at a stream.

Depending on the specific project and approach to information management, field teams may also be provided with an information management handbook that contains instructions for tracking samples and generating sampling status reports as well as using the computers and associated hardware and software. Field teams are also required to keep the field operations and methods manual available in the field for reference and to address questions pertaining to protocols that might arise.

1.5 QUALITY ASSURANCE

Large-scale and/or long-term monitoring programs such as those envisioned for EMAP require a rigorous quality assurance (QA) program that can be implemented consistently by all participants throughout the duration of the monitoring period. Quality assurance is a required element of all EPA-sponsored studies that involve the collection of environmental data (Stanley and Verner, 1986). Field teams should be provided a copy of the QA project plan (e.g., Chaloud and Peck, 1994 for EMAP-SW activities). The QA plan contains more detailed information regarding QA/QC activities and procedures associated with general field operations, sample collection, measurement data collection for specific indicators, and data reporting activities. A QA project plan will be prepared for the Western Pilot Study and distributed to all participants.

Quality control (QC) activities associated with field operations are integrated into the field procedures. Important QA activities associated with field operations include a comprehensive training program that includes practice sampling visits, and the use of a qualified museum facility or laboratory to confirm any field identifications of biological specimens. The overall sampling design for EMAP-SW related studies usually includes a subset of sites (10 to 15 percent) that are revisited within a single sampling period and/or across years (e.g., Larsen, 1997; Urquhart et al., 1998). Information from these repeat visits is used in part to describe overall sampling and measurement precision for the various ecological indicators.

1.6 LITERATURE CITED

- Barbour, M.T., J.B. Stribling, and J.R. Karr. 1995. The multimetric approach for establishing biocriteria and measuring biological condition. pp. 69-80 IN: W.S. Davis and T.P. Simon (eds.) *Biological Assessment and Criteria: Tools for Water Resource Planning and Decision-making*. Lewis Publishers, Chelsea, Michigan.
- Barbour, M.T., J. Gerritsen, G.E. Griffith, R. Frydenborg, E. McCarron, J.S. White, and M.L. Bastian. 1996. A framework for biological criteria for Florida streams using benthic macroinvertebrates. *Journal of the North American Benthological Society* 15(2):185-211.
- Barbour, M.T., J. Gerritsen, B.D. Snyder, and J.B. Stribling. 1999. *Rapid Bioassessment Protocols for Use in Streams and Wadeable Rivers: Periphyton, Benthic Macroinvertebrates, and Fish.* Second Edition. EPA/841-B-99-002. U.S. Environmental Protection Agency, Office of Water, Assessment and Watershed Protection Division, Washington, D.C.
- Baker, J.R., D.V. Peck, and D.W. Sutton (editors). 1997. *Environmental Monitoring and Assessment Program-Surface Waters: Field Operations Manual for Lakes*. EPA/620/R-97/001. U.S. Environmental Protection Agency, Washington, D.C.
- Blaustein, A.R. and D.B. Wake. 1990. Declining amphibian populations: a global phenomenon? *Trends in Ecology and Evolution* 5:203-204.
- Bury, R.B., P.C. Corn, K.B. Autry, F.F. Gilbert, and L.L.C. Jones. 1991. Aquatic amphibian communities in Oregon and Washington. pp. 353-362 IN: L.F. Ruggiero, K.B. Aubry, A.B. Carey, and M.H. Huff (coordinators). *Wildlife and Vegetation of Unmanaged Douglas-Fir Forests*. General Technical Report PNW-GRT-285. USDA Forest Service, Portland, Oregon.
- Chaloud, D. J., and D. V. Peck (eds.). 1994. Environmental Monitoring and Assessment Program: Integrated Quality Assurance Project Plan for the Surface Waters Resource Group. EPA 600/X-91/080. Revision 2.00. U.S. Environmental Protection Agency, Las Vegas, Nevada.
- Charles, D.F. 1985. Relationships between surface sediment diatom assemblages and lakewater characteristics in Adirondack lakes. *Ecology* 66:994-1011.

- Committee on Environment and Natural Resources. 1997. *Integrating the Nation's Environmental Monitoring and research Networks and Programs: A Proposed Framework.* March 1997 revision. Office of Science and Technology Policy, Washington, DC.
- Dixit, S.S., J.P. Smol, J.C. Kingston, and D.F. Charles. 1992. Diatoms: Powerful indicators of environmental change. *Environmental Science and Technology* 26:22-33.
- Fore, L.S., J.R. Karr, and R.W. Wisseman. 1996. Assessing invertebrate responses to human activities, evaluating alternative approaches. *Journal of the North American Benthological Society* 15:212-231.
- Hairston, N.G. 1987. *Community Ecology and Salamander Guilds*. Cambridge University Press.
- Hill, B.A., A.T. Herlihy, P.R. Kaufmann, R.J. Stevenson, F.H. McCormick, and C. Burch-Johnson. 2000. Use of periphyton assemblage data as an index of biotic integrity. *Journal of the North American Benthological Society* 19(1):50-67.
- Hughes, R.M. (ed.). 1993. *Stream Indicator and Design Workshop.* EPA/600/R-93/138. U.S. Environmental Protection Agency, Corvallis, Oregon.
- Hughes, R. M., P. R. Kaufmann, A. T. Herlihy, T. M. Kincaid, L. Reynolds, and D. P. Larsen. 1998. A process for developing and evaluating indices of fish assemblage integrity. Canadian Journal of Fisheries and Aquatic Sciences 55: 1618-1631.
- Karr, J.R. 1991. Biological integrity: a long neglected aspect of water resource management. *Ecological Applications* 1:66-84.
- Karr, J.R., K.D. Fausch, P.L. Angermeier, P.R. Yant, and I.J. Schlosser. 1986. *Assessing Biological Integrity in Running Waters: A Method and its Rationale.* Illinois Natural History Survey Special Publication 5. Champaign, IL.
- Kaufmann, P.R. (ed.). 1993. Physical Habitat. pp. 59-69 <u>IN</u>: R.M. Hughes (ed.). *Stream Indicator and Design Workshop*. EPA/600/R-93/138. U.S. Environmental Protection Agency, Corvallis, Oregon.

- Kaufmann, P.R., P. Levine, E.G. Robison, C. Seeliger, and D.V. Peck. 1999. Quantifying Physical Habitat in Wadeable Streams. EPA/620/R-99/003. U.S. Environmental Protection Agency, Washington, D.C..
- Kerans, B.L., and J.R. Karr. 1994. A benthic index of biotic integrity (B-IBI) for rivers of the Tennessee Valley. *Ecological Applications* 4:768-785.
- Klemm, D.J., P.A. Lewis, F. Fulk, J.M. Lazorchak. 1990. *Macroinvertebrate Field and Laboratory Methods for Evaluating the Biological Integrity of Surface Waters*. EPA/600/4-90/030. U.S. Environmental Protection Agency, Cincinnati, Ohio.
- Larsen, D.P. and S.J. Christie (eds.) 1993. *EMAP-Surface Waters* 1991 *Pilot Report*. EPA/620/R-93/003. U.S. Environmental Protection Agency, Washington, D.C.
- Lange-Bertalot, H. 1979. Pollution tolerance of diatoms as criterion for water quality estimation. *Nova Hedwigia* 64:285-304.
- Larsen, D.P. 1997. Sample survey design issues for bioassessment of inland aquatic ecosystems. *Human and Ecological Risk Assessment* 3:979-991.
- Lazorchak, J.M., Klemm, D.J., and D.V. Peck (editors). 1998. Environmental Monitoring and Assessment Program -Surface Waters: Field Operations and Methods for Measuring the Ecological Condition of Wadeable Streams. EPA/620/R-94/004F. U.S. Environmental Protection Agency, Washington, D.C.
- Lowe, R.L. 1974. *Environmental Requirements and Pollution Tolerance of Freshwater Diatoms*. U.S. Environmental Protection Agency, Environmental Monitoring Series, National Environmental Research Center, Cincinnati, Ohio.
- McCormick, F.H. 1993. Fish. pp. 29-36 **N**: R.M. Hughes (ed.). *Stream Indicator Work-shop*. EPA/600/R-93/138. U.S. Environmental Protection Agency. Corvallis, Oregon.
- McCormick, F.H., R. M. Hughes, P.R. Kaufmann, A. T. Herlihy, D.V. Peck, and J.L. Stoddard. In press. Development of an Index of Biotic Integrity for the Mid-Atlantic Highlands region. *Transactions of the American Fisheries Society.*
- Norris, R.H. 1995. Biological monitoring: the dilemma of data analysis. *Journal of the North American Benthological Society* 14:440-450.

- Peterson, S.A., L. Carpenter, G. Gutenspergen, and L.M. Cowardin (editors). 1997. *Pilot Test of Wetland Condition Indicators in the Prairie Pothole Region of the United States*. EPA/620/R-97/002. U.S. Environmental Protection Agency, Washington, D.C.
- Phillips, K. 1990. Where have all the frogs and toads gone? *Bioscience* 40:422-424.
- Plafkin, J.L., M.T. Barbour, K.D. Porter, S.K. Gross, and R.M. Hughes. 1989. *Rapid Bio-assessment Protocols for Use in Streams and Rivers: Benthic Macroinvertebrates and Fish.* EPA/440/4-89/001. U.S. Environmental Protection Agency, Washington, D.C.
- Reynoldson, T.B., R.C. Bailey, K.E. Day, and R.H. Norris. 1995. Biological guidelines for freshwater sediment based on Benthic Assessment Sediment (the BEAST) using a multivariate approach for predicting biological state. *Australian Journal of Ecology* 20:198-219.
- Rosenberg, D.M. and V.H. Resh. 1993. Freshwater Biomonitoring and Benthic Macro-invertebrtates. Chapman and Hall, New York.
- Sayler, G.S., M. Puziss, and M. Silver. 1979. Alkaline phosphatase assay for freshwater sediments: application to perturbed sediment systems. *Applied and Environmental Microbiology* 38:922-927.
- Simon, T.P. and J. Lyons. 1995. Application of the index of biotic integrity to evaluate water resources integrity in freshwater ecosystems. pp. 245-262 **IN**: W.S. Davis and T.P. Simon (eds.), *Biological Assessment and Criteria: Tools for Water Resource Planning and Decision Making*. Lewis Publishers, Boca Raton, Florida.
- Stanley, T.W., and S.S. Verner. 1986. The U.S. Environmental Protections Agency's quality assurance program. pp. 12-19 **IN**: J.K. Taylor and T.W. Stanley (eds.). *Quality Assurance for Environmental Measurements*. ASTM STP 867, American Society for Testing and Materials, Philadelphia, Pennsylvania.
- Stoddard, J.L. 1990. Plan for Converting NAPAP Aquatic Effects Long-Term Monitoring (LTM) Project to the Temporally Integrated Monitoring of Ecosystems (TIME) Project. U.S. Environmental Protection Agency, Environmental Research Laboratory, Corvallis, Oregon.

- Urquhart, N.S., S.G. Paulsen, and D.P. Larsen. 1998. Monitoring for policy-relevant regional trends over time. *Ecological Applications* 8:246-257.
- U.S. EPA. 1998. Environmental Monitoring and Assessment Program (EMAP): Research Plan 1997. EPA/620/R-98/002. U.S. Environmental Protection Agency, Washington, D.C.
- U.S. EPA. 2000. Ecological Assessment of Streams and Rivers in the Western United States: A Cooperative Effort Between the U.S. EPA and Western States and Tribal Nations. U.S. Environmental Protection Agency, Corvallis, Oregon
- Weber, C.I. 1973. Recent developments in the measurement of the response of plankton and periphyton to changes in their environment. pp. 119-138 **IN**: G. Glass (ed.), *Bioassay Techniques and Environmental Chemistry.* Ann Arbor Science Publishers, Ann Arbor, Michigan.
- Whittier, T.R. and S.G. Paulsen. 1992. The surface waters component of the Environmental Monitoring and Assessment Program (EMAP): an overview. *Journal of Aquatic Ecosystem Health* 1:119-126.
- Wright, J.F. 1995. Development and use of a system for predicting the macroinvertebrate fauna in flowing waters. *Australian Journal of Ecology* 20:181-197.

NOTES

NOTES

